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- 1 The impact of ocean acidification on the gonads of three key Antarctic benthic
- 2 macroinvertebrates
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12 Contributors

- 13 OD carried out the experiment, collected data, performed the analyses and wrote the paper;
- 14 SF performed the histological assay and analyses;
- MC and VA planned the experimental design and revised the manuscript, approving the submitted
 version

17 Abstract

CO₂ atmospheric pressure is increasing since industrial revolution, leading to a lowering of 18 19 the ocean surface water pH, a phenomenon known as ocean acidification, with several reported 20 effects on individual species and cascading effects on marine ecosystems. Despite the great 21 amount of literature on ocean acidification effects on calcifying organisms, the response of their reproductive system still remains poorly known. In the present study, we investigated the 22 histopathological effects of low pH on the gonads of three key macroinvertebrates of the Terra 23 Nova Bay (Ross Sea) littoral area: the sea urchin Sterechinus neumayeri, the sea star Odontaster 24 25 validus and the scallop Adamussium colbecki. After 1 month of exposure at control (8.12) and reduced (7.8 and 7.6) pH levels, we dissected the gonads and performed histological analyses to 26 27 detect potential differences among treatments. Results showed significant effects on reproductive conditions of A. colbecki and S. neumayeri, while O. validus did not show any kind of alteration. 28 29 Present results reinforce the need to focus on ocean acidification effects on soft tissues,

- 30 particularly the gonads, whose damage may exert large effects on the individual fitness, with
- 31 cascading effects on the population dynamic of the species.
- 32 Keywords: ocean acidification, histopathology, gonads, macrobenthos, Antarctica

33 **1. Introduction**

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The rising of atmospheric CO₂ concentration, since industrial revolution, has already caused the drop of ocean surface water pH of about 0.1 unit, a process known as ocean acidification (OA, Caldeira and Wickett 2003). OA is expected to exert a large number of impacts on marine organisms and ecosystems, but pathways and the extent of these impacts are still poorly understood, since organism response varies across species and even between life stages of the same species (e.g. Ries et al. 2009; Hendriks et al. 2010; Kroeker et al. 2010).

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The Antarctic marine environment is supposed to be one of the most threatened by rising atmospheric CO₂, because of the peculiar seawater physico-chemical parameters (McNeil and Matear 2008; Fabry et al. 2008). Antarctic invertebrates, in turn, are not expected to successfully cope with current climate change pattern, occurring at a faster rate than ever (Meredith and King 2005; Hofmann and Todgham 2010), because of their low metabolism (Peck 2002). Their response to climate change impacts still needs to be fully understood and an in deep investigation on polar species and ecosystems is strongly necessary.

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Although shell and skeleton dissolution are the most direct (and most studied) effects of OA, 50 51 several other studies have focused on different aspects, such as survival, metabolic rate, growth, etc. (e.g. Michaelidis et al. 2005; Widdicombe and Spicer 2008; Wittmann and Pörtner 2013). Yet, 52 53 despite this huge increase of knowledge, one aspect still remains poorly investigated: the effect of OA on the reproductive system. In fact, while a large number of papers deals with fertilization and 54 embryonic and larval development (e.g. Kurihara et al. 2008; Byrne 2012; Van Colen et al. 2012; 55 Byrne and Przeslawski 2013; Barros et al. 2013), so far only three papers have investigated the 56 effects of OA on the gonads (Kurihara et al. 2013; Uthicke et al. 2014 and Mos et al. 2016), finding 57 58 alterations in coelomic fluid ion compositions, reduced energy intake (leading to a delay in gonad 59 maturation) and a general poor condition in the gonads of specimens exposed to low pH.

61 In order to investigate the reproductive system, histological techniques are very useful to detect sublethal effects and histopatology falls in the middle of the relationship 'response time' vs 62 'ecological significance' (Thiéry et al. 2012), allowing to detect effects also in relatively short-term 63 experiments (in the range of few weeks). In fact, the histological survey of gonad tissues is often 64 used to detect the welfare of organisms both in field and in laboratory conditions (e.g. Vaschenko 65 et al. 1997; Lehmann et al. 2007; Martinez et al. 2014; Smaoui-Damak et al. 2006; Ortiz-Zarragoitia 66 67 and Cajaraville 2010). In this view, gonads are considered a 'sentinel' organ, since they are commonly the first organs to be affected by energy re-allocation in case of organism stress, when 68 69 a trade-off occurs between reproduction and somatic maintenance. This may be aggravated in calcifying organisms, which reserve a high energy percentage to the repair and maintenance of 70 71 hard tissues (Sokolova et al. 2012; Haag et al. 2016).

In the present study, we investigated the potential effects of OA on the gonads of three macrobenthic invertebrates, key species in the Antarctic littoral ecosystem: the sea urchin *Sterechinus neumayeri*, the sea star *Odontaster validus* and the scallop *Adamussium colbecki*. These species are endemic of High-Antarctic areas, circumpolar and locally very abundant, playing an important role in the littoral ecosystem of Terra Nova Bay (Ross Sea, Chiantore et al., 2002).

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78 S. neumayeri and O. validus are the most abundant echinoderms in shallow waters around Antarctica. These two echinoderms are generalist feeders, though S. neumayeri preferably feeds 79 on benthic algae and detritus, while O. validus is mainly a scavenger and predator (McClintock 80 81 1994; Brey et al. 1995; Gillies et al. 2012; Taboada et al. 2013). Both species are dioecious and 82 display a gametogenic cycle of 18-24 month, at least for females, with overlapped maturation of 83 two cohorts of oocytes (separated by a year), simultaneously present in their gonads (Pearse 1965; Pearse and Giese 1966). Because of their abundance and trophic role, S. neumayeri and O. 84 85 validus are considered key species in the shallow water benthic ecosystem (McClintock et al. 1988; McClintock 1994; Brey et al. 1995) and effects of climate change on these two echinoderms could 86 87 lead to dramatic community shifts.

88

The Antarctic scallop, *A. colbecki* represents a key link in the bentho-pelagic coupling (Chiantore et al. 1998) and a food source for higher trophic levels (Vacchi et al. 2000; Dell'Acqua et al. 2017). Despite being patchily distributed (Schiaparelli and Linse 2006), it can be locally very abundant,

such as in Terra Nova Bay (Chiantore et al. 1998), where A. colbecki 'beds' play a relevant role in
the coastal organic carbon flux and CO₂ sequestration in the shells.

S. neumayeri fertilization, embryo and larval development have been widely investigated in 94 regard to OA, showing an overall robustness, also when warming acts as a co-stressor (e.g. Ericson 95 96 et al., 2012; Collard et al., 2013; Morley et al 2016). So far, only one study has dealt with the effects of OA on O. validus, investigating its larval development and has found a decline in larval 97 98 survival at pH 7.6, together with a significantly decreased body width (Gonzalez-Bernat et al., 2013). The same stands for A. colbecki, for which only one study dealing with OA effects is 99 100 available: Benedetti et al. (2016) found a slight to moderate effect of low pH, but only in concert with other stressors (metals and/or warming). 101

The aim of the present study is to assess the effects of OA on gonadal development and the histological features of gonadal tissue of these three key Antarctic macroinvertebrates, in order to further elucidate benthic responses of animals in the Southern Ocean where near-future ocean acidification is predicted to be particularly pronounced (Fabry et al. 2009).

106 2. Material and methods

107 2.1. Specimens and field data collection

The experiment was performed at the Italian Mario Zucchelli Station (MZS), during the 108 2014-2015 Italian Antarctic Expedition. Specimens of A. colbecki, S. neumayeri and O. validus were 109 collected by SCUBA divers on 10th December 2014, at around 15 m depth in Tethys Bay (74° 110 41.407' S; 164° 06.311' E), about 2 km from MZS. Specimens were immediately transported to the 111 112 aquarium facility of the base and the shells of A. colbecki specimens were cleaned from epibionts 113 by gentle brushing. All the specimens were acclimated in two (O. validus, as predator, was kept 114 alone) 100 L refrigerated (-0.5°C) aquaria supplied with flow-through unfiltered seawater pumped from the water intake located in front of the station (at 6 m depth). 115

Littoral seawater parameters (temperature, salinity, and pH) were periodically measured using a CTD probe (Ocean Seven 310 CTD - Idronaut, Brugherio, Italy; Table 1), from 10th December 2014 to 27th January 2015, always at the same time of the day (10:30 to 12:00), both at the site of the water intake and at the site where the investigated specimens were collected.

120 2.2. Experimental set up

After a two-week period to allow the individuals to adjust to the laboratory conditions, the experiment started using a flow through system, in order to avoid O₂ depletion and uncontrolled 123 pH reduction due to respiration of the animals. Seawater was first pumped to a 100 L aquarium, 124 where the water was cooled to controlled -0.7°C, and afterwards to three 50 L header tanks, two of which were set as treatment and one served as a control (unmanipulated water). Two nominal 125 pH treatment levels, 7.8 and 7.6, were chosen according to literature (Cao et al., 2007; Gonzalez-126 127 Bernat et al., 2013) and were reached by bubbling pure CO₂ in two of the three header tanks. The addition of CO₂ was made through two independent electronic valves, in connection with two pH 128 electrodes set on 7.8 and 7.6 and regulated by a continuous pH-stat system (IKS Aquastar, 129 Karlsbad, Germany). A third pH electrode was mounted on the pipe that fed the control header 130 131 tank in order to monitor the natural littoral seawater pH.

Eight specimens per species were subjected to each pH level: in order to control for a 132 potential 'tank effect', the individuals of each species were equally divided between two tanks for 133 each pH level. Consequently, we set up six tanks (experimental units or replicates; 20 x 25 x 20 cm) 134 135 for each pH level (2 for each species). Each tank contained 4 specimens for a total of 24 individuals per species across the different tanks. The specimens for experimental exposure were randomly 136 137 selected, with the only restriction of the size range in order to perform the experiment on adult 138 specimens only. For each species, means and standard deviations of the 24 individuals were 139 calculated, resulting in 76 ± 3 mm shell height (SH, the distance from the umbo to the opposite 140 side of the shell) for A. colbecki, 37 ± 3 mm test diameter (without spines) for S. neumayeri and 43 141 \pm 3 mm in mean distance from center of disc to arm tip (R) for *O. validus*.

142 Each experimental unit was directly fed, by gravity, from one of the header tanks with an individual pipe, regulated manually by PVC valves at a rate of 150 ml/min, which assured a water 143 144 renewal rate of 60%/h in each experimental unit. After filling the experimental units, the water 145 was discharged through a relief hole. All the tanks were covered with transparent lids to avoid gas exchange and specimen escape. The cover was only removed every 2-3 days for about 5 minutes, 146 147 in order to clean the tank bottom by siphoning. The organisms were not fed, but, since unfiltered seawater was taken in, diatoms and other detrital material were available in the water. This 148 system, of course, could supply food for the filter feeding A. colbecki and, partially, the generalist 149 S. neumayeri, but could not meet the food demand of O. validus. Yet, both the sea urchin and the 150 151 sea star can stand long periods of time without food with no apparently ill effects (Brockington et 152 al. 2001; Agüera et al. 2016).

153 The pH electrodes of the pH-stat system were intercalibrated every 3-4 days on total scale 154 (pH_T) using TRIS buffer solutions with a salinity of 35 psu (Dickson et al., 2007) and cross checked 155 against the pH_T values measured with two different multiprobes that were previously calibrated 156 on Antarctic littoral waters: Ocean Seven 310 CTD and C6 Muti-sensor Platform (Turner Design, San Jose, CA, USA). Temperature, salinity and pH were recorded from each header tank with 157 multiprobes throughout the experimental period, while triplicate seawater samples for total 158 159 alkalinity (TA) measurements were collected once a week, poisoned with HgCl₂ and stored at +4°C. In Italy, TA was determined at the Polytechnic University of Ancona (Italy) using an open cell 160 161 potentiomentric titration according to Dickson et al. (2007) procedures and standards. Temperature, salinity, pH_T and TA were used to calculate pCO_2 of the three pH levels with 162 163 SWCO2_V2 (http://http://neon-old.otago.ac.nz/research/kah/software/swco2/index.html) software, using the equilibrium constants of Millero et al. (2006), since the lowest value of their 164 temperature range (0 – 40°C) is close to Antarctic waters. In addition, further duplicate seawater 165 samples were periodically taken downstream from each header in order to measure nutrient 166 concentration. The experiment lasted from 26th December 2014 to 28th January 2015 (34 days). At 167 the end of the experiment, all specimens were measured, weighed and dissected. Gonads were 168 169 dissected, weighed and fixed separately for each specimen in Bouin solution, then rinsed in 170 ethanol 70% and stored at +4°C in order to return the samples to Italy. The gonado-somatic index 171 (GSI) was assessed as follows:

172 GSI = gonad ww (g) * 100/body ww (g)

173 ww = wet weight.

174

175 2.3. Sample processing and statistical analyses

176 Once in Italy, the gonads were further dissected for the analysis. For S. neumayeri and O. 177 validus, whose gonads are constituted of five portions, each of them was split into 2 halves; 5 out of the 10 halves for each specimen (24 sea urchins and 24 sea stars) were used for histology. In 178 total, we analyzed 120 gonad portions for each species. For A. colbecki, we used three 179 uncontiguous pieces from the middle of the gonad of each individual specimen (24 individuals), for 180 181 a total of 72 gonad slices for this species. The selected portions were embedded in paraffin and cut with a microtome in order to obtain 6 µm thick sections on microscope slides. Afterwards, 182 Hematoxylin-Eosin staining (Bio Optica Spa, Milan, Italy) was performed, mounting the cover slip 183 with Eukitt (Kindler GmbH, Freiburg, Germany) and letting the slides dry for 24 hours. Since, 184 185 preliminarily, a large mucus production in the gonoduct epithelium of A. colbecki specimens was noticed, additional slides were prepared (only for this species) to be stained with Alcian Blue-P.A.S 186

reaction for mucopolysaccharides (Bio Optica Spa, Milan, Italy). Sections were examined with a Leica DMRB light microscope and images were acquired with a Leica CCD camera DFC420C. Pictures were acquired at 10x and 20x magnification, covering at least 80% of the surface. Such high coverage of the gonad histological survey allowed for a reliable estimation of the potential damages. As far as the mucus production analysis, pictures of the Alcian Blue – P.A.S. stained slides were acquired, covering at least 80% of the total duct epithelium.

The gametogenic stage (GS) of the dissected specimens was assessed following Pearse and Giese (1966) for *S. neumayeri*, Pearse (1965) for *O. validus* and Berkman et al. (1991) for *A. colbecki*. Since during summer the three investigated species are at the end of the gametogenesis (*S. neumayeri*, Pearse and Giese 1966) or in maturation (*O. validus* and *A. colbecki*, Pearse 1965; Chiantore et al. 2001), no gonads in recovery stage were observed. The assigned stages were: early maturation = 1, advanced maturation = 2, ripe = 3, initial spawning = 4, advanced spawning/partially spent = 5, spent = 6.

Given the fundamental lack of reported knowledge on specifically OA effects on gonads, 200 we decided to compare our histological evidences with considered published baseline of 'well-201 202 state' gonads of sea urchins, scallops and sea stars (e.g. Belkhedim et al. 2014; Zheng et al. 2014; 203 Baeta et al. 2016), including the species here investigated (Pearse 1965; Pearse and Giese 1966; 204 Berkman et al. 1991). Thanks to literature on altered gonads because of other stressors, such as hydrocharbons (Schäfer and Köhler, 2009) and temperature (Delorme and Sewell, 2016), we were 205 able to ascribe gonads (or parts of) to 'good' or 'poor' conditions. We recorded any sign of 206 207 'anomalies' in our specimens and we computed the ratio of the total altered area on the total 208 photographed area (anomaly ratio). For mucus analysis, we computed the ratio between the 209 length of the membrane with mucus overproduction and the total photographed membrane length. This ratio also allowed to pool females and males in the statistical analysis to provide 210 211 statistical robustness.

The Kruskal-Wallis test (hereafter KWt), performed using the R software package PMCMR (Pohlert, 2014; R Core Team, 2013), was applied to assess differences between treatments. Unfortunately, KWt does not allow for nested design, but it was selected because of its robustness for such small sample size. The KWt test was applied on the tank level and not on the pH treatment level, according to Cornwall and Hurd (2016). In case of significance, the Nemenyi-Damico-Wolf-Dunn test for post-hoc comparisons among tanks was performed. The tested response variables were: GSI, GS and anomaly ratio.

219 3. Results

220 3.1. Field and aquaria saeawater variables

221 The experimental set up, including the acidification system, provided stable seawater 222 parameters for the whole experimental duration. Range of littoral seawater variables, at 6 (water 223 intake) and 15 m (collection site) are displayed in Table 1. The cooling system in the aquarium provided an experimental temperature range in agreement with the natural variability at the 224 collection site of the treated specimens (Tab. 2). Both temperature and pH values in the littoral 225 226 waters changed over the season (Tab. 1), so that the experimental treatment 7.8 actually fell in the range of the natural field variability. The trend of the nutrient concentration along the 227 228 experimental period is displayed in Fig. 1. The flow-through system assured the nutrients and 229 cathbolites to be maintained under a potential toxic threshold (Suckling et al., 2015). A seasonal 230 trend for nutrient concentrations was observed, as expected for the period, concurrent with the 231 algal bloom: on 10th January all the nutrients reached the lowest levels, few days after the pack-ice 232 break (pers. obs.).

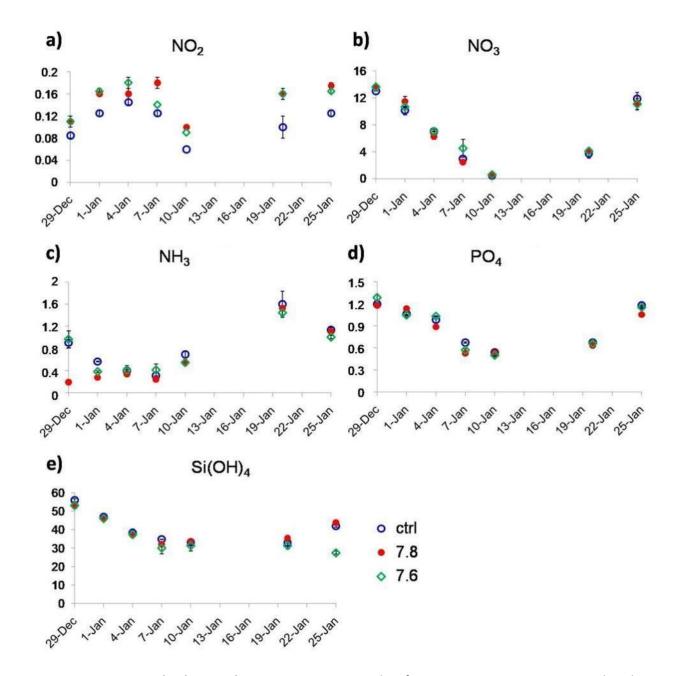
Tab. 1. Field seawater variables. Means and standard deviations of the littoral seawater variables
 recorded at the water intake site (6 m) and at the specimen collection site (15 m) during the
 experimental period.

Site	Date	Temperature (°C)	Salinity (psu)	рН
	10 th Dec 2014	-1.7 ± 0.04	34.55 ±0.01	7.78±0.00
	15 th Dec 2014	-1.42 ± 0.01	34.72± 0.00	7.88± 0.00
	18 th Dec 2014	-0.59 ± 0.01	34.58 ± 0.00	7.99 ± 0.00
	20 th Dec 2014	-0.54 ± 0.01	34.60 ± 0.01	8.00± 0.01
Water intake (6	29 th Dec 2014	-0.81 ± 0.03	34.41 ± 0.01	8.05 ± 0.01
m)	2 nd Jan 2015	-0.18 ± 0.01	34.52 ± 0.01	8.05 ± 0.00
	23 rd Jan 2015	0.10 ± 0.25	33.30 ± 0.30	8.04 ± 0.02
	24 th Jan 2015	0.34 ± 0.00	33.00 ± 0.00	8.06 ± 0.00
	27 th Jan 2015	0.42 ± 0.05	33.66 ± 0.04	8.02 ± 0.01
Collection site	10 th Dec 2014	-1.75 ± 0.00	34.71 ± 0.00	7.78 ± 0.00
(15 m)	15 th Dec 2014	-1.46 ± 0.00	34.74 ± 0.00	7.86 ± 0.00
(13 m)	19 th Dec 2014	-0.58 ± 0.01	34.64 ± 0.00	7.98 ± 0.00

20 th Dec 2014	-0.62 ± 0.01	34.63 ± 0.00	7.95 ± 0.00
29 th Dec 2014	-0.64 ± 0.02	34.54 ± 0.01	7.99 ± 0.00
2 nd Jan 2015	-0.23 ± 0.02	34.59 ± 0.00	8.00 ± 0.00
23 rd Jan 2015	-0.25 ± 0.00	33.69 ± 0.00	8.01 ± 0.00
24 th Jan 2015	-0.30 ± 0.04	34.28 ± 0.03	8.02 ± 0.00
27 th Jan 2015	-0.16 ± 0.04	34.17 ± 0.02	8.01 ± 0.00

Tab. 2. Experimental seawater variables. Means and standard deviations of the seawater
 variables recorded in the three treatments throughout the experiment, from 26th December 2014
 to 28th January 2015. pCO₂ are calculated with SWCO2 software, using equilibrium constants from
 Millero et al. (2006).

Nominal treatment	Temperature (°C)	Salinity (psu)	рН	TA (μmol/kg _{sw})	pCO2 (ppm)
control pH	-0.395 ± 0.12	33.06± 0.02	8.12 ± 0.05	2208.5± 75.7	381.7 ± 31.2
7.8	-0.362 ± 0.09	33.05± 0.01	7.814 ± 0.02	2238.3± 118.2	730.3 ± 35.5
7.6	-0.369 ± 0.12	33.09 ± 0.02	7.625 ± 0.02	2254.6± 89.2	1085.4 ± 55.9



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Fig. 1. Nutrient trends during the experiment. Trends of nutrient concentrations in the three
 treatments along the experimental period. Empty circles indicate control treatment, filled circles
 represent pH 7.8 and diamonds pH 7.6. Sampling date is reported on *x*-axis and nutrient
 concentration (µmol/l) on *y*-axis. (this figure is 1.5-column fitting image and it should be in color in
 the online version only)

248 3.2. Histological assay

Overall anomalies found in the experimental specimens with respect to treatment are displayed in Tab. 3. Single species response to OA treatments are described in the following sections. Tab. 3. List of histological anomalies. Observed anomalies in the gonads associated to the
nominal treatment and number of individuals in which the given type of anomaly was detected.
Some individuals displayed more than a single type of anomaly. Percentage of individuals is
calculated, for each pH, as number of specimens displaying anomalies on a total of 8 individuals.

Species	рН	Alteration	% ind
A. colbecki	Ctrl	Leakage of germinal tissue outside the acinus	
	7.6	Disruption of the acinus basal membrane, with concurrent leakage of germinal tissue outside the acinus	75
		Fibrosis	25
		Disruption of the trabeculae and connective tissue	
		Hemocyte infiltration	
		Neoplasia (overgrowth of undifferentiated germinal cells)	12.5
	Ctrl	Spermatozoa detached flagella (head missing)	12.5
		Spermatozoa detached flagella (head missing)	12.5
		Intense desquamation of ovary sac tissue, including the	62.5
	7.8	germinal tissue, and concurrent fibrosis	
	7.8	Hermaphroditism	12.5
		Atretic oocytes	25
		Neoplasia (overgrowth of undifferentiated germinal cells)	12.5
		Spermatozoa detached flagella (head missing)	12.5
		Intense desquamation of ovary tissue, including ed the	
S. neumayeri		germinal tissue, and concurrent fibrosis	
		Hermaphroditism	12.5
		Neoplasia (overgrowth of undifferentiated germinal cells)	
	7.6		25
	7.0		
		Disruption of acinus basal membrane, with concurrent	
		disorganization of the germinal tissue (that lost the radial	37.5
		architecture) and fibrosis	
		Diffuse lipofuscin-like pigments (LPP, Vaschenko et al. 2012)	12.5

O. validus	7.9	Desquamation of follicle tissue with concomitant hemocytes	
	7.0	infiltration	

256 3.2.1. Adamussium colbecki

Kruskal-Wallis test for GSI provided significant results (p = 0.043) and Dunn Test for posthoc comparisons indicates significant higher values for the pH 7.6 tanks compared to all the other treatments, except ctrlb, that significantly differs from ctrla (Fig. 2). Both ctrl and pH 7.8 tanks accounted for a gonad stage ranging from 1 (early maturation) to 3 (ripe gonads, only two specimens), while pH 7.6 showed a higher mean value (4, initial spawning), as 6 specimens on 8 were in spawning and the others accounted for ripe gonads. KWt and Dunn Test confirmed these observations, being *p*-value = 0.015 due to the tanks in pH 7.6.

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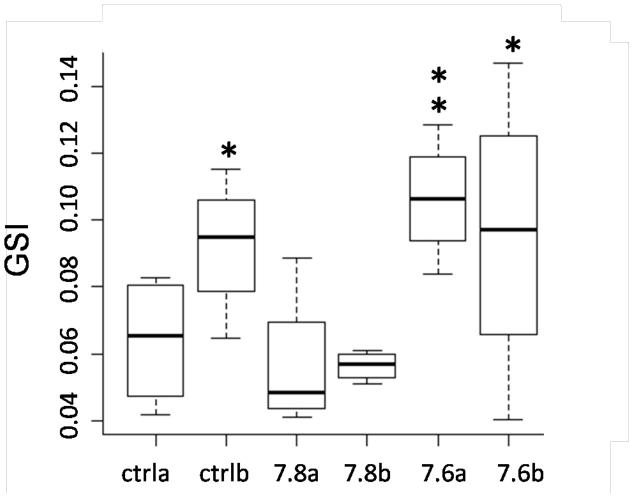


Fig. 2. Adamussium colbecki GSI vs tank level. Boxplot showing the difference, among tanks, in
the gonado-somatic index (GSI) of A. colbecki. Middle line: median; boxes: 25-75 percentiles;
whiskers: lowest and highest data points. A. colbecki GSI values on y-axis; tank level on x-axis: the

nominal pH is indicated by 'ctrl', '7.8', '7.6', while the letters 'a' or 'b' refer to the tank replicates.
Boxes with the same number of asterisks are statistically not different. N = 4 per individual tank.

271 (this figure is a single column fitting image)

The mucus presence in the gonoducts epithelium did not show significant differences among tanks (p = 0.650), as well as the occurrence of anomalies ratio (p = 0.060). Despite the lack of significance, we found different types of anomaly in some specimens at pH 7.6, indicating that some kind of stress occurred, although with a large inter-individual variability. Herein, we describe the most representative anomalies, those occurred in 50% or more specimens.

Fig. 3 shows gonads of female and male specimens maintained under control conditions, compared with the ones at pH 7.6, where anomalies were found. Fig. 3a and 3b show gonads in the control specimens that are in very good conditions, either with ripe ovaries or mature acini. Conversely, specimens from pH 7.6 display some level of anomalies: massive hemocyte infiltration (Fig. 3c and 3d, female and male, respectively) and disruption of male follicle basal membrane with leakage of germinal cells (Fig. 3e).

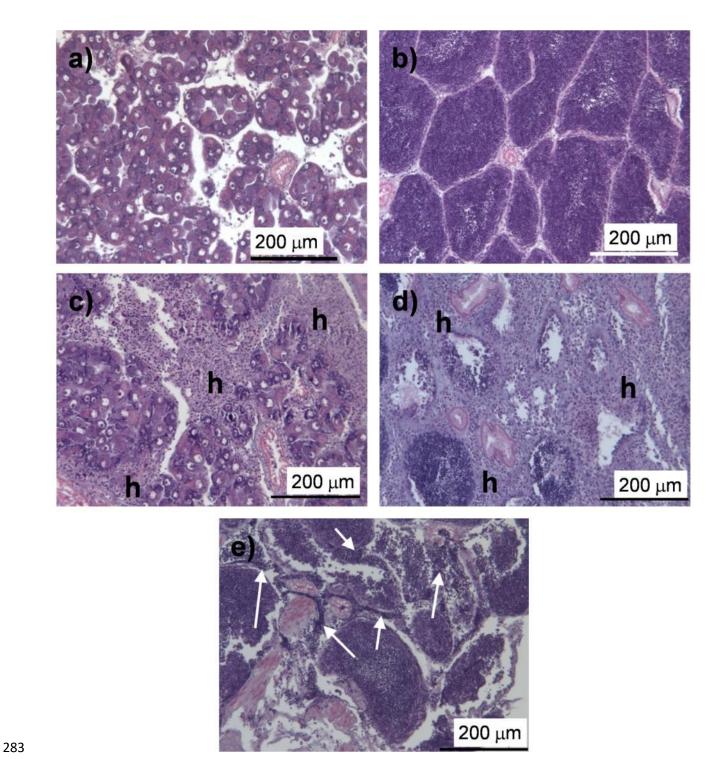


Fig. 3. Adamussium colbecki gonad histology. Representative gonad tissue of *A. colbecki* maintained at control and pH 7.6. (a, b) Normal histology of a female and a male gonad from control pH, with ovaries and testis in advanced maturation. (c, d) Female and male gonads from pH 7.6, showing a massive hemocyte intrusion in the follicles (indicated with 'h'). (e) Gonad of a male specimen from pH 7.6, showing disruption of acinus epithelium and leakage of germinal cells (white arrows) between acini. (this figure is 2-column fitting image and it should be in color in the online version only) Fig. 4 shows two male gonads stained with Alcian Blue-PAS. Mucous components are highlighted by a more intense blue along the gonoduct epithelium. Around half of the specimens maintained at pH 7.6 showed larger mucus production compared to the control specimens, but the difference was not statistically significant

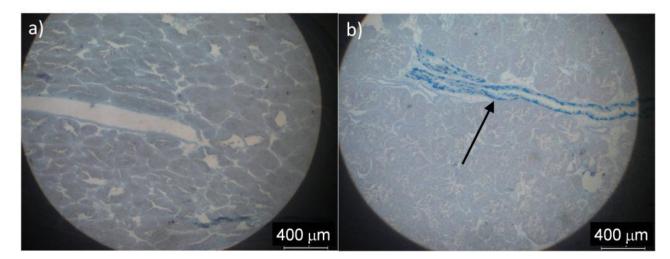
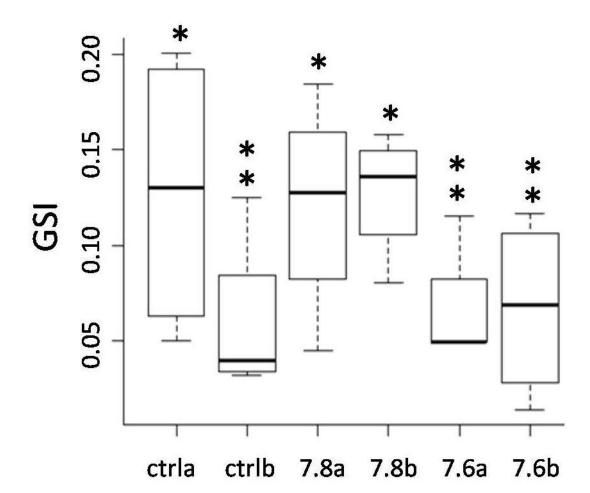


Fig. 4. Adamussium colbecki gonad histological evidence of mucus overproduction. Gonad tissue of *A. colbecki* maintained at control pH (a) and at pH 7.6 (b), stained with Alcian Blue-P.A.S. Mucus production is detectable along the gonoduct epithelium (see the intense blue coloration indicated by the black arrow in b), where the mucous components bind the stain <u>(this figure is 2-column</u> fitting image and it should be in color in the online version only)

301 *3.2.2. Sterechinus neumayeri*

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302 GSI of *S. neumayeri* did not significantly differ among tanks (p = 0.103), with no detectable pattern among tanks either from different or same pH treatment. High individual variability was 303 304 observed, as from the boxes width and the variability among tanks within the same pH (Fig. 5). All 305 the specimens were in advanced spawning or spent stage, showing a gonad stage 5 or 6, with no differences among tanks (p = 0.409). Conversely, the anomaly ratio indicated a tank effect, which 306 is ascribable to the treatment (p = 0.046). In fact, the post-hoc comparison shows that control 307 tanks did not differ, but were statistically different from tanks at pH 7.8 and 7.6, that were, in turn, 308 309 very similar among each other.

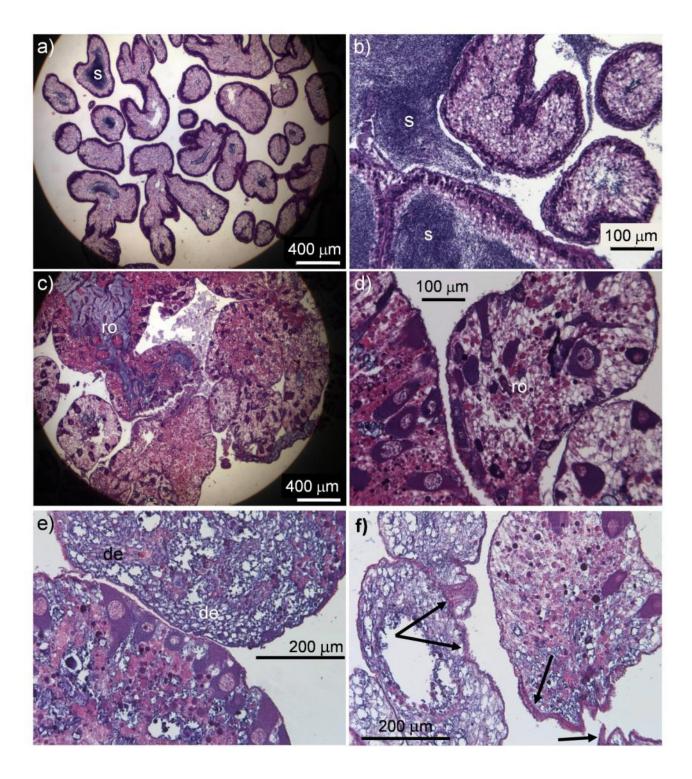


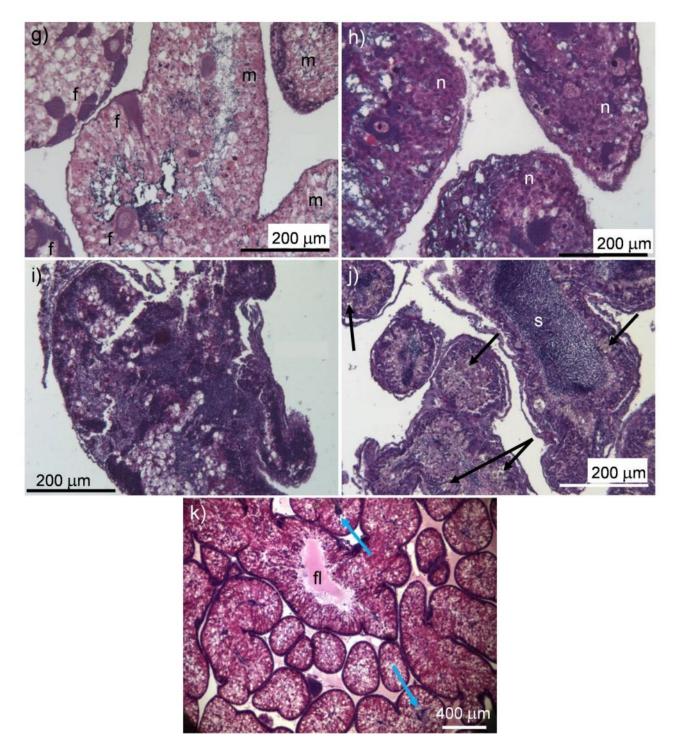
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Fig. 5. *Sterechinus neumayeri* GSI at tank level. Boxplot showing the difference, among tanks, in the gonado-somatic index (GSI) of *S. neumayeri*. Middle line: median; boxes: 25-75 percentiles; whiskers: lowest and highest data points. *S. neumayeri* GSI values on *y*-axis; tank level on *x*-axis: the nominal pH is indicated by 'ctrl', '7.8' or '7.6', while letters 'a' or 'b' refer to the tank replicates. Boxes with the same number of asterisks are statistically not different. N = 4 per individual tank. (this figure is a single column fitting image)

317

318 We found a high variability of anomalies, even between tanks within the same pH. The anomalies that occurred at higher frequency are described as representative; few anomalies, 319 320 occurring at lower frequency, are described in terms of their relevance associated with experimental or environmental stressors in literature. The different anomalies detected in the 321 gonads of low pH treated *S. neumayeri*, compared with the control specimens, are displayed in Fig. 322 6. Panels a, b, c and d show male and female gonads in control specimens, with asynchronous 323 324 ovaries and acini, either in spawning or spent stage. Panels from e to k show different types of anomalies that were found either at pH 7.8 or 7.6, sometimes in both: intense desquamation of 325 326 germinal tissue and concurrent fibrosis (Fig. 6e and 6f), hermaphroditism, with either female, male and miscellaneous follicles (Fig. 6g), neoplasia (Fig. 6h), disorganization of the germinal tissue with
loss of the radial architecture (Fig. 6i), several spread anomalies (including basal membrane
disruption) with the presence of lipofuscin-like pigments (LLP, Vaschenko et al. 2012) throughout
the tissue (Fig. 6j), and sperm flagella detached from the head (Fig. 6k).





333

Fig. 6. *Sterechinus neumayeri* gonad histology. Gonad tissue of *S. neumayeri* maintained at control pH, pH 7.8 or and 7.6. (a, b) Representative histology of a male gonad showing testis either at the end of the spawning stage or spent (s = sperms). (c, d) Representative histology of a female gonad showing ovary sacs either at the end of the spawning stage or spent (ro = relict oocytes). (e, f) Representative female gonads from pH 7.8 and 7.6, respectively, showing an intense desquamation of the germinal tissue (de = desquamation) and fibrosis of the follicle membrane (black arrows). (g) Hermaphroditic specimen from pH 7.8, showing either distinct female and male 341 follicles, as well as miscellaneous follicles (f =female germinal tissue, m = male germinal tissue). (h) 342 Neoplasia (indicated with n) in a female specimen from pH 7.6. (i) Male gonad of a specimen from pH 7.6. (i) Male gonad of a specimen from pH 7.6, displaying a disorganized germinal tissue that 343 lost the radial architecture along with disruption of the acinus membrane. (j) Diffuse several 344 anomalies in a specimen from pH 7.6, including presence of lipofuscin-like pigments (black arrow), 345 s = sperms. (k) Detached flagella of spermatozoa, both inside lumina and between acini (fl = 346 347 flagella, blue arrows =normal sperms). (this figure is 2-column fitting image and it should be in <u>color in the online version only</u>) 348

349 3.3.3. Odontaster validus

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GSI of *O. validus* significantly differed among tanks (p = 0.038); this response seems not to be ascribable to the pH treatment, but, rather to a high variability between tanks, even within the same pH (Fig. 7). The overall GS ranged from 1 to 5, with the majority of the specimens (14 out of 24) in stage 2 (advanced maturation), but no significant difference was detected among tanks (p =0.874). Also the anomaly ratio did not show significant differences among tanks ($p \approx$ 1). The only alteration we found is displayed in Fig. 8, in a single specimen from pH 7.8

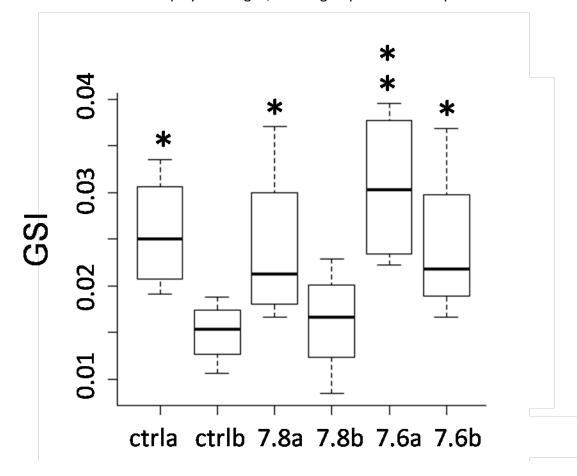
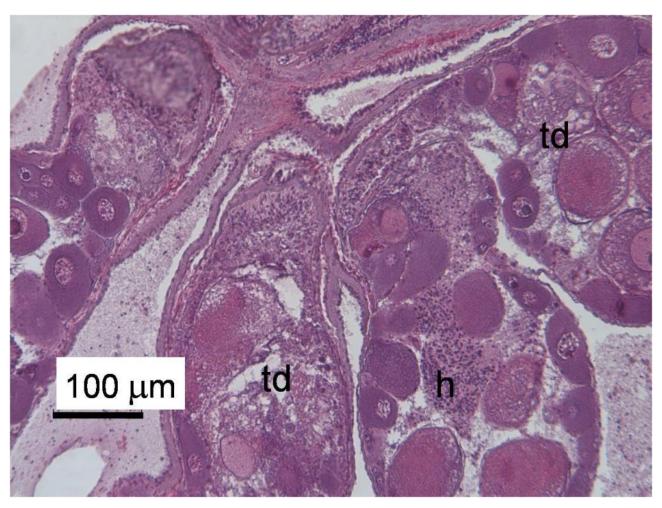


Fig. 7. Odontaster validus GSI at tank level. Boxplot showing the difference in the gonado-somatic index (GSI) of *O. validus* between tanks. Middle line: median; boxes: 25-75 percentiles; whiskers: lowest and highest data points. *O. validus* GSI values on *y*-axis; tank level on *x*-axis: the nominal pH is indicated by 'ctrl', '7.8' or '7.6', while letters 'a' or 'b' refer to the tank replicates. Boxes with the same number of asterisks are statistically not different. N = 4 per individual tank <u>(this figure is a</u> <u>single column fitting image)</u>

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Fig. 8. Odontaster validus gonad histology. Gonad tissue of a female O. validus kept at pH 7.8,
 displaying tissue disruption (td) and hemocyte inflitration (h). (this figure is a single column fitting
 image and it should be in color in the online version only)

369 4. Discussion

The aim of the present study was to provide a comparable investigation on OA response in three key benthic Antarctic species from the same habitat, simultaneously exposed to the same experimental treatment in a relatively short-term experiment. 373 In the natural environment the pH showed a remarkable variation during the short summer 374 season following physical and biological processes: at the beginning of the season, the lower seawater pH mirrors the high respiration processes occurring under the sea-ice and the total 375 absence of gas exchange (Schram et al. 2015). With the progression of the summer season and the 376 377 concomitant seawater temperature rise, the sea-ice bottom probably started to melt and break up, allowing partial gas exchange and the bloom of the sympagic algae. The peak of this process 378 was reached on 10th January, but field measures close to that date were not possible because of 379 380 the sea ice instability. Notably, the nominal pH 7.8 fell in the range of the littoral seawater at Terra 381 Nova Bay during late spring. Basically, the specimens maintained at pH 7.8 experienced the advent of the summer season (phytoplankton bloom clearly visible in the aquarium water, higher 382 383 temperature and lower salinity), but an unchanged surrounding pH.

Overall, in the controlled experimental conditions, a diffused effect of low pH was detected 384 385 in A. colbecki and S. neumayeri, but not in O. validus, which did not display a clear response to OA as far as the three investigated variables: GSI, GS and anomaly ratio in gonad tissue. Although we 386 387 measured and weighed the specimens at the beginning and at the end of the experiment, we did 388 not perform any statistical analyses on body size and weight, as these species are known to grow 389 very slowly, even less than 1 gram (S. neumayeri: Brey et al. 1995; Suckling et al. 2015) or 2 mm (A. 390 colbecki: Heilmayer et al. 2003) per year, so that growth rate effects could not be detected over 391 the short experimental timeframe.

392 4.1. Adamussium colbecki

393

While GS was affected by pH 7.6, nothing can be inferred about GSI. Indeed, a significant 394 difference between the two control tanks was detected for the GSI, probably due to a high inter-395 396 individual variability that reduced the statistical power of the test. All the male specimens at pH 7.6 were in the spawning stage, while the other treatments accounted mostly for late maturation. 397 398 Premature spawning in response to environmental stress has already been observed in clams and mussels naturally or experimentally exposed to pollutants (e.g. Smaoui-Damak et al. 2006; 399 Gonzalez-Fernandez et al. 2016). Premature spawning seems to be a way to get rid of the affected 400 401 gametes or to avoid gamete alteration in case of adverse environmental conditions (Gonzalez-402 Fernandez et al. 2016). Synchronous spawning is fundamental in broadcast invertebrates to 403 ensure successful fertilization (Lewis and Ford, 2012) and asynchrony between males and females

404 could seriously affect the population dynamic of *A. colbecki*, known to display an intermittent 405 recruitment (Berkman et al. 1991).

Hemocyte aggregation inside bivalve follicles is a regular physiological occurrence during recovery stage, when hemocytes perform relict gamete resorption (Dorange and Le Pennec, 1989), but are normally excluded in any other gametogenic stage (Donaghy et al. 2009). Anomalous hemocyte infiltration in follicles is regarded as an important biomarker of inflammatory response, occurring in affected tissues (Donaghy et al. 2009; Cuevas et al. 2015), in bivalves either naturally or experimentally exposed to endocrine disruptors (e.g. Ortiz-Zarragoitia and Cajaraville 2010; Cuevas et al. 2015; Gonzalez-Fernandez et al. 2016).

Mucus production can be associated to the stress of the experimental procedures (Heinonen et al. 2007), but also to an immune response to oxidative stress (Allam and Espinosa, 2016). Since *A. colbecki* shows signs of stress under environmental hypercapnia (Benedetti et al. 2016) and given the high cost of mucus production (Davies and Hawkins 1998), the low pH treated scallops with a higher mucus production may have suffered from this additional energy demand.

418 4.2. Sterechinus neumayeri

In this species, no effects on GSI and GS were detected, in agreement with Suckling et al. (2015), who did not detect differences even after 2 years of exposure to acidified seawater; the differences recorded among tanks do not follow any pattern, but seem to be related to a high inter-individual variability. The same variability can be detected in the anomaly ratio, as we found different types of anomalies in different percentages, even between tanks of the same pH.

The most diffuse anomaly was a desquamation of gonad tissue, that occurred intensively in 424 female specimens, both at pH 7.8 and 7.6. Concurrently, atretic oocytes and disruption of part of 425 the second cohort of oocytes were found (Fig. 6e and 6f) as well as follicle membrane fibrosis (Fig. 426 427 6f). Schäfer and Köhler (2009) reported the same anomaly in the gonads of female *Psammechinus miliaris* exposed to fenanthren. These anomalies, other than destroy part of the new oocytes, may 428 affect the energy budget of the sea urchin because of the effort required to repair and reconstruct 429 the ovary tissue. Alteration in gonad development can also affect the fitness of the offspring 430 431 (Khristoforova et al. 1984; den Besten et al. 1991), as seems to be confirmed by the findings of 432 Suckling et al. (2015). The authors reported alterations in S. neumayeri larvae obtained from the 433 fertilization of eggs and sperms released by adult individual maintained at low pH (7.7 and 7.5). 434 Following three peculiar anomalies are described, not because of their frequency of

435 occurrence (very low), but because they are rare in echinoids in general, and usually associated, in

436 literature, to endocrine disruptors. Three of our treated sea urchins had stage 2 neoplasia (Carballal et al. 2015): one female at pH 7.8, one male and the hermaphrodite (female follicle) 437 specimen at pH 7.6. Neoplasia is the proliferation of germinal cells that prevent the normal 438 gamete maturation (Barber, 1996; Carballal et al. 2015); so far, it has been described for only a 439 440 few bivalves (Carballal et al. 2015 and references therein) and one gastropod only (Gagnaire et al. 2009) in relation to endocrine disruptors (Tay et al. 2003; Carballal et al. 2015). Two sea urchin 441 442 specimens, one from pH 7.8 and one from pH 7.6 showed hermaphroditism, a trait never reported for S. neumayeri and very rare for echinoids in general (Boolootian and Moore, 1956; Bernard 443 444 1977 and references therein; Byrne 1990 and references therein). In bivalves, hermaphroditism can occur in case of environmental stress (Tay et al. 2003; Ortiz-Zarragoitia and Cajaraville, 2010 445 446 and references therein) and, in some cases, oocytes become relict and are substituted by sperm, because of less energy is required by male gametes (Wintermyer and Cooper, 2007; Ortiz-447 448 Zarragoitia and Cajaraville, 2010). Such process is detectable in the central follicle in Fig. 6g. One male at pH 7.6 displayed a massive presence of LLP inside follicles during the spawning stage (Fig. 449 450 6j). LLP normally occurs in the gonads during recovery stage only (Vaschenko et al. 2012; Delorme 451 and Sewell, 2016), but is known to occur otherwise in case of oxidative stress (Schäfer et al. 2011). 452 Finally, two male specimens (one from pH 7.8 and one from 7.6) displayed a massive presence of 453 detached flagella, either inside the acinus lumen or between acini. We did not find any literature 454 reference for this anomaly, but we can hypothesize that the amount of healthy sperm, able to 455 swim and reach the oocytes, is unavoidably reduced.

456

Odontaster validus

457 The sea star O. validus did not display any sufferance from low pH exposure. In fact, no 458 effects were detected for any of the investigated variables, although tank 7.6a was slightly different from the other tanks with regard to GSI (Fig. 7). Similar to S. neumayeri, the variability of 459 460 the response seems to be due to an overall high inter-individual variability, since no pattern can be detected, either related to tanks or pH. As far as anomaly ratio, all the specimen gonads were in 461 good conditions with only one specimen at pH 7.8 showing altered gonad tissue. Opposite to the 462 adult resistance to OA reported in the present study, Gonzalez-Bernat et al. (2013) found reduced 463 464 survival and overall morphological alterations in larvae exposed to pH 7.6, corresponding to 1129.6 μ atm (similar to 1085.4 μ atm, which is the pCO₂ at pH 7.6 in our experiment). 465

466 4.3. Comparison of species

467 The response of the three species was very variable, both in terms of effect size and 468 typology. Indeed, while A. colbecki clearly displayed an advanced GS at the lowest pH, the GSI and gonad tissue did not seem to be significantly affected. Conversely S. neumayeri only displayed 469 diffuse anomalies in the gonad tissue, but no altered development. A particular case is O. validus, 470 which did not display any response to OA. While the experimental system was identical for the 471 three species, the only variables that can be considered affecting the response comparison are the 472 473 feeding condition and the gametogenesis stage. Food quality and availability can improve 474 invertebrate resistance to OA effects (Asnaghi et al. 2013), providing more energy to maintain and 475 repair tissues. Since phytoplankton was available in the aquarium water during the algal bloom, coincident with the experimental period, this could have been a source of resistance for the filter-476 477 feeding A. colbecki, which was in the gonad maturation stage. Probably thanks to the energy gained by feeding on the phytoplankton, the scallop was able to maintain the gonad tissue in good 478 479 condition. But the perception of an unfavorable surrounding environment may have been the trigger for premature release of gametes. S. neumayeri, in spawning stage, was the only species 480 that displayed a diffuse alteration of the gonad tissue. As mainly a deposit feeder, the sea urchin, 481 482 although not experiencing a real starvation, was probably not able to feed as usually during the 483 summer season. As well as other sea urchins, S. neumayeri can reallocate energy from gonads to 484 other tissues in case of starvation (Russel 1995; Brockington and Clarke 2001); the specimens treated at pH 7.8 and 7.6 may have reallocated their energy, favoring, most probably, the hard 485 486 tissue (Haag et al. 2016). Because the animals were in the gamete release stage, this energy 487 reallocation would not have affected the current reproduction of the sea urchin, but, rather, the second cohort of oocytes, which require energy for the germinal cell repair. A longer-term 488 489 experiment would be necessary to understand the gonad resilience of the sea urchin, following a 490 prolonged low pH exposure. O. validus, being preferably a scavenging/predator, is the only species 491 that could have been considered under starvation, yet, it is the only species that did not display any response to the treatments. The resistance of *O. validus* to the unfavorable conditions may be 492 explained by the very low somatic maintenance costs that allow the sea star to withstand even 493 494 long periods of starvation (Agüera et al. 2015). Moreover, Peck et al. (2008) showed a high 495 resistance of O. validus to rising temperature; together with our results, the sea star seems to 496 display a peculiar ability to cope with climate change features.

497 **5.** Conclusions

498 The response of the three Antarctic species here investigated was very variable, both in terms of effect size and typology. Histopathology with regards to OA, so far largely 499 underestimated, needs to be further investigated in the light of its endocrine alteration effects, 500 together with investigation on the energetic cost of their resistance/resilience to altered seawater 501 conditions. This may be of particular interest for calcifying species, since hard tissues can demand 502 up to 75% of the energy budget. The energy requirement is up to four times higher than the 503 amount required for reproduction (Sokolova et al. 2012). Studies performed during the winter 504 505 season are mostly required, when food scarcity potentially reduces gonad resistance and the 506 coastal seawater is expected to act as a source of CO₂ (Schram et al. 2015). Finally, the different response of the three species to OA needs to be evaluated in the framework of species interaction 507 and food web structure. In fact, if the predator O. validus will successfully cope with OA, opposite 508 to its prey A. colbecki, the potential cascading effects may be larger than the expected on basis of 509 510 a single species response.

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512 References

513 Agüera, A., Collard, M., Jossart, Q., Danis, B., 2016. Parameter Estimations of Dynamic Energy

514 Budget (DEB) Model over the Life History of a Key Antarctic Species: The Antarctic Sea Star

515 Odontaster validus Koehler, 1906. PLoS ONE 10 (10): e0140078.
516 https://doi.org/10.1371/journal.pone.0140078

Allam, B., Espinosa, E.P., 2016. Bivalve immunity and response to infections: Are we looking at the
right place? Fish & Shellfish Immunology 53, 4-12. http://dx.doi.org/10.1016/j.fsi.2016.03.037

519 Asnaghi, V., Chiantore, M., Mangialajo, L., Gazeau, F., Francour, P., Alliouane, S., Gattuso, J.P.,

520 2013. Cascading Effects of Ocean Acidification in a Rocky Subtidal Community. PLoS ONE 8(4), 521 e61978. https://doi.org/10.1371/journal.pone.0061978.t004

522 Baeta, M., Galimany, E., Ramón, M., 2016. Growth and reproductive biology of the sea star 523 Astropecten aranciacus (Echinodermata, Asteroidea) on the continental shelf of the Catalan Sea 524 (northwestern Mediterranean). Helgoland Marine Research 70: 1. 525 https://doi.org/10.1186/s10152-016-0453-z

526 Barber, B.J., 1996. Effects of gonadal neoplasms on oogenesis in softshell clams, *Mya arenaria*.

527 Journal of Invertebrates Pathology 67(2), 161-168. https://doi.org/10.1006/jipa.1996.0024

- 528 Barros, P., Sobrala, P., Range, P., Chícharo, L., Matias, D., 2013. Effects of sea-water acidification 529 on fertilization and larval development of the oyster *Crassostrea gigas*. Journal of Experimental 530 Marine Biology and Ecology 440, 200-206. https://doi.org/10.1016/j.jembe.2012.12.014
- 531 Belkhedim, L., Dermeche, S., Chahrour, F., Boutiba, Z., 2014. Physiological indices and 532 reproduction in the sea urchin *Paracentrotus lividus* (Lamarck, 1816) echinodermata echinoïdae in 533 the west coast of Algeria. IJRRAS 18(2), 173-181
- Benedetti, M., Lanzoni, I., Nardi, A., d'Errico, G., Di Carlo, M., Fattorini, D., Nigro, M., Regoli, F., 2016. Oxidative responsiveness to multiple stressors in the key Antarctic species, *Adamussium colbecki*: Interactions between temperature, acidification and cadmium exposure. Marine environmental research 121, 20-30. https://doi.org/10.1016/j.marenvres.2016.03.011
- 538 Berkman, P.A., Waller, T.R., Alexander, S.P., 1991. Unprotected larval development in the Antarctic
- scallop Adarnussium colbecki (Mollusca: Bivalvia: Pectinidae). Antarctic Science 3(2), 151-157.
- 540 https://doi.org/10.1017/S0954102091000184
- 541 Bernard, F.R., 1977. Fishery and Reproductive Cycle of the Red Sea Urchin, Strongylocentrotus
- franciscanus, in British Columbia. Journal of the Fisheries Research Board of Canada, 34(5), 604610, https://doi.org/10.1139/f77-095
- den Besten, P.J., van Donselaar, E.G., Herwig, H.J., Zandee, D.I., Voogt, P.A., 1991. Effects of
 cadmium on gametogenesis in the sea star *Asterias rubens* L. Aquatic Toxicology 20(1–2), 83-94.
 https://doi.org/10.1016/0166-445X(91)90043-9
- Boolootian, R.A., Moore, A.R., 1956. Hermaphroditism in echinoids. The Biological Bulletin 111 (3),
 328-335. https://doi.org/10.2307/1539139
- 549 Brey, T., Pearse, J., Basch, L., McClintock, J., Slattery, M., 1995. Growth and production of 550 *Sterechinus neumayeri* (Echinoidea: Echinodermata) in McMurdo Sound, Antarctica. Marine 551 Biology 124, 279-292. https://doi.org/10.1007/BF00347132
- 552 Brockington, S., Clarke, A., 2001. The relative influence of temperature and food on the 553 metabolism of a marine invertebrate. Journal of Experimental Marine Biology and Ecology, 258, 554 87–99. https://doi.org/10.1016/S0022-0981(00)00347-6
- 555 Brockington, S., Clarke, A., Chapman, A.L.G., 2001. Seasonality of feeding and nutritional status
- 556 during the austral winter in the Antarctic sea urchin Sterechinus neumayeri. Marine Biology,
- 557 139(1), 127–138. https://doi.org/10.1007/s002270100561

- 558 Byrne, M., 1990. Annual reproductive cycles of the commercial sea urchinParacentrotus lividus
- 559 from an exposed intertidal and a sheltered subtidal habitat on the west coast of Ireland. Mar. Biol.
- 560 104, 275-289. https://doi.org/10.1007/BF01313269
- 561 Byrne, M., 2012. Global change ecotoxicology: Identification of early life history bottlenecks in 562 marine invertebrates, variable species responses and variable experimental approaches. Marine 563 Environmental Research, 76, 3–15. https://doi.org/10.1016/j.marenvres.2011.10.004
- 564 Byrne, M., Przeslawsk, R., 2013. Multistressor Impacts of Warming and Acidification of the Ocean 565 on Marine Invertebrates' Life Histories. Integrative and Comparative Biology 53(4), 582–596. 566 https://doi.org/10.1093/icb/ict049
- 567 Caldeira, K., Wickett, M.E., 2003. Anthropogenic carbon and ocean pH. Nature 425, 365.
 568 https://doi.org/10.1038/425365a
- Cao, L., Caldeira, K., Jain, A.K., 2007. Effects of carbon dioxide and climate change on ocean
 acidification and carbonate mineral saturation. Geophysical Research Letters 34, L05607.
 https://doi.org/10.1029/2006GL028605
- Carballal, M., Barber, B.J., Iglesias, D., Villalba, A., 2015. Neoplastic diseases of marine bivalves.
 Journal of Invertebrate Pathology 131, 83–106. https://doi.org/10.1016/j.jip.2015.06.004
- Chiantore, M., Cattaneo-Vietti, R., Albertelli, G., Misic, C., Fabiano, M., 1998. Role of filtering and
 biodeposition by *Adamussium colbecki* in circulation of organic matter in Terra Nova Bay (Ross
 Sea, Antarctica). Journal of Marine Systems 17, 411-424. https://doi.org/10.1016/S09247963(98)00052-9
- 578 Chiantore, M., Cattaneo-Vietti, R., Berkman, P.A., Nigro, M., Vacchi, M., Schiaparelli, S., Albertelli,
 579 G., 2001. Antarctic scallop (*Adamussium colbecki*) spatial population variability along the Victoria
- 580 Land Coast, Antarctica. Polar Biology 24, 139-143. https://doi.org/10.1007/s00300000191
- 581 Chiantore, M., Cattaneo-Vietti, R., Elia, L., Guidetti, M., Antonini, M., 2002. Reproduction and 582 condition of the scallop *Adamussium colbecki* (Smith 1902), the sea-urchin *Sterechinus neumayeri* 583 (Meissner 1900) and the sea-star *Odontaster validus* (Koehler 1911) at Terra Nova Bay (Ross Sea): 584 different strategies related to inter-annual variations in food availability. Polar Biology 25, 251–
- 585 255. https://doi.org/10.1007/s00300-001-0331-1
- Collard, M., De Ridder, C., David, B., Dehairs, F., Dubois, P., 2013. Could the acidbase status of
 Antarctic sea urchins indicate a better-than expected resilience to near-future ocean acidification?
 Global change biology 21(2), 605-617. https://doi.org/10.1111/gcb.12735

- 589 Cornwall, C.E., Hurd., C.L., 2016. Experimental design in ocean acidification research: problems 590 and solutions. ICES Journal of Marine Science 73(3), 572–581. 591 https://doi.org/10.1093/icesjms/fsv118
- 592 Cuevas, N., Zorita, I., Costa, P.M., Franco, J., Larreta, J., 2015. Development of histopathological 593 indices in the digestive gland and gonad of mussels: Integration with contamination levels and 594 effects of confounding factors. Aquatic Toxicology 162, 152–164. 595 https://doi.org/10.1016/j.aquatox.2015.03.011
- 596 Davies, M.S., Hawkins, S.J., 1998. Mucus from Marine Molluscs. Advances in Marine Biology 34, 1597 71. https://doi.org/10.1016/S0065-2881(08)60210-2
- 598 Dell'Acqua, O., Brey, T., Vacchi, M., Chiantore, M., 2017. Predation impact of the notothenioid fish 599 *Trematomus bernacchii* on the size structure of the scallop *Adamussium colbecki* in Terra Nova Bay
- 600 (Ross Sea, Antarctica). Polar Biology 40, 1557-1568. https://doi.org/10.1007/s00300-017-2077-4
- Delorme, N.J., Sewell, M.A., 2016. Effects ofwarmacclimation on physiology and gonad
 development in the sea urchin *Evechinus chloroticus*. Comparative Biochemistry and Physiology,
 Part A 198, 33–40. https://doi.org/10.1016/j.cbpa.2016.03.020
- Dickson, A.G., Sabine, C.L., Christian, J.R., 2007. Guide to best practice for ocean CO₂ measurments, PICES Special Publication 3, Sidney, British Columbia, North Pacific Marine Science Organization
- Donaghy, L., Lambert, C., Choi, K.S., Soudant, P., 2009. Hemocytes of the carpet shell clam (*Ruditapes decussatus*) and the Manila clam (*Ruditapes philippinarum*): Current knowledge and
- 609 future prospects. Aquaculture 297, 10-24. https://doi.org/10.1016/j.aquaculture.2009.09.003
- Dorange, G., Le Pennec, M., 1989. Ultrastructural study of oogenesis and oocytic degeneration
- in Pecten maximus from the Bay of St. Brieuc. Marine Biology 103, 339-348.https://doi.org/10.1007/BF00397268
- Ericson, J.A., Ho, M.A., Miskelly, A., King, C.K., Virtue, P., Tilbrook, B., Byrne, M., 2012. Combined
 effects of two ocean change stressors, warming and acidification, on fertilization and early
 development of the Antarctic echinoid *Sterechinus neumayeri*. Polar Biology 35, 1027-1034.
 https://doi.org/10.1007/s00300-011-1150-7
- Fabry, V.J., Seibel, B.A., Feely, R.A., Orr, J.C., 2008. Impacts of ocean acidification on marine fauna
 and ecosystem processes. ICES Journal of Marine Science 65, 414–432.
 https://doi.org/10.1093/icesjms/fsn048

- Fabry, V.J., McClintock, J.B., Mathis, J.T., Grebmeier, J.M., 2009. Ocean Acidification at HighLatitudes: The Bellwether. Oceanography 22(4), 160-171.
 https://doi.org/10.5670/oceanog.2009.105
- 623 Gagnaire, B., Gagné, F., André, C., Blaise, C., Abbaci, K., Budzinski, H., Dévier, M.H., Garric, J., 2009.
- 624 Development of biomarkers of stress related to endocrine disruption in gastropods: Alkali-labile
- phosphates, protein-bound lipids and vitellogenin-like proteins. Aquatic Toxicology 92(3), 155-167.
- 626 https://doi.org/10.1016/j.aquatox.2009.01.012
- 627 Gillies, C.L., Stark, J.S., Johnstone, G.J., Smith, S.D.A., 2012. Carbon flow and trophic structure of an
- Antarctic coastal benthic community as determined by δ^{13} C and δ^{15} N. Estuarine, Coastal and Shelf Science 97, 44e57. dx.doi.org/10.1016/j.ecss.2011.11.003
- Gonzalez-Bernat, M.J., Lamare, M., Baker, M., 2013. Effects of reduced seawater pH on
 fertilisation, embryogenesis and larval development in the Antarctic seastar *Odontaster validus*.
 Polar biology 36, 235-247. https://doi.org/10.1007/s00300-012-1255-7
- González-Fernández, G, Albentosa, M., Campillo, J.A., Viñas, L., Franco, A., Bellas, J., 2016. Effect of
 mussel reproductive status on biomarker responses to PAHs: Implications for large-scale
 monitoring programs. Aquatic Toxicology 177, 380–394.
 https://doi.org/10.1016/j.aquatox.2016.06.012
- Haag, N., Russell, M.P., Hernandez, J.C., 2016. Effects of spine damage and microhabitat on 637 638 resource allocation of the purple sea urchin Strongylocentrotus purpuratus (Stimpson 1857). 639 Journal of Experimental Marine Biology and Ecology 482, 106-117. https://doi.org/10.1016/j.jembe.2016.05.005 640
- Heilmayer, O., Brey, T., Chiantore, M., Cattaneo-Vietti, R., Arntz, W.E., 2003. Age and productivity
 of the Antarctic scallop, *Adamussium colbecki*, in Terra Nova Bay (Ross Sea, Antarctica). Journal of
 Experimental Marine Biology and Ecology 288, 239–256. https://doi.org/10.1016/S00220981(03)00020-0
- Heinonen, K.B., Ward, J.E., Holohan, B.A., 2007. Production of transparent exopolymer particles
 (TEP) by benthic suspension feeders in coastal systems. Journal of Experimental Marine Biology
 and Ecology 341, 184–195. https://doi.org/10.1016/j.jembe.2006.09.019
- Hendriks, I.E., Duarte, C.M., Álvarez, M., 2010. Vulnerability of marine biodiversity to ocean
 acidification: A meta-analysis. Estuarine, Coastal and Shelf Science 86, 157–164.
 https://doi.org/10.1016/j.ecss.2009.11.022

- Hofmann, G.E., Todgham, A.E., 2010. Living in the Now: Physiological Mechanisms to Tolerate a
 Rapidly Changing Environment. Annual Review of Physiology 72, 127–45.
 https://doi.org/10.1146/annurev-physiol-021909-135900
- 654 Khristoforova, N.K., Gnezdilova, S.M., Vlasova, G.A., 1984. Effect of cadmium on gametogenesis
- and offspring of the sea urchin *Strongylocentrotus intermedius*. Marine Ecology Progress Series 17,
- 656 9-14. https://doi.org/10.3354/meps017009
- Kroeker, K.J., Kordas, R.L., Crim, R.N., Singh, G.G., 2010. Ecology Letters 13, 1419–1434.
 https://doi.org/10.1111/j.1461-0248.2010.01518.x
- Kurihara , H., Asai, T., Kato, S., Ishimatsu, A., 2008. Effects of elevated pCO₂ on early development
 in the mussel *Mytilus galloprovincialis*. Aquatic Biology 4(3), 225-233.
 https://doi.org/10.3354/ab00109
- Kurihara, H., Yin, R., Nishihara, G.N., Soyano, K., Ishimatsu A., 2013. Effect of ocean acidification on
 growth, gonad development and physiology of the sea urchin *Hemicentrotus pulcherrimus*.
- 664 Aquatic Biology 18(3), 281-292. https://doi.org/10.3354/ab00510
- Lehmann, D.W., Levine, J.F., Law, J.M., 2007. Polychlorinated Biphenyl Exposure Causes Gonadal
 Atrophy and Oxidative Stress in *Corbicula fluminea* Clams. Toxicologic Pathology 35, 356–365.
 https://doi.org/10.1080/01926230701230288
- Lewis, C., Ford, A.T., 2012. Infertility in male aquatic invertebrates: A review. Aquatic Toxicology
 120–121, 79–89. https://doi.org/10.1016/j.aquatox.2012.05.002
- Martinez, A.A., Suárez, P., Ruiz, Y., Vidal, A., San Juan, F., 2014. *In Vitro* Toxicity and
 Histopathological Effects Induced in the Mantle Tissue of Males of *Mytilus galloprovincialis* During
- 672 Short-Term Exposure to a Tar Mixture, in: Sauvé G. (Eds.), Molluscan Shellfish Safety. Springer,
- 673 Dordrecht, pp 229-238
- McClintock, J.B., Pearse, J.S., Bosch, I., 1988. Population structure and energetics of the common
 antarctic sea star Odontaster validus. Marine Biology 99, 235–246.
 https://doi.org/10.1007/BF00391986
- McClintock, J.B., 1994. Trophic biology of antarctic shallow-water echinoderms. Marine Ecology
 Progress Series 11, 191-202. https://doi.org/10.3354/meps111191
- McNeil, B.I., Matear, R.J., 2008. Southern Ocean acidification: A tipping point at 450-ppm
 atmospheric CO₂. PNAS 105 (48): 18860-18864. https://doi.org/10.1073/pnas.0806318105

Meredith, M.P., King, J.C., 2005. Rapid climate change in the ocean west of the Antarctic Peninsula
during the second half of the 20th century. Geophysical Research Letters 32, L19604,
https://doi.org/10.1029/2005GL024042

684 Michaelidis et al 2005. Effects of long-term moderate hypercapnia on acid–base balance and 685 growth rate in marine mussels *Mytilus galloprovincialis*. Marine Ecology progress Series 293, 109–

- 686 118. https://doi.org/10.3354/meps293109
- Millero, F.J., Graham, T.B., Huang, F., Bustos-Serrano, H., Pierrot, D., 2006. Dissociation constants
 of carbonic acid in seawater as a function of salinity and temperature. Marine Chemistry 100, 80 –
 94. https://doi.org/10.1016/j.marchem.2005.12.001
- 690 Morley, S.A., Suckling, C., Clark, M.S., Cross, E.L., Peck, L.S., 2016. Long-term effects of altered pH
- and temperature on the feeding energetics of the Antarctic sea urchin, *Sterechinus neumayeri*.
- 692 Biodiversity 17(1-2), 34-45. http://dx.doi.org/10.1080/14888386.2016.1174956
- Mos, B., Byrne, M., Dworjanyn, S.A., 2016. Biogenic acidification reduces sea urchin gonad growth
 and increases susceptibility of aquaculture to ocean acidification. Marine Environmental Research
- 695 113, 39-48. https://doi.org/10.1016/j.marenvres.2015.11.001
- 696 Ortiz-Zarragoitia, M., Cajaraville, M.P., 2010. Intersex and oocyte atresia in a mussel population
- 697 from the Biosphere's Reserve of Urdaibai (Bay of Biscay). Ecotoxicology and Environmental Safety
- 698 73(5), 693-701. https://doi.org/10.1016/j.ecoenv.2010.04.002
- Pearse, J.S., 1965. Reproductive periodicities in several contrasting populations of *Odontaster validus* Koehler, a common Antarctic asteroid. Antarctic Research Series 5, 39-85.
 https://doi.org/10.1029/AR005p0039
- Pearse, J.S., Giese, A.C., 1966. Food, reproduction and organic constitution of the common
 Antarctic echinoid *Sterechinus neumayeri* (Meissner). Biological Bulleetin 130(3), 387-401.
 https://doi.org/10.2307/1539745
- Peck, L.S., 2002. Ecophysiology of Antarctic merine ectotherms: limits to life. Polar Biology 25, 3140. https://doi.org/10.1007/s003000100308
- Peck, L.S., Webb, K.E., Miller, A., Clark, M.S., Hill, T., 2008. Temperature limits to activity, feeding
- and metabolism in the Antarctic starfish *Odontaster validus*. Marine Ecology Progress Series, 358,
- 709 181–189. https://doi.org/10.3354/meps07336
- Pohlert, T., 2014. The Pairwise Multiple Comparison of Mean Ranks Package (PMCMR). R package.
- 711 http://CRAN.R-project.org/package=PMCMR.

- R Core Team (2013). R: A language and environment for statistical computing. R Foundation for
 Statistical Computing, Vienna, Austria. URL http://www.R-project.org/
- Ries, J.B., Cohen, A.L., McCorkle, D.C., 2009. Marine calcifi ers exhibit mixed responses to CO₂induced ocean acidification. Geology 37(12), 1131–1134. https://doi.org/10.1130/G30210A.1
- Russell, M.P., 1998. Resource allocation plasticity in sea urchins: rapid, diet induced, phenotypic
 changes in the green sea urchin, *Strongylocentrotus droebachiensis* (Müller). Journal of
 Experimental Marine Biology and Ecology 220, 1–14. https://doi.org/10.1016/S00220981(97)00079-8
- Schäfer, S., Köhler, A., 2009. Gonadal lesions of female sea urchin (*Psammechinus miliaris*) after
 exposure to the polycyclic aromatic hydrocarbon phenanthrene. Marine Environmental Research
 68, 128–136. https://doi.org/10.1016/j.marenvres.2009.05.001
- Schäfer, S., Abele, D., Weihe, E., Köhler, A., 2011. Sex-specific biochemical and histological
 differences in gonads of sea urchins (*Psammechinus miliaris*) and their response to phenanthrene
 exposure. Marine Environmental Research 71(1), 70-78.
- 726 https://doi.org/10.1016/j.marenvres.2010.10.004
- Schiaparelli, S., Linse, K., 2006. A reassessment of the distribution of the common Antarctic scallop *Adamussium colbecki* (Smith, 1902). Deep-Sea Research, Part II 53, 912-920.
 https://doi.org/10.1016/j.dsr2.2006.02.004
- Schram, J.B., Schoenrock, K.M., McClintock, J.B., Amsler, C.D., Angus, R.A., 2015. Multi-frequency
 observations of seawater carbonate chemistry on the central coast of the western Antarctic
 Peninsula. Polar Research 34, 25582. http://dx.doi.org/10.3402/polar.v34.2558
- Smaoui-Damak, W., Rebai, T., Berthet, B., Hamza-Chaffai A., 2006. Does cadmium pollution affect 733 734 reproduction in the clam Ruditapes decussatus? A one-year case study. Comparative Biochemistry and Physiology Part C: Toxicology & Pharmacology 735 143(2), 252-261. 736 https://doi.org/10.1016/j.cbpc.2006.02.009
- Sokolova, I.M., Frederich, M., Bagwe, R., Lannig, G., Sukhotin, A.A., 2012. Energy homeostasis as
 an integrative tool for assessing limits of environmental stress tolerance in aquatic invertebrates.
- 739 Marine Environmental Research 79, 1-15. https://doi.org/10.1016/j.marenvres.2012.04.003
- Suckling, C.C., Clark, M.S., Richard, J., Morley, S.A., Thorne, M.A., Harper, E.M., Peck, L.S., 2015.
 Adult acclimation to combined temperature and pH stressors significantly enhances reproductive
 outcomes compared to short-term exposures. Journal of Animal Ecology 84(3), 773-84.
 https://doi.org/10.1111/1365-2656.12316

Taboada, S., Núñez-Pons, L., Avila, C., 2013. Feeding repellence of Antarctic and sub-Antarctic
benthic invertebrates against the omnivorous sea star *Odontaster validus*. Polar Biology, 36, 13-25.
https://doi.org/10.1007/s00300-012-1234-z

Tay, K.L., Teh, S.J., Doe, K., Lee, K., Jackman, P., 2003. Histopathologic and histochemical
biomarker responses of Baltic clam, Macoma balthica, to contaminated Sydney Harbour sediment,
Nova Scotia, Canada. Environmental Health Perspectives, 111(3), 273–280.
https://doi.org/10.1289/ehp.5437

- Thiéry, A., De Jong, L., Issartel, J., Moreaux, X., Saez, G., Barthélémy, P., Bestel, I., Santaella, C., 751 Achouak, W., Auffan, M., Rose, J., Bottero, J-Y., 2012. Effects of metallic and metal oxide 752 nanoparticles in aquatic and terrestrial food chains. Biomarkers responses in invertebrates and 753 bacteria. International 754 Journal of Nanotechnology 9(3-7), 181-203. https://doi.org/10.1504/IJNT.2012.045326 755
- Uthicke, S., Liddy, M., Nguyen, H.D., Byrne, M., 2014. Interactive effects of near-future
 temperature increase and ocean acidification on physiology and gonad development in adult
 Pacific sea urchin, *Echinometra* sp. A. Coral Reef 33(3), 831–845. https://doi.org/10.1007/s00338014-1165-y
- Vacchi, M., Cattaneo-Vietti, R., Chiantore, M., Dalu, M., 2000. Predator-prey relationship between 760 761 the nototheniid fish Trematomus bernacchii and the Antarctic scallop Adamussium colbecki at 762 Terra Nova Bay (Ross Sea). Antarctic science 12(1), 64-68. https://doi.org/10.1017/S095410200000092 763
- Van Colen, C., Debusschere, E., Braeckman, U., Van Gansbeke, D., Vincx M., 2012. The Early Life
 History of the Clam *Macoma balthica* in a High CO₂ World. PLoS ONE 7(9): e44655.
 https://doi.org/10.1371/journal.pone.0044655
- Vaschenko, M.A., Syasina, I.G., Zhadan, P.M., Medvedeva, L.A., 1997. Reproductive function state
 of the scallop *Mizuhopecten yessoensis* Jay from polluted areas of Peter the Great Bay, Sea of
 Japan. Hydrobiologia 352(1–3), 231–240
- Vaschenko, M.A., Zhadan, P.M., Aminin, D.L., Almyashova, T.N., 2012. Lipofuscin-like pigment in
 gonads of Sea Urchin *Strongylocentrotus intermedius* as a potential biomarker of marine pollution:
- a field study. Archives of Environmental Contamination and Toxicology 62(4), 599–613.
 https://doi.org/10.1007/s00244-011-9733-4

Widdicombe, S., Spicer, J.I., 2008. Predicting the impact of ocean acidification on benthic
biodiversity: What can animal physiology tell us? Journal of Experimental Marine Biology and
Ecology, 366(1-2), 187-197. https://doi.org/10.1016/j.jembe.2008.07.024

Wintermyer, M.L., Cooper, K.R., 2007. The development of an aquatic bivalve model: Evaluating
the toxic effects on gametogenesis following 2,3,7,8-tetrachlorodibenzo-*p*-dioxin (2,3,7,8-TCDD)
exposure in the eastern oyster (*Crassostrea virginica*). Aquatic Toxicology 81, 10–26.
https://doi.org/10.1016/j.aquatox.2006.10.005

- Wittmann, A.C., Pörtner, H.O., 2013. Sensitivities of extant animal taxa to ocean acidification.
 Nature Climate Change 3, 995-1001. https://doi.org/10.1038/NCLIMATE1982
- 783 Zheng, B.H., An, L.H., Chang, H., Liu, Y., Jiang, Z.Q., 2014. Evidence for the presence of sex steroid
- 784 hormones in Zhikong scallop, *Chlamys farreri*. Journal of Steroid Biochemistry & Molecular Biology
- 785 143, 199–206. https://doi.org/10.1016/j.jsbmb.2014.03.002

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