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	Management of harmful benthic dinoflagellates requires targeted sampling methods and alarm thresholds					
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17	Abstract					

18 Concern regarding Benthic Harmful Algal Blooms (BHABs) is increasing since some harmful 19 benthic species have been identified in new areas. In the Mediterranean basin, the most common 20 harmful benthic microalgae are Ostreopsis cf. ovata and Prorocentrum lima, which produce palytoxin-like compounds and okadaic acid respectively, and the need to implement monitoring 21 activities has increased. However, a general agreement on appropriate strategies (e.g. sampling 22 season, definition of alarm thresholds, etc.) is still lagging behind, especially for P. lima, whose 23 24 proliferation dynamics are still poorly known.

The aim of this study is to highlight similarities and differences between these two benthic 25 26 dinoflagellates, providing a six-year time series of their co-occurrence in relation to environmental variables (in particular, hydrodynamic sea conditions). Our data report species specific effects of 27 the hydrodynamic regime in terms of affinity of microalgal cells with the benthic substrate, 28 29 suggesting that P. lima has a more strictly "benthic behavior" compared to O. cf. ovata.

30 These results highlight the particular need to implement a specific sampling strategy and define benthic alarm thresholds, triggering future applications in a management perspective. 31

32 1. Introduction

33 Benthic Harmful Algal Blooms (BHABs) refer to the proliferation of benthic microalgae, which may represent a serious threat for other organisms (including humans) and the environment. In 34 temperate regions, BHABs started receiving large interest in the last decade, since certain harmful 35 36 benthic species, mainly of tropical origin, have been identified in new areas, gaining the attention of

37 both the scientific community and public governance.

38 As an example, Ostreopsis spp. have been recorded in many locations along the Mediterranean 39 (Accoroni et al., 2016), Australian (Verma et al., 2016), New Zealand (Shears and Ross, 2009) and Japanese coasts (Parsons et al., 2012). This genus is responsible for the Palytoxicosis (foodborne 40 poisoning) and other irritative symptoms (through direct contact and aerosols), respectively in the 41 42 tropics and in temperate areas (Berdalet et al., 2015).

The genus Gambierdiscus has been reported in Micronesia - Macaronesia (Pérez-Arellano et al., 43 44 2005), Mediterranean Sea (Aligizaki et al. 2008), Canary Islands (Fraga et al. 2011), Japan (Nishimura et al., 2013 and 2014) and The Cook Island (Smith et al., 2016); this genus is involved 45 46 in the Ciguatera Fish Poisoning, an endemic human syndrome in tropical and subtropical areas, 47 presently spreading to temperate latitudes.

Species belonging to the genus Prorocentrum spp., and especially Prorocentrum lima, are more 48 49 cosmopolitan than the other BHABs presenting a larger geographical distribution (Aligizaki et al. 2009); some Prorocentrum spp. are responsible for Diarrheic Shellfish Poisoning (DSP) (Berdalet 50 et al., 2015). 51

In the Mediterranean Sea, the most common benthic harmful dinoflagellates are *Ostreopsis* cf.
 ovata and *P. lima*.

55 The former species lives associated with a variety of biotic and abiotic substrata (macroalgae, seagrasses, benthic invertebrates, sand, pebbles and rocks), attached through mucilaginous 56 filaments, which lump together microalgal cells (Totti et al., 2010). In general, O. cf. ovata 57 proliferates during the summer season, from June to September (except for the Adriatic Sea, 58 59 where the maximum abundance is usually reached in October; Accoroni et al., 2012), when sea temperature exceeds 24 °C, especially in sheltered and shallow-water areas (Mangialajo et al., 60 2008; Pezzolesi et al., 2012). During long lasting blooms, its mucilaginous matrix can form a 61 brownish mat that can easily be re-suspended by mechanical action or local hydrodynamics: 62 consequently, the concentration of O. cf. ovata in the water column is related to its abundance in 63 64 the benthic habitat (Mangialajo et al., 2011).

65 This dinoflagellate occurs in almost all Italian coastal waters (ISPRA, 2013) and produces a wide range of palytoxin-like compounds (PLTXs) (Ciminiello et al., 2010, 2012a and 2012b), which can 66 67 affect both humans and marine organisms. Data related to Ostreopsis spp. toxins and their effects are reported in several studies (e.g. Deed and Schwartz, 2010; Ramos, 2010; Alloisio et al., 2016). 68 Considering all available data on O. ovata toxicity and the Acute Reference Dose (ARfD: amount of 69 70 substance that can be consumed over a short period of time without any appreciable health risk for consumers) set by the European Food Safety Authority (ESFA) at 0.2 µg of PLTXs equivalent / kg 71 72 of body weight, the Italian National Institute of Health has recently updated the guidelines for assessing and managing human health risks associated to Ostreopsis (Funari et al., 2014; 2015). 73 74 According to these guidelines, monitoring activities are planned according to observed bloom 75 phase (Routine, Alert and Emergency). In particular, the "Alert status" starts when the cell density 76 of O. cf. ovata in the water column is between 10 000 and 30 000 cells L-1, with a prolonged period 77 (7-10 days) of low hydrodynamic conditions and high temperature.

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Knowledge on P. lima proliferation dynamics, conversely, is very limited, although this species has 79 been frequently recorded in several Mediterranean areas (Catalan coasts, Spain: Vila et al., 2001; 80 North Aegean Sea: Aligizaki et al., 2009; Italian coasts: Simoni et al., 2004; Ingarao et al., 2007; 81 82 Cyprus: author's unpublished data). In general, it shows a golden-brown chloroplast and a centrally located pyrenoid in each valve view, smooth theca and valve and marginal pores, slightly visible by 83 84 light microscope; howeverit displays a wide variability in cell shape and size (Aligizaki et al. 2009). 85 Due to its production of okadaic acid (OA), along with other DSP toxins, P. lima induces diarrheic 86 symptoms by acting as a potent protein phosphatase inhibitor in metabolic processes; chronic 87 exposure to these toxins is hypothesized to stimulate gastrointestinal tumors (Aune & Yndestad, 1993). Regarding toxicity of the Mediterranean P. lima strains, there is lack of information and, so 88 far, no official alarm threshold (in terms of cell concentration) has been established in this region. 89

However, OA-group toxin levels are monitored in Europe and food enterprises must ensure that
bivalve molluscs placed on the market for human consumption do not exceed the current
regulatory limit of 160 µg of OA kg⁻¹ shellfish meat and the Acute Reference Dose fixed by EFSA is
0.3 µg kg⁻¹ body weight. Some European countries recently adopted the threshold of 100 cells L⁻¹
(Ireland, UK) or 500 cells L⁻¹ (Mediterranean coast of Spain) to activate more accurate monitoring,
for example, before the harvesting and sale of aquaculture products (ICES Report, 2015).

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Monitoring of benthic microalgae in the Mediterranean basin recently increased due to the related 97 concerns and, so far, the most common applied strategy is based on seawater sampling and, in a 98 few cases, on macrophyte collection: herein estimation of microalgal species abundances is 99 assessed through light microscopy after detachment from the substrate. While Ostreopsis 100 101 monitoring has been widely discussed and some projects focused on sampling and counting 102 optimization (e.g. www.m3-habs.net), a general agreement on appropriate strategies (e.g. sampling season, definition of alarm thresholds) is still lagging behind, especially regarding other 103 104 potentially toxic species, such as P. lima.

Being the quantification of benthic microalgae more difficult to standardize, due to the spatial and temporal scales of variability of the benthic system, different sampling approaches have been tested (see GEOHAB 2012; Tester et al., 2014; Jauzein et al., 2016). The lack of an appropriate and standardized sampling approach may be responsible for problems in properly measuring of their abundances and perform comparisons across locations.

The aim of this study, is to provide data on proliferation dynamics of *P. lima* along Genoa coast (Italy, North-western Mediterranean Sea), highlighting similarities and eventual differences compared to *O.* cf. *ovata* in terms of "ecological behavior", which could push to the adoption of specific and more targeted sampling strategies, using a six year monitoring data set.

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115 2. Materials and Methods

116 <u>2.1. Sampling strategy</u>

Data presented in this study refer to the BHABs monitoring campaign carried out at Quarto dei
Mille (Genoa, Italy) during the summer seasons from 2010 to 2015, for a total of 87 sampling
occasions: from June 24th to 1st of September 2010, from June 14th to August 31th 2011, from June
20th to October 10th 2012, from June 7th to September 23th 2013, from June 13th to September 9th
2014 and from June 18th to 10th September 2015.

Sampling was performed by collecting seawater samples and macrophytes (the brown alga
 Halopteris scoparia), following the commonly applied protocol(as reported e.g. in Mangialajo et al.,

2011. Sampling was performed in the morning once a week, incremented to 2-3 times per week 124 during the bloom period, keeping a similar sampling time (from 8:00 to 10:00 am) to avoid 125 fluctuations due to potential diel variations between benthic and planktonic pools. Planktonic 126 samples (1 liter of seawater, in duplicate or triplicate) were collected at about -0.3m depth, before 127 collection of benthic ones (5-10 g fresh weight of algae, at -0.5 m deep) in order to avoid a 128 mechanical re-suspension of benthic cells. Samples were fixed with formalin (2%) from 2010 to 129 130 2013 and with acidic Lugol solution followed by storage at 4°C in 2014 and 2015, according to an improved and standardized protocol developed in the framework of the M3-HABs Project 131 (http://m3-habs.net). 132

Cells abundance in seawater samples was assessed using the Utermöhl sedimentation method 133 (100 ml column settled for 24 h). In regard to benthic samples, the protocol was slightly changed in 134 135 2014 adding a rinsing step after the shaking procedure used to detach epiphytic cells. From 2010 to 2013 macrophytes and surrounding seawater were vigorously shaken to detach benthic cells 136 and 3 ml of seawater were settled in the Utermöhl chamber. In 2014 and 2015, isolation of 137 epiphytic cells was performed by vigorously shaking (for 10 sec) the macrophyte samples, then 138 139 rinsing with 100 ml of filtered seawater (0.2 µm) and shaking two times. Counting was performed 140 on 1 ml, using the Sedgewick Rafter chamber.

Abundances of dinoflagellates were expressed as cells g^{-1} fresh weight of macrophyte (fwm) in the case of benthic samples and in cells L⁻¹ for seawater ones. Identification of the two species was carried out through light microscopy.

144 2.2. Environmental data

Seawater temperature and salinity data were recorded *in situ* using a CTD probe (Hanna 9328),
concurrently with sample collection. Data regarding sea condition, reported as days with calm sea
(wave height < 0.5 m) or rough sea (waves > 0.5 m), have been provided by DHI srl
(https://worldwide.dhigroup.com/it).

149 2.3. Data and statistical analyses

Alarm thresholds for *O*. cf. *ovata* used in this work were defined according to the Italian guideline 2014 (10 000 - 30 000 cells L⁻¹, according to Funari et al., 2015). As far as *P*. *lima*, the threshold concentration of 500 cells L⁻¹ was taken into account, as from the ICES Report 2015.

153 Correlation analyses (separately for both target species) were performed between microalgal 154 densities in the water column and on macrophytes to investigate the effects of sea conditions on 155 the re-suspension of cells from the substrate. In addition, for assessing the potential different 156 behavior of *P. lima* and *O.* cf. *ovata* in response to the hydrodynamic regime, the ratio between 157 concentration in the water column and on macrophytes (cells L⁻¹ *versus* cells g⁻¹ fwm) was plotted separately for calm (wave height <0.5 m) or rough sea-conditions (wave height > 0.5 m for more than 6 hours in the two days before sampling). Analysis was performed only on data collected during blooming periods of 2010, 2011, 2012, 2013 and 2015 (from 1st of July to August 15th; 56 total sampling occasions). Summer 2014 has been excluded due to the unavailability of wave height data. Statistical analyses were carried out using the software R (R Development Core Team, 2008).

164 3. Results

165 <u>3.1. Monitoring data</u>

Maximum cell density values of *O*. cf. *ovata* and *P*. *lima* and related alarm thresholds, seawater temperature and hydrodynamic conditions are summarized in Table 1. Concentrations of both species both in the water column and epiphytic on *H. scoparia*, recorded during 2010, 2011, 2012, 2013, 2014 and 2015 summer seasons are plotted in Fig. 1 and 2.

Ostreopsis cf. ovata reached the highest abundance in the water column (51 719 cells L⁻¹), as well 170 as on macrophytes (2 289 100 cells g⁻¹ fwm), during summer 2015, which was characterized by 171 notably good weather and calm sea conditions. In fact, in summer 2015 the average seawater 172 173 temperature (25.9 °C) and percentage of days with calm sea (75 %) were higher than in the other years. On the contrary, the lowest peak concentration in the water column was recorded in 2011 (4 174 770 cells L-1), in conjunction with only 42.3 % of days with calm sea and quite low average 175 seawater temperature (24.7 °C). Lowest abundance on macrophytes was recorded in summer 176 2014, which was again characterized by comparatively low temperatures (23.8 °C) and unstable 177 178 sea conditions (authors observation). Alarm threshold of this species (> 30 000 cells L⁻¹) was exceeded on three days during 2010, one day in 2012, 4 days in 2013 and one day in 2015. 179

Also *P. lima* achieved the highest concentrations in 2015, reaching 530 cells L⁻¹ in the water column and 74 176 cells g⁻¹ fwm. Conversely, lowest maximum densities in the water column occurred in summers 2010 and 2013 (141 and 115 cells L⁻¹ respectively). Threshold was exceeded only once in 2015.

184 In general, taking into account both species, no relevant health problems nor impacts on the 185 marine fauna were recorded in the studied area during the years considered.

186 <u>3.2. Proliferation dynamics and ecological behavior</u>

Significant correlations between cells in the water column and on macrophytes were reported for both species (p < 0.001 for both *O*. cf. *ovata* and *P*. *lima*) but the *O*. cf. *ovata* model was notably stronger ($R^2 = 0.79$) than the one for *P*. *lima* ($R^2 = 0.15$). Cell densities recorded with calm or rough sea conditions (see MM section for definition) in the water and on the macrophytes did not show 191 species specific patterns according to sea conditions (Fig. 3), although species specific effects of the hydrodynamic regime can be inferred in terms of affinity of microalgal cells with the benthic 192 193 substrate (Fig. 4). In particular, the ratio between planktonic and benthic cells of O. cf. ovata is more variable in both calm and rough sea conditions, suggesting that its cells more autonomously 194 move from the substrate to the water column (regardless of seawater hydrodynamics). On the 195 other hand, data regarding P. lima show lower ratio between planktonic and benthic (so less 196 197 specimens occur in seawater relative to concentration on the benthic substrate), no sea condition effects and lower variability within sea conditions (calm vs rough), suggesting a strictly "benthic 198 behavior" and a high degree of attachment to the benthic substrate. 199

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201 4. Discussion

The epiphyte dinoflagellate assemblage constituted by Ostreopsis spp., P. lima, C. monotis is a 202 203 well known group of species, frequently co-occurring in the Mediterranean Sea (Tindall & Morton 204 1998; Vila et al., 2001). While Ostreopsis species have been more regularly monitored in different 205 Mediterranean areas and data regarding O. cf. ovata in terms of cell abundances and distribution (Mangialajo et al, 2011, Penna et al., 2010), ecology (Accoroni et al., 2012; Asnaghi et al., 2012; 206 Carnicer et al., 2015), toxins content (Dell'Aversano et al., 2014) and toxic effects (Giussani et al., 207 2016 and references within) are becoming more conspicuous, the other species still need to be 208 209 more comprehensively addressed.

In the framework of the present study, a 6-year time series of summer monitoring data on Ostreopsis cf. ovata and Prorocentrum lima, both epiphytic on macroalgae and suspended in seawater, and related environmental variables (in particular seawater temperature and hydrodynamic conditions) were collected, showing that summer proliferations of O. cf. ovata and P. lima commonly occur along the Genoa coast (Ligurian Sea, Italy).

215 O. cf. ovata showed a clear intra-seasonal trend, comparable among the different years and similar 216 for cells sampled in seawater and on macroalgae, with peaks of abundance recorded between mid July and the beginning of August, as also previously observed in other NW-Mediterranean areas 217 218 (Mangialajo et al., 2011). In years characterized by a steep increase in seawater temperature, with heat peaks after the first-half of July, proliferation trends were more evident compared to years (i.e. 219 220 2011 and 2014) when seawater temperatures remained lower and more constant along the summer season (Fig. 1 and 2), confirming the relevant role of seawater temperature rise in driving 221 bloom events, as already highlighted by laboratory (Manganelli et al., 2007; Graneli et al., 2008), 222 223 field (Mangialajo et al., 2008; Totti et al., 2010; Cohu et al., 2013) and model (Asnaghi et al., 2017) 224 studies.

On the contrary, *P. lima* did not show any particular pattern related to seawater temperature, neither for cells recorded in seawater nor on macroalgae. Cells on macroalgae and in seawater showed quite stable abundances during the sampling periods across the considered years, without any particular detectable intra-seasonal trend (Figs. 1 and 2).

229 Similar findings on the relative abundance of the two species are provided by another monitoring 230 study performed throughout the whole year along the Catalan coast from July 1997 to July 1998 (Vila et al., 2001). This study showed that higher densities were reached during warmer months. In 231 232 particular, maximum recorded abundance of *P. lima* was around 700 cells g⁻¹ fwm (and only few 233 cells in seawater) but the dominant species was always Ostreopsis sp. in all investigated media 234 (seawater, macrophytes and sand), which reached the highest density of 596 000 cells g⁻¹ fwm in July 1997. On the contrary, in a recent study reporting a massive proliferation of P. lima (about 45 235 236 000 cells g⁻¹ fwm) during summer 2011 along French Catalan coast and Languedoc, Ostreopsis spp. was not found (Blanfuné et al. 2015). Higher abundances of the Prorocentrum group and a 237 slightly delayed proliferation period was reported by a two-year monitoring study performed along 238 the North Aegean Sea (Aligizaki et al., 2009), which carried out a monthly survey from August 239 2003 to December 2005. "P. lima species-complex" was the most common epiphytic species 240 241 (among Prorocentrum genus), reaching relatively high densities both on macrophytes (133 167 cells g⁻¹ fwm on Cymodocea nodosa in November 2005) and in seawater (11 280 cells L⁻¹ at the 242 end of October 2003). This species-complex was recorded all year round, with a temperature 243 range from 10.0 to 29.5 °C, but maximum abundances occurred during late autumn. In the same 244 study, other dinoflagellate toxins producers belonging to the Prorocentrum group were reported to 245 246 bloom during the summer season: P. levis (Faust et al. 2008; up to 918 cells g-1 fwm in late June or early July) and P. borbonicum (Ten-Hage et al. 2000; up to 942 cells g⁻¹ fwm in August -247 September), both found almost only on benthic samples. 248

The alarm threshold range considered for *O*. cf. *ovata* (10 000 - 30 000 cells L⁻¹) was exceeded in several occasions during the study period (Table 1), with the exception of 2011 and 2014, while *P*. *lima* exceeded the set threshold only once (500 cells L⁻¹) in 2015. It is noteworthy that in 2015, an actual much larger bloom of *P*. *lima* on macroalgae occurred (average cells g⁻¹ fwm around 10 fold higher than in the other years), but such remarkable increase was not equally mirrored by the concentration in seawater.

The different ratio between concentration in seawater and on macroalage displayed by the two microalgae (Fig. 4) suggests a quite different behavior of the two species, independently from hydrodynamic conditions. In fact, although both species are benthic, *O. cf. ovata* is much more prone to move from the substrate to seawater, and a general good correlation between the two compartments has been observed in the present study (Fig. 3a) as well as in previous ones (Magialajo et al., 2011). As a consequence of this behavior, setting alarm threshold for the cell concentration in seawater for *O. ovata* is actually meaningful, because the concentration in seawater is closely mirroring the one on macroalgae. Differently, for *P. lima* the threshold set in seawater seems not to be representative of the actual proliferation occurring on the substrate, suggesting the need for the definition of targeted benthic thresholds.

Similarly, *P. lima* behavior in the Gulf of Maine (Northeastern USA coast), has been described emphasizing its characterization as an epiphytic or benthic dinoflagellate, and strengthening the argument that monitoring techniques designed for the water column would be ineffective in this case (Maranda et al. 2007).

It is noteworthy that *P. lima*, given its strictly benthic behavior, is expected to be less dangerous for human health, in terms of accidental ingestion or inhalation of aerosols; nevertheless, it could become an issue for aquaculture activities due to the high density that it can reach in fouling macroalgal biomass around mollusc lines (Maranda et al. 2007a). Additionally, it is noteworthy that the toxin level is not always correlated with the mere number of algal cells (Maranda et al. 2007b; Vannucci et al. 2010).

An appropriate sampling strategy, specifically designed for benthic species (e.g. collection of natural or artificial substrates), is necessary in order to accurately assess potential threatening effects of *P. lima*, more than for other benthic microalgae, such as *Ostreopsis*.

Providing a six-year time series of the co-occurrence of *P. lima* and *O.* cf. *ovata* and their ecological behavior in relation to environmental variables, the present study highlights the particular need to implement a specific sampling strategy and define benthic alarm thresholds for strictly benthic microalgae, such as *P. lima*, triggering future applications in a management perspectives.

283

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490 Figure legends

Figure 1. Densities (± sd) of *O*. cf. *ovata* (black) and *P. lima* (grey) in the water column and on the
macrophyte *H. scoparia* during summer seasons 2010, 2011 and 2012 at Quarto dei Mille, Genoa
(Italy). Dotted line indicates seawater temperature. Dark-gray shade highlights *O*. cf. *ovata*threshold range and light-gray dotted line indicates *P. lima* threshold.

Figure 2. Densities (± sd) of O. cf. *ovata* (black) and *P. lima* (grey) in the water column and on the
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(Italy). Dotted line indicates seawater temperature. Dark-gray shade highlights *O. cf. ovata*threshold range and light-gray dotted line indicates *P. lima* threshold.

- Figure 3. Correlation between *O*. cf. *ovata* (left; p < 0.001 and $R^2 = 0.79$) and *P*. *lima* (right; p < 0.001 and $R^2 = 0.15$) cell concentrations in seawater (cells L⁻¹) and on macrophyte (cells g⁻¹ fwm). Circles refer to data collected in calm sea conditions, triangles to wave height> 0.5 m.
- 502 Figure 4. Boxplots of ratio between cells in seawater and on macrohyte (O. cf. ovata on the left; P.
- lima on the right) in calm and rough (< or > 0.5 m high wave; see MM section) sea conditions, during the blooming periods (1st of July – August 15th of 2010, 2011, 2012, 2013 and 2015).