

# Accepted Manuscript

Rhizosphere response to nickel in a facultative hyperaccumulator

Stefano Rosatto, Enrica Roccotiello, Simone Di Piazza, Grazia Cecchi, Giuseppe Greco, Mirca Zotti, Luigi Vezzulli, Mauro Mariotti



PII: S0045-6535(19)31108-7

DOI: <https://doi.org/10.1016/j.chemosphere.2019.05.193>

Reference: CHEM 23928

To appear in: *ECSN*

Received Date: 23 July 2018

Revised Date: 9 May 2019

Accepted Date: 22 May 2019

Please cite this article as: Rosatto, S., Roccotiello, E., Di Piazza, S., Cecchi, G., Greco, G., Zotti, M., Vezzulli, L., Mariotti, M., Rhizosphere response to nickel in a facultative hyperaccumulator, *Chemosphere* (2019), doi: <https://doi.org/10.1016/j.chemosphere.2019.05.193>.

This is a PDF file of an unedited manuscript that has been accepted for publication. As a service to our customers we are providing this early version of the manuscript. The manuscript will undergo copyediting, typesetting, and review of the resulting proof before it is published in its final form. Please note that during the production process errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain.

1 **Rhizosphere response to nickel in a facultative hyperaccumulator**

2

3 Stefano Rosatto<sup>1</sup>, Enrica Roccotiello<sup>1\*</sup>, Simone Di Piazza<sup>2</sup>, Grazia Cecchi<sup>2</sup>, Giuseppe Greco<sup>2</sup>, Mirca  
4 Zotti<sup>2</sup>, Luigi Vezzulli<sup>3</sup>, Mauro Mariotti<sup>1</sup>

5

6 Stefano Rosatto<sup>1</sup>, stefano.rosatto@gmail.com

7 Enrica Roccotiello<sup>1</sup>, enrica.roccotiello@unige.it

8 Simone Di Piazza<sup>2</sup>, simone.dipiazza@unige.it

9 Grazia Cecchi<sup>2</sup>, grazia.cecchi@edu.unige.it

10 Giuseppe Greco<sup>2</sup>, giuseppe.greco@edu.unige.it

11 Mirca Zotti<sup>2</sup>, mirca.zotti@unige.it

12 Luigi Vezzulli<sup>3</sup>, luigi.vezzulli@unige.it

13 Mauro Mariotti<sup>1</sup>, m.mariotti@unige.it

14

15 DISTAV-Department of Earth, Environment and Life Sciences, University of Genoa, Corso Europa  
16 26, 16132, Genova. <sup>1</sup>Laboratory of Plant Biology, <sup>2</sup>Laboratory of Mycology, <sup>3</sup>Laboratory of  
17 Microbiology

18

19 Corresponding author: \*Enrica Roccotiello, DISTAV-Department of Earth, Environment and Life  
20 Sciences, University of Genoa, Corso Europa 26, 16132, Genova. <sup>1</sup>Laboratory of Plant Biology  
21 e-mail: enrica.roccotiello@unige.it

22 phone +39 010 353 8047

23

24

25

26

## 27 Abstract

28 This study faces the characterization of the culturable microbiota of the facultative Ni-  
29 hyperaccumulator *Alyssoides utriculata* to obtain a collection of bacterial and fungal strains for  
30 potential applications in Ni phytoextraction.

31 Rhizosphere soil samples and adjacent bare soil associated with *A. utriculata* from serpentine and  
32 non-serpentine sites were collected together with plant roots and shoots. Rhizobacteria and fungi  
33 were isolated and characterized genotypically and phenotypically. Plants and soils were analyzed  
34 for total element concentration using Inductively Coupled Plasma Mass Spectrometry (ICP-MS).

35 Serpentine and non-serpentine sites differ in terms of elements concentration in soil, plant roots and  
36 shoots. Ni and Co are significantly higher on serpentine site, while Ca is more abundant in non-  
37 serpentine site.

38 Bacteria and fungi were significantly more abundant in rhizosphere than in bare soil and were  
39 dominated by genera *Arthrobacter*, *Bacillus* and *Streptomyces*, *Penicillium* and *Mucor*. The genus  
40 *Pseudomonas* was only found in rhizospheric serpentine soils (< 2 % of total serpentine isolates)  
41 and with *Streptomyces* sp. showed highest Ni-tolerance up to 15 mM. The same occurred for  
42 *Trichoderma* strain, belonging to the *harzianum* group (< 2 % of the total microfungal count) and  
43 *Penicillium ochrochloron* (< 10 % of the total microfungal count, tolerance up to Ni 20 mM).  
44 Among serpentine bacterial isolates, 8 strains belonging to 5 genera showed at least one PGPR  
45 activity (1-Aminocyclopropane-1-Carboxylic Acid (ACC) deaminase activity, production of indole-  
46 3-acetic acid (IAA), siderophores and phosphate solubilizing capacity), especially genera *Pantoea*,  
47 *Pseudomonas* and *Streptomyces*. Those microorganisms might thus be promising candidates for  
48 employment in bioaugmentation trials.

49

## 50 Highlights

- 51 • Only 10% of hyperaccumulators have their rhizosphere examined

- 52 • *A. utriculata* is a facultative Ni-hyperaccumulator that thrives in serpentine soils
- 53 • Rhizobiota of *A. utriculata* seems to be limited by serpentine soil conditions
- 54 • *Pseudomonas*, *Streptomyces*, *T. harzianum* group and *P. ochrochloron* hypertolerate Ni
- 55 • Rhizobiota with PGP traits and high Ni tolerance can improve plant phytoextraction

56

57 **Keywords:** microfungal strain, metal uptake, microbiota, PGPR, rhizobacteria, root area.

58

## 59 1. Introduction

60 The rhizosphere, i.e., the soil-root interface for the assimilation of soil elements and the uptake of  
61 metals, is characterized by feedback loops of interactions among root processes, physical and  
62 chemical soil characteristics, and microbial dynamics (Wenzel et al. 2004; Comerford, 2005; Alford  
63 et al., 2010; Kidd et al., 2017).

64 Among natural metalliferous soils, ultramafic serpentinitic soils usually have geochemical traits,  
65 which include an elevated concentration of magnesium (Mg) and iron (Fe), a low calcium (Ca):Mg  
66 ratio and elevated concentrations of trace elements such as nickel (Ni) (bioavailable Ni 7 to >100  
67 mg kg<sup>-1</sup>; total Ni 500-8000 mg kg<sup>-1</sup>), chromium (Cr), and cobalt (Co) (Reeves and Baker, 2000;  
68 Freitas et al., 2004; Ghaderian et al., 2007; Turgay et al., 2012; Van der Ent et al., 2013; Roccotiello  
69 et al., 2015a; Kidd et al., 2018). They are also known for their paucity of macronutrients such as  
70 nitrogen (N), potassium (K), or phosphorus (P) (Nkrumah et al., 2018).

71 Specialized plants known as hyperaccumulators are able to live and reproduce on those soils  
72 without experiencing toxicity symptoms (Rascio and Navari-Izzo, 2011). Those plants are also  
73 cultivated to accumulate more than 1% leaf metals from soils and to translocate them to the  
74 harvestable biomass (Chaney et al., 2007). In addition, *facultative hyperaccumulators able to live*  
75 *on both metalliferous and non-metalliferous sites are of a key importance because of their*  
76 *environmental plasticity respect to metal stress (Pollard et al., 2014) with consequent better*

77 *adaptation to metal-disturbed habitat*. In the rhizosphere of Ni-hyperaccumulators, the optimal  
78 plant growth, metal tolerance, and increased Ni uptake are strongly influenced by the native  
79 microbial community (Jing et al., 2007; Abouddrar et al., 2013; Rue et al., 2015). . Recently, studies  
80 on the role of microbiota in the rhizosphere of metalliferous plants have encouraged the  
81 development of remediation technologies employing native soil microorganisms for metal  
82 phytoextraction (Zhuang et al., 2007; Lebeau et al., 2008; Ma et al., 2009a; Glick, 2010; Sessitsch  
83 et al., 2013; Cabello-Conejo et al., 2014). Besides, soil metal contamination exerts selective  
84 pressure and changes on microbial communities and functional diversity (Kelly and Tate, 1998;  
85 Liao and Xie, 2007). Soil microorganisms (bacteria and fungi) can be metal-tolerant (Mengoni et  
86 al., 2001; Abou-Shanab et al., 2007; Iram et al., 2012; Turgay et al., 2012; Álvarez-López et al.,  
87 2016; Kidd et al., 2017; Thijs et al., 2017;) i.e., able to thrive in metal-rich soil (Gadd, 1992; Lima  
88 de Silva et al., 2012).

89 *Serpentine bacteria were studied for the ability to mobilize metals and promote plant growth as in*  
90 *the case of Microbacterium, Arthrobacter, Agreia, Bacillus, Micrococcus, Stenotrophomonas,*  
91 *Kocuria, and Variovorax interacting with the obligate Ni-hyperaccumulating species Noccaea*  
92 *caerulescens (J.Presl & C.Presl) F.K.Mey. (=Thlaspi c. J.Presl & C.Presl), Pseudomonas acting in*  
93 *synergy with roots of Odontarrhena bertolonii (Desv.) L. Cecchi & Selvi (=Alyssum b. Desv.),*  
94 *Microbacterium and Sphingomonas belonging to the rhizosphere of Odontarrhena muralis complex*  
95 *(=Alyssum m.) and members of the genus Burkholderia collected in the rhizosphere of Pycnandra*  
96 *acuminata (Pierre ex Baill.) Swenson & Munzinger (=Sebertia a. Pierre ex Baill.) and Psychotria*  
97 *douarrei (Beauvis.) Däniker, together with other nickel-resistant strains, like Hafnia alvei,*  
98 *Pseudomonas mendocina, Acinetobacter, Comamonas acidovorans, and Agrobacterium*  
99 *tumefaciens (Schlegel et al., 1991; Stoppel and Schlegel, 1995, Mengoni et al., 2001; Idris et al.,*  
100 *2004; Barzanti et al., 2007; Abouddrar et al., 2013). In addition, Plant Growth Promoting*  
101 *Rhizobacteria (PGPR) can be found in association with the root system of hyperaccumulators*

102 increasing plant growth, biomass development, and protecting plant from stresses (Benizri et al.,  
103 2001; Abouddrar et al., 2013; de Souza et al., 2015).

104 Previous studies isolated fungi like *Aspergillus*, *Botrytis*, *Clonostachys*, *Eurotium*, *Penicillium*,  
105 *Rhodotorula*, and *Trichoderma* from the rhizosphere of the facultative Ni-hyperaccumulator  
106 *Alyssoides utriculata* (Roccotiello et al. 2010; 2015a; 2016) growing on serpentine soils (Zotti et  
107 al., 2014; Roccotiello et al., 2015b; Cecchi et al., 2017b). Most of them were also metal-  
108 accumulator (Zotti et al., 2014; Roccotiello et al., 2015b; Cecchi et al., 2017b). Also, rhizosphere  
109 fungi have the potential to assist the growth of hyperaccumulator plants in metal-rich soil and to  
110 increase their metal uptake (Husna et al., 2017; Thijs et al., 2017). Many studies were carried out  
111 on rhizospheric mycorrhizal communities of serpentine Ni-hyperaccumulator plants (Amir et al.,  
112 2013; Husna et al., 2017), but few of them investigated the role of serpentine non-mycorrhizal fungi  
113 associated with plant rhizosphere (Pal et al., 2006; Urban et al., 2008).

114 For instance, a comparison among the serpentine rhizospheric bacterial biodiversity associated  
115 with some subspecies of the Ni-hyperaccumulator *A. serpyllifolium* Desf., the Ni-excluder *Dactylis*  
116 *glomerata* L., and the non-hyperaccumulator *Santolina semidentata* Hoffmanns. & Link, revealed  
117 that *A. serpyllifolium* subspecies hosted a greater density of bacteria than the non-  
118 hyperaccumulator species (Álvarez-López et al. 2016). This selective increase of Ni-tolerant  
119 bacteria in the rhizosphere was correlated with enhanced Ni bioavailability in soil (Becerra-Castro  
120 et al., 2009).

121 *Alyssoides utriculata* is an evergreen shrub, Ni-facultative hyperaccumulator, able to concentrate  
122 up to 2000 mg kg<sup>-1</sup> Ni in shoots when growing on serpentine sites (Roccotiello et al., 2010, 2015a;  
123 2015b; 2016). Despite the medium-high ability to concentrate Ni in shoots, this species is of a key  
124 importance because it is a native Mediterranean hyperaccumulator that can be exploited for  
125 improved phytoremediation purposes in this climate.

126 Nowadays, a deep comparison between serpentine and non-serpentine microbiota of facultative Ni-  
127 hyperaccumulators like *A. utriculata* is missing. The inoculation of some Ni-tolerant bacterial and  
128 fungal strains in the rhizosphere of hyperaccumulators, specifically facultative-hyperaccumulators,

129 can increase the efficiency of phytoextraction by promoting the development of root biomass (Ma et  
130 al., 2009a) and enhancing Ni accumulation in plant organs (Ma et al., 2011a).

131 This study is aimed at characterizing the culturable microbiota associated with the rhizosphere of  
132 the facultative Ni-hyperaccumulator *A. utriculata* from serpentine and non-serpentine sites and to  
133 obtain a collection of bacterial and fungal strains for application in Ni phytoextraction. Ni tolerance  
134 of rhizosphere bacteria and fungi and PGPR traits of isolated bacterial strains have been also  
135 evaluated.

136 The paper shows the first results of an interdisciplinary approach focused on the assessment of the  
137 whole rhizospheric components in a facultative hyperaccumulator.

138

## 139 **2. Materials and methods**

### 140 *2.1 Sampling sites, plant and soil sample collection*

141 The facultative Ni-hyperaccumulator *Alyssoides utriculata* and related soil were sampled from  
142 serpentine (S) and non-serpentine sites (NS). The S is located at Beigua Geopark (44°27'41.4"N  
143 8°40'03"E) in the eastern Ligurian Alps. The soils derived from ultramafic bedrocks like  
144 serpentines and eclogitic metagabbros (Capponi and Crispini, 2008; Marsili et al., 2009). The NS  
145 site was the locality of Glori in the NW of Liguria (43°57'19"N 7°50'08" E), geologically  
146 characterized by flysch and clay marl (Giammarino et al., 2010).

147 Five shoots of adult plants of *A. utriculata* from non-fruiting branches and roots replicates (1 plant  
148 = 1 replicate) were collected in July 2016 (fruiting stage) from the S and NS sites (n = 5 each site).

149 Ni-hyperaccumulation in plants was assessed with colorimetric dimethylglyoxime test (DMG 1% in  
150 ethanol 95%, Sigma-Aldrich) (Charlot, 1964), then plants were quantitatively evaluated with ICP-  
151 MS.

152

### 153 *2.2 Rhizospheric soil sampling*

154 Rhizosphere soil from S and NS site was collected from five different plants (n= 5 each site). Plants  
155 were carefully dug out with an intact root system and the soil tightly adhering to the roots was  
156 collected. The rhizosphere soil was obtained by agitating roots and sampling the soil still attach to  
157 the roots according to Khan et al. (2015). Each rhizospheric soil sample was placed into a plastic  
158 bag to avoid microbial mixing between the soils and transported into a refrigerated box. In  
159 laboratory, samples were processed as described in '*Isolation and identification of culturable*  
160 *bacteria*'.

161 No vegetation was observed in the bare soil from (Khan et al., 2015) S and NS site. Bulk soil  
162 samples were collected from five different points which were 20 m away from the vegetation to a  
163 depth of approximately 15 cm.

164 Bare and rhizospheric soil samples not immediately processed were stored at -20°C for about eight  
165 months.

166

### 167 2.3 Plant and soil sample analysis

168 In the laboratory, the plant samples were thoroughly rinsed first with tap water and then with  
169 deionized water to remove dust and soil particles. After oven-drying (60°C, 48 h), the leaves and  
170 roots were separately powdered using a ball mill (Retsch MM2000, Haan, Germany), before the  
171 chemical analyses.

172 Soil samples were air-dried, particle size distribution analysis was carried out by wet-sieving for the  
173 fraction >50 µm and, the composition of the fine fraction (<50 µm) was determined by pipette  
174 procedure after dispersion of the sample with sodium hexametaphosphate, (NaPO<sub>3</sub>)<sub>6</sub>.

175 The pH was measured with the potentiometric method in a 1:2.5 soil:water suspension. The cation  
176 exchange capacity (CEC) were determined with BaCl<sub>2</sub>-triethanolamine at pH 8.2.



177 The soil samples were oven-dried at 60°C for 48 h before being sieved through a 2.0 mm mesh. The  
178 soil and plant fractions were analyzed for Ni, Ca, Mg, Co and Cr concentration by means of a  
179 Inductively Coupled Plasma Mass Spectrometry (ICP-MS).

180 The accuracy of the results was checked processing BCR-100 'beech leaves' reference material  
181 (JRC-IRMM, 2004). Plant and soil metal concentrations were expressed on a dry weight basis  
182 (DW).

183

#### 184 *2.4 Isolation and identification of culturable bacteria*

185 Bacteria were extracted from 2.5 g of fresh soil by using 25 ml of sterile saline solution. Aliquots (1  
186 ml) were serially diluted with NaCl 0.9% w/v (Khan et al., 2015; Turgay et al., 2012) and spread on  
187 Tryptic Soy Agar (TSA, Sigma-Aldrich) (Idris et al., 2004; Vieira and Nahas, 2005; Aboudrar et al.,  
188 2013; Rue et al., 2015) and amended with 50 mg l<sup>-1</sup> of cycloheximide (Sigma-Aldrich) to inhibit the  
189 microfungus growth (Visioli et al., 2014).

190 Plates (n= 90, each site and soil type) were incubated in the dark at 27±1°C for 72 h (Barzanti et al.,  
191 2007; Sheng et al., 2008; Luo et al., 2011; Aboudrar et al., 2013). Bacterial colonies with distinct  
192 morphologies (color, shape, size, opacity, etc.) were selected from the plates and repeatedly re-  
193 streaked onto fresh agar medium prepared as previously described to obtain pure bacterial colonies  
194 (Turgay et al., 2012; Khan et al., 2015). Relative concentration of identified morphotypes (Hanirah  
195 et al., 2015) was also determined and expressed as Colony Forming Units per g of dry soil (CFU g<sup>-1</sup>)  
196 (Rue et al., 2015). The most representative morphotypes were selected (n= 30, three replicates  
197 each isolate) to perform the DNA sequencing.

198 To identify isolates, DNA extraction from pure bacterial culture was performed by the boiling  
199 method (99°C, 10 min). PCR amplification of a 409 bp region was then performed using the  
200 universal primers Com1 (5'-CAGCAGCCGCGGTAATAC-3') and Com2 (5'-  
201 CCGTCAATTCCTTTGAGTTT -3'), amplifying positions 519–926 of the *Escherichia coli*  
202 numbering of the 16S rRNA gene (Schwieger and Tebbe, 1998). Each PCR was performed in a

203 total volume of 15  $\mu$ l in micro-test tubes (Eppendorf s.r.l., Milan, Italy). Reaction mixtures  
204 contained 1 $\times$  PCR buffer with 1.5 mM of magnesium chloride ( $MgCl_2$ ), deoxynucleoside  
205 triphosphate solution (200 mM each dATP, dCTP, dGTP and dTTP), primers Com1 and Com2 (0.5  
206 mM each), and 2.5 U of DNA polymerase (FastStart High Fidelity enzyme blend, Sigma Aldrich  
207 srl, Milan, Italy). The temperature profile for the PCR was as follows: an initial step of 10 min at  
208 95°C, followed by denaturation for 1.30 min at 94°C, annealing for 40 s at 55°C and primer  
209 extension for 40s at 72°C. After the 35th cycle, the extension step was prolonged for 10 min to  
210 complete synthesis of all strands, and then the samples were kept at 4°C until analysis.

211 Amplified fragments from the PCR reaction were purified using the High Pure PCR product  
212 purification kit (Roche Diagnostics, Mannheim, Germany) and sequenced using the automated ABI  
213 Prism 3730 DNA sequencer (Applied Biosystems). 16S rRNA gene sequence similarity was  
214 determined using the BLAST function of the CLC Genomics workbench (version 9.5.1).

215 The sequences were submitted in the NCBI GenBank® database under the accession numbers from  
216 MG661811 to MG661840.

217 The isolated strains were cryoconserved at -80°C in 20% glycerol in Luria Bertani (LB, Sigma-  
218 Aldrich) broth in the Laboratory of Microbiology of DISTAV (University of Genoa, Italy).

219

## 220 *2.5 Screening for PGPR traits and Ni-tolerance in bacteria*

221 Serpentine Bacterial strains were screened for their ability to grow on 1-Aminocyclopropane-1-  
222 Carboxylic Acid (ACC) as the sole N source, to produce indole-3-acetic acid (IAA) and  
223 siderophores and solubilize phosphorous. For ACC deaminase activity, following an incubation of  
224 24 h in Tryptic Soy Broth (TSB, Sigma-Aldrich) at 28°C, bacterial suspension was harvested by  
225 centrifugation (4500 g x 10 min) then a 1-ml aliquot was transferred to 50 ml sterile Dworkin and  
226 Foster (DF) mineral medium (Dworkin and Foster, 1958) added with 300  $\mu$ l of ACC (Alfa Aesar)  
227 instead of  $(NH_4)_2SO_4$  as nitrogen source. A 0.5 M solution of ACC (labile in solution) was  
228 previously filter-sterilized through 0.2  $\mu$ m pore size membrane and added to DF medium. The salts

229 minimal medium without N was used as control. The solution was incubated at 28°C for at least 24  
230 h (Penrose and Glick, 2003; Luo et al., 2011). Turbidity indicates positive growth.

231 Auxin IAA production was estimated using a spectrophotometric method (Bric et al., 1991).

232 Isolates were grown in TSB amended with tryptophan (trp, 1 mg/ml broth) at 32±2°C for 4 days.

233 Surnanant was mixed with Salkowski Reagent (1 ml 0.5 M FeCl<sub>3</sub> in 50 ml 35% HClO<sub>4</sub>) in the ratio  
234 of 1:1. After 25-30 min at room temperature, the development of pink color highlights IAA

235 production. The optical density was measured using spectrophotometer (Jenway 6300  
236 spectrophotometer) at 530 nm of absorbance and auxin concentration was determined using

237 standard curve of IAA (Ma et al., 2009b; Goswami et al., 2015). The siderophores production was  
238 determined after 5 days of incubation at 30°C on Chrome Azurol Sulfonate (CAS) agar (Schwyn

239 and Neilands, 1987), through the development of red-orange halo around the colony (Durand et al.,  
240 2016).

241 Phosphate solubilization activity was assessed by the formation of a clear halo around the colony on  
242 Pikovskaya's agar medium (Khan et al., 2015).

243 Afterwards, each PGPR isolate was tested for metal tolerance on Tryptic Soy Agar (TSA, Sigma-  
244 Aldrich) spiked with NiSO<sub>4</sub>\*6H<sub>2</sub>O (Sigma-Aldrich), at the concentration of 1, 5, 10, 15, and 20 mM

245 (n= 15 each strain).

246

#### 247 *2.6 Isolation and identification of culturable fungi*

248 Fungi were counted and isolated through a dilution plate technique (Cecchi et al., 2017b; Greco et  
249 al., 2017; Zotti et al., 2014) by using two different culture media: Malt Extract Agar added with

250 Chloramphenicol (MEA+C) and Rose Bengal agar (RB) (Greco et al. 2017). The dilution was  
251 obtained by mixing 1 g of soil with 100 ml of sterile water. Each sample was plated in duplicate, for

252 each dilution (1:50.000 and 1:100.000). The plates were then incubated at 24±1°C, in the dark, for  
253 14 days and checked daily.

254 The colonies forming unit (CFU) were counted for each fungal strain grown in plates. Then, these  
255 strains were isolated and subcultured onto Malt Extract Agar (MEA). The pure cultures were  
256 maintained on MEA slants and kept at 4°C.

257 All the fungal isolates were initially identified by a polybasic approach on the base of their micro-  
258 macromorphological, physiological, and molecular characteristics. The morphological identification  
259 was confirmed by molecular analysis. PCR amplification of  $\beta$ -tubulin gene was performed using  
260 Bt2a and Bt2b primers (Glass and Donaldson, 1995) and ITS region amplification using universal  
261 primers ITS 1F (5'-CTTGGTCATTTAGAGGAAGTAA-3') and ITS 4 (5'-  
262 TCCTCCGCTTATTGATATGC-3') (White et al., 1990; Gardes and Bruns, 1993). The PCR  
263 cycling parameters were the following: one cycle of 5 min at 95 °C; thirty-five cycle of 40 s at 94  
264 °C, 45 s at 55 °C, 1 min at 72 °C; one cycle of 10 min at 72 °C;  $\infty$  at 4 °C. The taxonomic  
265 assignment of the sequenced samples was carried out by means of the BLASTN algorithm thus  
266 allowing us to compare the sequences obtained in our study with the ones available in the GenBank  
267 database. The sequences were submitted in the NCBI GenBank® database under the accession  
268 numbers from MG836709 to MG850983.

269 The isolated strains were conserved in the culture collection of Mycological Laboratory of DISTAV  
270 (University of Genoa, Italy). These cultures were maintained by agar slants with periodic transfers  
271 and then cryoconserved (-20°C).

272

### 273 *2.7 Screening for Ni-tolerance in fungi*

274 Each fungal strain isolated from S site was tested for metal tolerance on Malt Extract Agar (MEA,  
275 Sigma-Aldrich) spiked with NiSO<sub>4</sub>\*6H<sub>2</sub>O at the concentration of 1, 5, 10, 15, and 20 mM (n= 15  
276 each strain).

277

### 278 *2.8 Data analysis*

279 The Bioaccumulation Factor ( $BF = C_{\text{shoot}}/C_{\text{soil}}$ ) and the Translocation Factor ( $TF = C_{\text{shoot}}/C_{\text{root}}$ ) were  
280 calculated.

281 The statistical analyses were performed with Statistica 8.0 (Statsoft Inc.) software.

282 The averages were presented with their standard deviations (SD). The t-test was used to evaluate the  
283 differences between serpentine and non-serpentine samples. Significance was considered at the  $P <$   
284 0.05 level.

285 The differences in the number of bacterial and fungal colonies was evaluated on different sites (S  
286 and NS) and substrates (R and B) by means of factorial ANOVA. Data significance was tested with  
287 a post hoc Tukey's Honest Significant Difference test (HSD) at  $P < 0.05$ .

288 The Principal Components Analysis (PCA) was performed as a multivariate display method to  
289 visualize the data structure. Significance was considered at the  $P < 0.05$  level.

290

### 291 **3. Results**

#### 292 *3.1 Soil and plant sample analysis*

293 Total nickel concentration in soils (Table 1) is significantly high on S site respect to NS site ( $Ni_S >$   
294  $1000 \text{ mg kg}^{-1}$ ;  $Ni_{NS} \sim 20 \text{ mg kg}^{-1}$ ,  $P < 0.001$ ), both in R and in B soil. On S site *A. utriculata*  
295 hyperaccumulates Ni as expected ( $Ni_{\text{roots}} \sim 200 \text{ mg kg}^{-1}$   $Ni_{\text{shoots}} \sim 1200 \text{ mg kg}^{-1}$ ) (Table 2). T-test  
296 reveals clear differences between S and NS sites ( $P < 0.05$ , Table 1) as regards soil (Ni, Ca, Co),  
297 plant roots (Ni, Ca, Mg, Cr) and shoots (Ni, Mg, Co).

298

299

Site	Soil type	Element concentration (mg kg <sup>-1</sup> )					CEC meq/100g	pH	Fine Earth fraction %			Parent material
		Ca	Mg	Co	Cr	Ni			Sand	Silt	Clay	
S	R	3249.30±645.54 <sup>a</sup>	9032.93±4878.95	183.06±44.92 <sup>a</sup>	1693.60±308.58	1491.48±624.09 <sup>a</sup>	24.00	7.3	56.10	34.30	9.60	Serpentine
S	B	4267.50±922.94 <sup>a</sup>	11957.34±1876.38	49.74±19.63 <sup>a</sup>	1770.30±137.88	1046.65±404.77 <sup>a</sup>						
NS	R	9288.04±66.76 <sup>b</sup>	7050.32±42.44	13.13±1.62 <sup>b</sup>	27.08±1.63	19.52±0.94 <sup>b</sup>	19.30	8.3	38.4	43.8	17.8	Marl
NS	B	9336.06±54.08 <sup>b</sup>	7095.98±88.31	10.88±4.15 <sup>b</sup>	19.42±6.67	19.46±1.14 <sup>b</sup>						

300

301 **Table 1** Bulk element concentration (mg kg<sup>-1</sup>) and main chemical and physical features in soil samples. S: Serpentine site, NS: non-serpentine site, R: Rhizosphere soil, B: Bare  
 302 soil. Data are mean±SD. n= 5 each site, each soil type. Significant differences were marked with different superscript letters in the same column (P < 0.05).

303

Site	Plant sample	Element concentration (mg kg <sup>-1</sup> )				
		Ni	Ca	Mg	Co	Cr
S	Shoot	1241.37±245.67 <sup>a</sup>	10142.29±1942.62	16342.84±2502.94 <sup>a</sup>	29.97±11.75 <sup>a</sup>	9.70±6.39
S	Root	203.04±49.75 <sup>c</sup>	2886.33±372.81 <sup>a</sup>	2374.12±228.19 <sup>c</sup>	1.46±1.17	3.76±1.53 <sup>a</sup>
NS	Shoot	10.64±0.45 <sup>b</sup>	9220.26±1766.01	21258.69±135.74 <sup>b</sup>	6.64±1.38 <sup>b</sup>	3.90±1.90
NS	Root	5.95±0.25 <sup>d</sup>	1317.18±252.2 <sup>b</sup>	2024.64±12.93 <sup>d</sup>	0.59±0.28	1.61±0.07 <sup>b</sup>

304

305 **Table 2** Bulk element concentration in shoots and roots of the facultative nickel-hyperaccumulator *Alyssoides*  
306 *utriculata* harvested from Serpentine (S) and non-serpentine (NS) sites. Data are mean±SD. n= 5 each site, each plant  
307 sample. Significant differences were marked with different superscript letters in the same column (P < 0.05).

308

309 In the serpentine site, the Bioaccumulation Factor (BF) for Ni is 0.8, while the Translocation Factor  
310 (TF) is greater than 6.

311

Strain name	Microbiota type	NCBI accession number	Taxonomy rank	Most closely related species (sequence similarity)	Soil compartment
SERP1	Bacterium	MG661811	Genus	<i>Pseudomonas</i> sp. (99%)	SR
SERP2	Bacterium	MG661822	Genus	<i>Stenotrophomonas</i> sp. (99%)	SR, NSR
SERP3	Bacterium	MG661833	Genus	<i>Streptomyces</i> sp. (99%)	SR
SERP4	Bacterium	MG661835	Genus	<i>Streptomyces</i> sp. (99%)	SR, SB, NSR, NSB
SERP5	Bacterium	MG661836	Family	<i>Enterobacteriaceae</i>	SR
SERP6	Bacterium	MG661837	Genus	<i>Bacillus</i> sp. (99%)	SR, SB, NSR, NSB
SERP7	Bacterium	MG661838	Genus	<i>Bacillus</i> sp. (99%)	SR, SB
SERP8	Bacterium	MG661839	Genus	<i>Bacillus</i> sp. (99%)	SR, SB, NSR, NSB
SERP9	Bacterium	MG661840	Genus	<i>Bacillus</i> sp. (99%)	SB
SERP10	Bacterium	MG661812	Genus	<i>Bacillus</i> sp. (99%)	SB, NSR, NSB
SERP11	Bacterium	MG661813	Genus	<i>Arthrobacter</i> sp. (99%)	SR, SB, NSR, NSB
SERP12	Bacterium	MG661814	Genus	<i>Bacillus</i> sp. (99%)	SR, SB, NSR, NSB

SERP13	Bacterium	MG661815	Genus	<i>Chryseobacterium</i> sp. (99%).	NSR
SERP14	Bacterium	MG661816	Genus	<i>Pantoea</i> sp. (99%)	SR, NSR
SERP15	Bacterium	MG661817	Genus	<i>Micrococcus</i> sp. (99%)	NSB
SERP16	Bacterium	MG661818	Genus	<i>Bacillus</i> sp. (99%)	NSB
SERP17	Bacterium	MG661819	Genus	<i>Streptomyces</i> sp. (99%)	SR, SB, NSR, NSB
SERP18	Bacterium	MG661820	Family	<i>Enterobacteriaceae</i>	SB, NSR
SERP19	Bacterium	MG661821	Genus	<i>Erwinia</i> sp. (99%)	SR, SB, NSR, NSB
SERP20	Bacterium	MG661823	Genus	<i>Flavobacterium</i> sp. (98%)	NSR
SERP21	Bacterium	MG661824	Genus	<i>Novosphingobium</i> sp. (100%)	SR, SB, NSR
SERP22	Bacterium	MG661825	Genus	<i>Curtobacterium</i> sp. (99%)	NSR
SERP23	Bacterium	MG661826	Genus	<i>Streptomyces</i> sp. (99%)	SR, SB, NSR, NSB
SERP24	Bacterium	MG661827	Genus	<i>Leucobacter</i> sp. (99%)	NSR
SERP25	Bacterium	MG661828	Genus	<i>Pantoea</i> sp. (100%)	SR, NSR
SERP26	Bacterium	MG661829	Genus	<i>Micrococcus</i> sp. (100%)	SR, SB, NSR, NSB
SERP27	Bacterium	MG661830	Genus	<i>Cronobacter</i> sp. (99%)	NSR
SERP28	Bacterium	MG661831	Genus	<i>Stenotrophomonas</i> sp. (99%)	NSR
SERP29	Bacterium	MG661832	Genus	<i>Sphingobacterium</i> sp. (100%)	NSR
SERP30	Bacterium	MG661834	Genus	<i>Massilia</i> sp. (100%)	NSR
Serp01S	Fungus	MG836709	Species	<i>Cladosporium cladosporioides</i> (100 %)	SR, SB
Serp03S	Fungus	MG850978	Species	<i>Penicillium ochrochloron</i> (99 %)	SR, SB
Serp04S	Fungus	MG850979	Species	<i>Penicillium canescens</i> (100 %)	SR, NSR
Serp05S	Fungus	MG836710	Species	<i>Trichoderma harzianum</i> (99 %)	SR, SB, NSB
Serp06S	Fungus	MG850980	Species	<i>Aspergillus niger</i> (100 %)	SB, NSB
Serp08S	Fungus	MG850981	Species	<i>Penicillium lanosum</i> (98 %)	NSR
Serp11S	Fungus	MG850982	Species	<i>Penicillium atramentosum</i> (99 %)	NSR, NSB
Serp13S	Fungus	MG850983	Species	<i>Penicillium canescens</i> (99 %)	SR, NSR

312

313 **Table 3** The diversity (strain name, accession number, *taxon*) of isolated culturable bacteria and fungi from the  
314 rhizosphere (R, n= 5) of *Alyssoides utriculata* growing on serpentine (S, n= 5) and non-serpentine soil (NS, n= 5) and  
315 from the adjacent bare soil (B, n= 5).

316



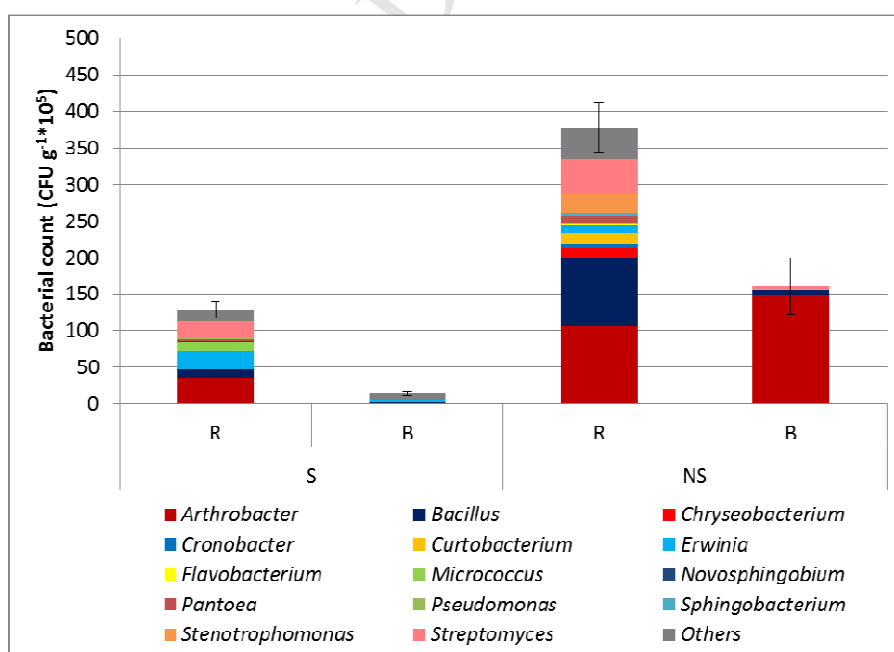
## 317 3.2 Isolation and identification of culturable bacteria

318 In the R soil the bacterial colonies from S and NS site are significantly more abundant than B soils  
 319 (Figure 1). Bacterial counts from NS site are higher than that from S site.

320 Among all the isolated bacteria, 30 recurring morphotypes are counted and identified with a  
 321 morphological approach followed by a molecular identification. Among the isolates, the most  
 322 frequent strain is *Arthrobacter* which accounts for about 30% of average bacterial count of the  
 323 rhizospheric S site (SR). Besides, there is a prevalence of *Enterobacteriaceae* which accounts for  
 324 about 20% of the total bacterial colonies isolated. Among others isolates, the most recurrent are  
 325 *Bacillus*, *Erwinia*, and *Streptomyces*. The bacterial count indicates that *Novosphingobium*, *Pantoea*,  
 326 and *Stenotrophomonas* are present in the SR and NSR soil, while the metallophilic *Pseudomonas*  
 327 thrives only in SR soil (Table 3).

328 Post-hoc test reveals a very significant difference between SR and NSR soils ( $P < 0.001$ ), and  
 329 between NSR and NSB soils ( $P < 0.001$ ), in terms of total culturable bacterial count ( $\text{CFU g}^{-1}$ ).

330



331

332 **Figure 1.** Average bacterial count expressed as Colony Forming Unit ( $\text{CFU g}^{-1}$ ) and diversity (genus, comprising  
 333 different strains) in the rhizospheric (R) soil of facultative Ni-hyperaccumulator *A. utriculata* and adjacent bare (B) soil  
 334 on serpentine (S) and non-serpentine (NS) site. Data are mean $\pm$ SD. N=90 each site, each soil type.

335

336 *3.3 PGPR traits and Ni-tolerance in bacteria*

337 Among the serpentine bacterial isolates, 8 strains belonging to 5 genera show to possess more than  
 338 one PGPR activity (Table 4): *Pseudomonas*, *Stenotrophomonas*, *Streptomyces*, *Pantoea* and  
 339 *Erwinia*. Specifically, strain SERP14 and SERP25 (*Pantoea* sp.) have all four traits and they exhibit  
 340 the highest production of IAA at 530 nm of absorbance.

341 SERP1 (*Pseudomonas* sp.), SERP2 (*Stenotrophomonas* sp.), and SERP19 (*Erwinia* sp.) show a  
 342 great synthesis of siderophores as well as *Pantoea* previously cited, while SERP3, SERP4 and  
 343 SERP23 (*Streptomyces* sp.) can solubilize phosphate as SERP19 (*Erwinia* sp.) and genus *Pantoea*  
 344 and to grow on ACC as the sole source of N.

345 All the bacterial colonies considered highlight a metal-tolerance at low Ni concentrations (up to 5  
 346 mM of Ni), Table 5. Half of these isolates tolerate concentrations up to 10 mM of Ni and only  
 347 SERP01 (*Pseudomonas* sp.) and SERP04 (*Streptomyces* sp.) can be cultivated on Ni 15 mM.

348

Strain name	Microbiota type	ACC deaminase <sup>a</sup>	IAA synthesis (mg l <sup>-1</sup> )	Siderophores Production <sup>b</sup> (halo Ø, mm)	Phosphate solubilization <sup>c</sup> (halo Ø, mm)
SERP1	<i>Pseudomonas</i> sp.	-	6.0±0.1	+ (90)	+ (5)
SERP2	<i>Stenotrophomonas</i> sp.	-	4.8±0.1	+ (90)	-
SERP3	<i>Streptomyces</i> sp.	+	-	-	+ (3)
SERP4	<i>Streptomyces</i> sp.	+	-	+ (12)	+ (3)
SERP6	<i>Bacillus</i> sp.	-	-	-	-
SERP7	<i>Bacillus</i> sp.	-	-	+ (15)	-
SERP8	<i>Bacillus</i> sp.	-	-	-	-
SERP11	<i>Arthrobacter</i> sp.	-	-	+ (45)	-
SERP12	<i>Bacillus</i> sp.	-	-	-	-
SERP14	<i>Pantoea</i> sp.	+	10.2±0.2	+ (90)	+ (4)
SERP19	<i>Erwinia</i> sp.	-	-	+ (90)	+ (5)

SERP21	<i>Novosphingobium</i> sp.	-	-	-	+ (2)
SERP23	<i>Streptomyces</i> sp.	+	-	-	+ (2)
SERP25	<i>Pantoea</i> sp.	+	20.5±0.3	+ (90)	+ (15)
SERP26	<i>Micrococcus</i> sp.	-	-	-	-

349

350 **Table 4** Plant growth promoting traits of culturable bacteria isolated from the rhizosphere of *A. utriculata* on Serpentine351 soil (S), n= 8 each strain. The halo refers to Petri dishes of Ø 90 mm. <sup>a</sup> Growth on ACC: - none; + turbidity; <sup>b</sup>352 Siderophores production: - no halo; + halo; <sup>c</sup>Phosphate solubilization: - no halo; + halo.

353

Strain name	Microbiota type	Ni 1 mM	Ni 5 mM	Ni 10 mM	Ni 15 mM	Ni 20 mM
SERP1	<i>Pseudomonas</i> sp.	+	+	+	+	-
SERP2	<i>Stenotrophomonas</i> sp.	+	+	+	-	-
SERP3	<i>Streptomyces</i> sp.	+	+	+	-	-
SERP4	<i>Streptomyces</i> sp.	+	+	+	+	-
SERP14	<i>Pantoea</i> sp.	+	+	-	-	-
SERP19	<i>Erwinia</i> sp.	+	+	-	-	-
SERP23	<i>Streptomyces</i> sp.	+	+	-	-	-
SERP25	<i>Pantoea</i> sp.	+	+	-	-	-

354

355 **Table 5** Nickel tolerance (mM) of culturable PGPR isolated from the rhizosphere of *A. utriculata* growing on the

356 Serpentine soil (S), n= 15 each strain. Growth: - absence; + presence

357

358 

### 3.4 Isolation and identification of culturable fungi

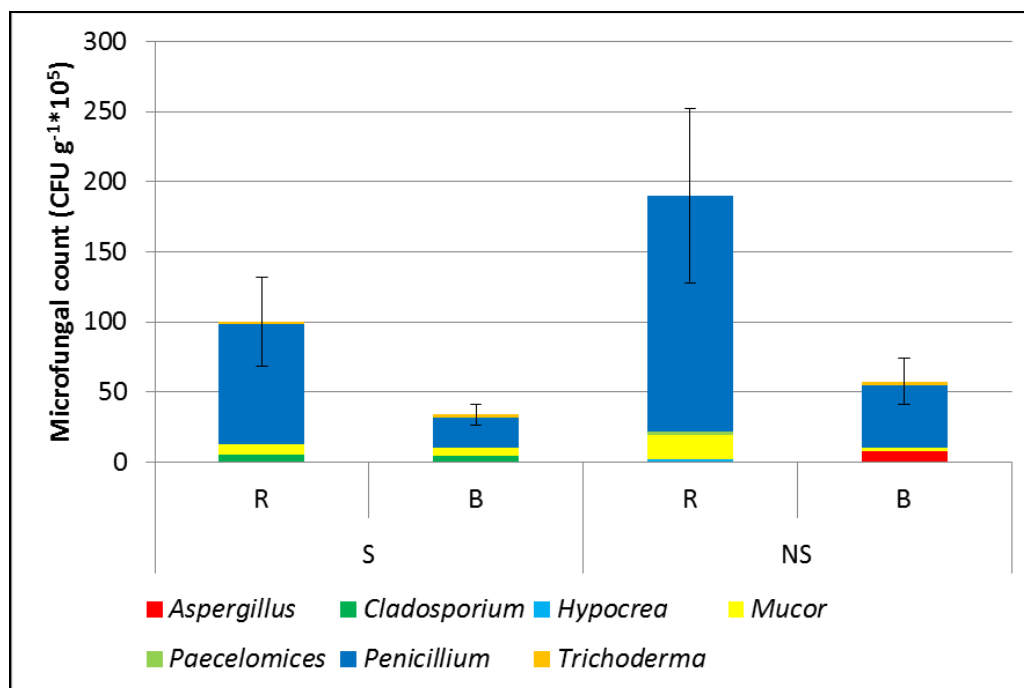
359 Figure 2 summarizes the fungal presence in S and NS sites both in R and B soils. Fungi preferably

360 colonize NSR soil respect to SR. The most recurrent genera are *Aspergillus*, *Penicillium*,361 *Cladosporium*, and *Trichoderma* (Table 3).362 *Penicillium* is the most abundant strain comprising the 84% of total colonies, and about 85% of363 isolates from SR soils. Although genus *Cladosporium* has a low frequency (2% of the total), it was

364 exclusively isolated from S soils. We observed a slightly greater abundance of microorganisms in R

365 than B soil and in NS than S site. Post-hoc test reveals a significant difference ( $P < 0.01$ ) between  
 366 NSR and NSB.

367



368

369 **Figure 2.** Microfungal count expressed as Colony Forming Unit (CFU g<sup>-1</sup>) and diversity (genus, comprising different  
 370 strains) in the rhizospheric (R) soil of facultative Ni-hyperaccumulator *A. utriculata* and adjacent bare (B) soil on  
 371 serpentine (S) and non-serpentine (NS) site. Data are mean±SD. N= 20 each site, each soil type.

372

### 373 3.5 Ni-tolerance in fungi

374 Among the fungal strains isolated from S site, *Cladosporium cladosporoides* (Fresen.) G.A. de  
 375 Vries does not grow in the presence of Ni, *Trichoderma harzianum* and *Penicillium canescens*  
 376 tolerate up to 5 mM of NiSO<sub>4</sub>\*6H<sub>2</sub>O, while *Penicillium ochrochloron* is able to tolerate up to 20  
 377 mM (Table 6).

378

Strain name	Species	Ni 0	Ni 1	Ni 5	Ni 10	Ni 15	Ni 20
Serp01S	<i>Cladosporium cladosporoides</i>	+	-	-	-	-	-
Serp03S	<i>Penicillium ochrochloron</i>	+	+	+	+	+	+

<b>Serp04S</b>	<i>Penicillium canescens</i>	+	+	+	-	-	-
<b>Serp05S</b>	<i>Trichoderma harzianum</i>	+	+	+	-	-	-

379

380 **Table 6.** Nickel tolerance (mM) of culturable fungi isolated from the rhizosphere of *A. utriculata* growing on the  
381 Serpentine soil (S), n= 15 each strain.

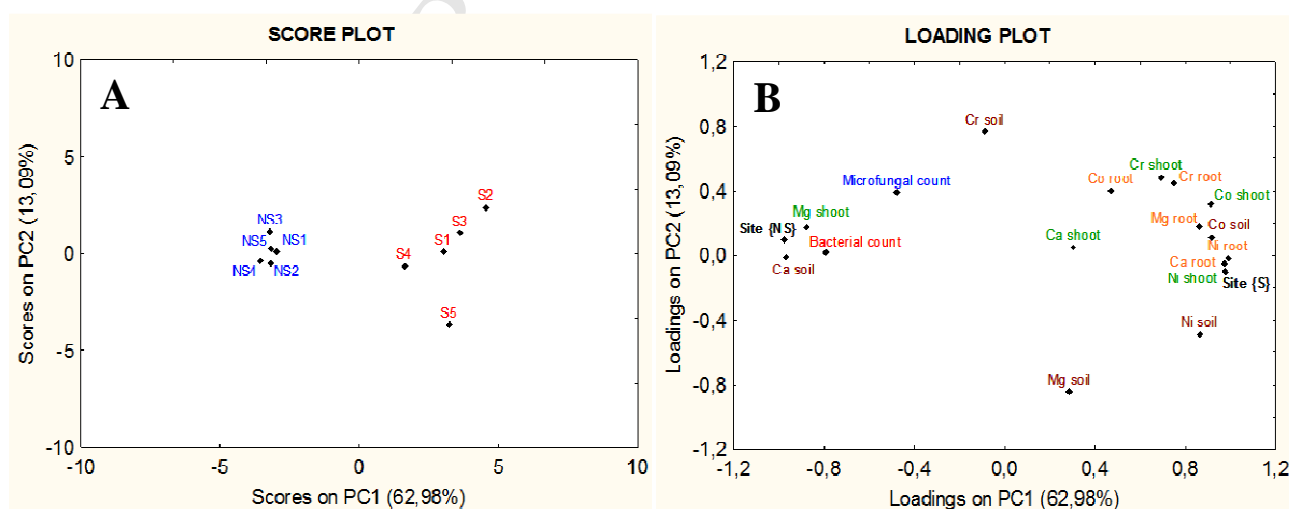
382

383

### 384 3.6 Plant and rhizosphere microbiota response to soil metals

385 The S and NS samples constitute two well-defined cluster, highlighting a good homogeneity of the  
386 starting samples (Fig. 3A). The Loadings plot in Figure 3B show the samples data (element  
387 concentration in plant and soil and microbiota count) distribution on the first two principal  
388 components. Most of the elements in soil, root and shoot (especially Ni and Co) is associated with  
389 positive scores on PC1. Both Loading plots presented explain a total variance of 76.07%. It is  
390 noteworthy that Ca is the only soil element associated with negative scores of PC1; indeed, S soil  
391 samples are characterized by scarcity of nutrients such as Ca and presence of phytotoxic elements  
392 such as Ni and Co. Microfungal and bacterial counts are associated with negative scores on PC1,  
393 without a clear response respect to soil element.

394



395

396

397 **Figure 3.** Score plot and Loading plot of a Principal Component Analysis. A) Score plot of 10 R soil samples  
398 distributed in S (red) and NS (blue) sites; B) loading plot of the bacterial and the fungal counts and element  
399 concentrations in soil and plant. The samples are distinguished by different colors depending on the origin, brown: soil,  
400 orange: root, green: shoot, red: bacteria, blue: fungi. Factor 1 (62.98% variation explained) vs Factor 2 (13.09%  
401 variation explained). Variables are indicated by symbol ♦.

402

403

#### 404 **4. Discussion**

405 Only 10% of hyperaccumulators have their rhizosphere examined (Alford et al., 2010; Visioli et al.,  
406 2015). Up to date facultative hyperaccumulators have been poorly studied (Pollard et al., 2014) and  
407 no facultative Ni-hyperaccumulator Mediterranean plant species have been studied in combination  
408 with their rhizosphere microbiota for phytoremediation purposes.

409 The differences between serpentine and non-serpentine sites in terms of element concentration is  
410 significant because of the serpentine higher level of Ni, Co, Cr, the Ca-deficiency, and the low  
411 Ca/Mg ratio (0.36) typical of harsh serpentinitic environmental conditions (Ghasemi et al., 2015;  
412 Kidd et al., 2018). The differences in terms of leaf Ni and Co concentration, in *A. utriculata*,  
413 between sites, could reflect soil bioavailability of these elements (Reeves, 2006; Ghasemi and  
414 Ghaderian, 2009). As expected, only *A. utriculata* from serpentine soils shows typical  
415 hyperaccumulator traits (shoot Ni >1000 mg kg<sup>-1</sup>; TF>>1, BF~1)(Reeves, 1992; van der Ent et al.,  
416 2013).

417 Even the culturable microbiota of the facultative Ni-hyperaccumulator *A. utriculata* seems to be  
418 limited by the serpentine soil conditions resulting less abundant than in non-serpentine site with  
419 preferential colonization of rhizospheric soil. The higher microbiota density in serpentine  
420 rhizosphere respect to bare soil is a common feature of hyperaccumulators (Mengoni et al., 2001;  
421 Wenzel et al., 2004; Álvarez-López et al., 2016; Lopes et al., 2016; Benizri and Kidd, 2018). This  
422 result partially reflects the

423 'rhizosphere effect'. It implies that the plant exuding a large number of compounds that can be  
424 used as nutrition sources by microbes to proliferate and colonize the root surrounding area (Morgan  
425 et al., 2005; Rovira, 1965; Segura et al., 2009; Smalla et al., 2001; Wenzel et al., 2004). Among the  
426 identified bacterial morphotypes, genera *Arthrobacter*, *Bacillus*, *Erwinia*, *Micrococcus*,  
427 *Novosphingobium*, *Pantoea*, *Pseudomonas*, and *Streptomyces* are distinguished as they have already  
428 been isolated on serpentine soil from the rhizosphere of Ni-hyperaccumulating plants (Mengoni et  
429 al., 2001; Abou-Shanab et al., 2003; Idris et al., 2004; Ma et al., 2009; Turgay et al., 2012; Visioli et  
430 al., 2015).

431 Bacterial and fungal strains are limited by serpentine soil conditions. This is probably due to the  
432 highly spatial and temporal heterogeneous soil structure (Abbott et al., 2015; Jansson and  
433 Hofmockel, 2018) which is affected by elements and nutrients distribution (i.e. C, N, P, Xue et al.,  
434 2018; Zhang et al., 2018). We cannot exclude that this bacterial and fungal community, native of  
435 serpentinitic soil may induce promoting effects on the growth of *A. utriculata*, immobilizing or  
436 uptaking Ni, as already demonstrated in *N caerulea* (Abouddrar et al., 2013). Further studies will  
437 clarify this point. The presence of some PGPR bacterial strains like *Pseudomonas* sp. and  
438 *Streptomyces* sp. that can be cultivated on Ni 15 mM and microfungi able to grow on Ni up to 3  
439 mM (Cecchi et al., 2017b) could support this idea.

440 Most of culturable bacterial strains exhibit one or more plant growth-promoting activities that  
441 potentially support the growth of *Alyssoides utriculata*, reducing metal stress and increasing Ni  
442 uptake and accumulation in aerial parts. One third of serpentinitic isolates shows ACC deaminase  
443 that determines the reduction of ethylene stress with inhibition of plant development at rhizosphere  
444 level (Glick, 2010; Glick and Stearns, 2011; Sessitsch et al., 2013) and produces IAA that enhances  
445 the root surface area and plants nutrients uptake (Shagol et al., 2014). Moreover, the synthesis of  
446 IAA can stimulate plant cell proliferation and cell elongation can induce the transcription of the  
447 enzyme that catalyses the synthesis of ACC (ACC sintase) (Khan et al., 2015).

448 Five bacterial strains produce siderophores and this could suggest the ability to alter metal solubility  
449 in the rhizosphere stimulating the plant growth (Lodewyckx et al., 2002; Idris et al., 2004; Tripathi  
450 et al., 2005; Barzanti et al., 2007; Ma et al., 2009b) but further investigations can clarify this point.  
451 Half of the tested bacteria is positive for the phosphate solubilization activity which increase the  
452 availability of P for the plant (Fitz and Wenzel, 2002) and the metal solubility through the bacterial  
453 phosphatase-mediated dissolution of metal phosphates (Alford et al., 2010). In addition, our study  
454 shows that the PGPR strains isolated on serpentine soils often exhibit a high tolerance to Ni (up to  
455 15 mM in the culture medium) according to Schlegel et al. (1991). We found *Pseudomonas* only in  
456 the rhizospheric soil of the S site suggesting potential metallophilic behavior. This genus is known  
457 for its PGP traits (Li and Ramakrishna, 2011; Ma et al., 2011) and Ni tolerance (growing up to 10  
458 mM Ni, Turgay et al. 2012). Previous studies revealed a high biomass yield in *Brassica* spp.  
459 inoculated with a serpentine selected strain of *Pseudomonas* (Freitas et al. 2004), highlighting the  
460 protective role of this bacterium against the inhibitory effects of Ni (Rajkumar and Freitas, 2008a;  
461 Ma et al., 2009). Moreover, an increased uptake of Ni, Cu and Zn was recorded in the root system  
462 of *Ricinus communis* L. inoculated with *Pseudomonas* (Rajkumar and Freitas, 2008b). Inoculation  
463 of metal-tolerant serpentinic strains such as *Bacillus* sp. and *Pseudomonas* sp. seems to be  
464 effective, directly by increasing the accumulation of metal in plant tissues and indirectly stimulating  
465 of plant growth, protecting the plant from the inhibition of growth given by soil Ni, thus reducing  
466 its toxicity (Ma et al., 2009b; Rajkumar et al., 2013; Rajkumar and Freitas, 2008a; Zaidi et al.,  
467 2006).

468 Although genera *Arthrobacter*, *Bacillus*, and *Streptomyces* were isolated from all soil type, some  
469 authors have shown their beneficial effects on different plants growing on disturbed soils types  
470 (Dimkpa et al., 2008, 2009; Gray and Smith, 2005; Y. Ma et al., 2011). All these aspects suggest  
471 that specific strains showing good activity, colonization potential and PGPR traits will be useful in  
472 enhancing bioavailability, phytoextraction and phytostabilization performance by plants (Sessitsch et  
473 al., 2013).



474 The lower amount of microfungus strains in bare respect to rhizosphere soils can be related to the  
475 lower availability of nutrients, water, and oxygen commonly provided by the rhizosphere  
476 (Söderström, 1975; Kjøller and Struwe, 1982). The isolated microfungus strains of *Penicillium*,  
477 *Aspergillus*, *Trichoderma*, and *Mucor* are commonly saprotrophic microfungi with species found in  
478 metal-polluted habitats worldwide (Kubatova et al., 2002; Massaccesi et al., 2002; Gadd and  
479 Fomina, 2011). Among *Penicillium*, *Mucor*, *Cladosporium*, and *Trichoderma* isolated on serpentine  
480 soils, only *Trichoderma* strain, belonging to the *harzianum* group, was also isolated from the  
481 rhizosphere of *A. utriculata* in metal-contaminated sites. This work confirmed the capability of this  
482 fungal strain to uptake Ni as described by Roccotiello et al. (2015b) and Cecchi et al. (2017b). *T.*  
483 *harzianum* was known to hyperaccumulate Ni up to 500 mg L<sup>-1</sup> (Roccotiello et al. 2015b) and other  
484 metals (Cecchi et al., 2017a, 2017b; Zotti et al., 2014) with an amazing bioaccumulation capacity  
485 (up to 11,000 mg Ni kg<sup>-1</sup>) (Cecchi et al., 2017b). Moreover, *Penicillium canescens* and *P.*  
486 *ochrochloron* also showed the capacity to tolerate Ni up to 5mM and 20 mM, respectively. The first  
487 is a typical rhizospheric species able to live until 50 cm depth in soil, while the second is known for  
488 its high tolerance to copper and other metals (Domsch et al., 2007). Besides, *P. expansum*  
489 represents a cosmopolitan species able to colonize polluted soils and extreme environments and to  
490 bioaccumulate high metal concentrations (Cecchi et al., 2017b; Di Piazza et al., 2017). Moreover,  
491 *Aspergillus* strain belonging to the *niger* group represents one of the more studied fungus about  
492 biocorrosion, bioalteration, and bioaccumulation of toxic metals, due to the high amount of  
493 secondary organic acid production (Gadd, 2007).

494 Although the culturable techniques are not representative of the total phylogenetic diversity of soil  
495 microbiota (i.e., less than 1% of bacteria can be cultured in laboratory - Pham and Kim, 2012), the  
496 characterization of culturable bacterial and fungal strains is essential for future field studies using  
497 bioinoculants as natural chelators in the rhizosphere of hyperaccumulator and facultative  
498 hyperaccumulator species. The microbiota characterization of the facultative hyperaccumulator  
499 species plays a key role because this could help isolate the culturable bacteria and fungi that

500 contribute to the hyperaccumulator phenotype (Visioli et al., 2015), using microbiota as inoculum  
501 of the rhizosphere to improve the root development, the metal uptake, and the phytoremediation of  
502 metal contaminated soils. The rhizosphere of *A. utriculata* can be an excellent model to enhance Ni  
503 uptake and a better growth using a microbiota associated with its root system.

504

505

## 506 **5. Conclusion**

507 The characterization of the culturable microbiota of the facultative Ni-hyperaccumulator *Alyssoides*  
508 *utriculata* provide a collection of bacterial and fungal strains for potential applications in Ni  
509 phytoextraction.

510 Bacteria and fungi were significantly more abundant in rhizosphere than in bare soil and were  
511 dominated by genera *Arthrobacter*, *Bacillus*, *Erwinia*, *Micrococcus*, *Novosphingobium*, *Pantoea*,  
512 *Pseudomonas*, and *Streptomyces*, *Penicillium*, *Aspergillus*, *Trichoderma*, and *Mucor*. Interestingly,  
513 the genus *Pseudomonas* was only found in rhizospheric serpentine soils and with *Streptomyces* sp.  
514 showed highest Ni-tolerance up to 15 mM. Among serpentine bacterial isolates, eight strains  
515 belonging to five genera showed at least one PGPR activity. Similarly, *P. ochrochloron* and  
516 *Trichoderma* strain belonging to the *harzianum* group exhibit great Ni-tolerance. This latter also has  
517 a high Ni uptake ability.

518 The rhizosphere of *A. utriculata* can be an excellent model to enhance Ni uptake and plant growth  
519 using a microbiota associated with its root system. This multidisciplinary research on the dynamic  
520 microenvironment known as the rhizosphere has the perspective to provide guidance for using Ni-  
521 hyperaccumulator species and the associated rhizobiota to remediate Ni contaminated soils.

522

523

## 524 **6. Acknowledgements**

525 This research was performed in the frame of the PhD in Biology applied to Agriculture and Environment,  
526 (DISTAV-Department of Earth, Environmental and Life Sciences University of Genoa, Italy), XXXI cycle.

527 Authors wish to thank C. Grande, G. Tassistro and A. Borello, Laboratory of Microbiology of DISTAV-  
528 University of Genoa, for their technical support.

529

530

## 531 7. References

532 Abbott, K.C., Karst, J., Biederman, L.A., Borrett, S.R., Hastings, A., Walsh, V., Bever, J.D., 2015. Spatial  
533 Heterogeneity in Soil Microbes Alters Outcomes of Plant Competition. PLOS ONE 10, e0125788.

534 <https://doi.org/10.1371/journal.pone.0125788>

535 Aboudrar, W., Schwartz, C., Morel, J.L., Boularbah, A., 2013. Effect of nickel-resistant rhizosphere bacteria  
536 on the uptake of nickel by the hyperaccumulator *Noccaea caerulea* under controlled conditions.

537 J. Soils Sediments 13, 501–507. <https://doi.org/10.1007/s11368-012-0614-x>

538 Abou-Shanab, R., Angle, J., Delorme, T., Chaney, R., Van Berkum, P., Moawad, H., Ghanem, K., Ghazlan,  
539 H., 2003. Rhizobacterial effects on nickel extraction from soil and uptake by *Alyssum murale*. New

540 Phytol. 158, 219–224. <https://doi.org/10.1046/j.1469-8137.2003.00721.x>

541 Alford, É.R., Pilon-Smits, E.A.H., Paschke, M.W., 2010. Metallophytes—a view from the rhizosphere. Plant  
542 Soil 337, 33–50. <https://doi.org/10.1007/s11104-010-0482-3>

543 Álvarez-López, V., Prieto-Fernández, Á., Becerra-Castro, C., Monterroso, C., Kidd, P.S., 2016.  
544 Rhizobacterial communities associated with the flora of three serpentine outcrops of the Iberian  
545 Peninsula. Plant Soil 403, 233–252. <https://doi.org/10.1007/s11104-015-2632-0>

546 Amir, H., Lagrange, A., Hassaine, N., Cavaloc, Y., 2013. Arbuscular mycorrhizal fungi from New  
547 Caledonian ultramafic soils improve tolerance to nickel of endemic plant species. Mycorrhiza 23,  
548 585–595. <https://doi.org/10.1007/s00572-013-0499-6>

549 Barzanti, R., Ozino, F., Bazzicalupo, M., Gabrielli, R., Galardi, F., Gonnelli, C., Mengoni, A., 2007.  
550 Isolation and Characterization of Endophytic Bacteria from the Nickel Hyperaccumulator Plant  
551 *Alyssum bertolonii*. Microb. Ecol. 53, 306–316. <https://doi.org/10.1007/s00248-006-9164-3>

552 Becerra-Castro, C., Monterroso, C., García-Lestón, M., Prieto-Fernández, A., Acea, M.J., Kidd, P.S., 2009.  
553 Rhizosphere Microbial Densities and Trace Metal Tolerance of the Nickel Hyperaccumulator

- 554 *Alyssum Serpyllifolium* Subsp. *Lusitanicum*. Int. J. Phytoremediation 11, 525–541.  
555 <https://doi.org/10.1080/15226510902717549>
- 556 Benizri, E., Baudoin, E., Guckert, A., 2001. Root Colonization by Inoculated Plant Growth-Promoting  
557 Rhizobacteria. Biocontrol Sci. Technol. 11, 557–574. <https://doi.org/10.1080/09583150120076120>
- 558 Benizri, E., Kidd, P.S., 2018. The role of the rhizosphere and microbes associated with hyperaccumulator  
559 plants in metal accumulation, in: Agromining: Farming for Metals. Springer, pp. 157–188.
- 560 Bric, J.M., Bostock, R.M., Silverstone, S.E., 1991. Rapid *In Situ* Assay for Indoleacetic Acid Production by  
561 Bacteria Immobilized on a Nitrocellulose Membrane. Appl. Environ. Microbiol. 57, 535–538.
- 562 Cabello-Conejo, M.I., Becerra-Castro, C., Prieto-Fernández, A., Monterroso, C., Saavedra-Ferro, A., Mench,  
563 M., Kidd, P.S., 2014. Rhizobacterial inoculants can improve nickel phytoextraction by the  
564 hyperaccumulator *Alyssum pintodasilvae*. Plant Soil 379, 35–50. [https://doi.org/10.1007/s11104-](https://doi.org/10.1007/s11104-014-2043-7)  
565 [014-2043-7](https://doi.org/10.1007/s11104-014-2043-7)
- 566 Capponi, G., Crispini, L., 2008. Note illustrative alla Carta Geologica d'Italia alla scala 1:50.000. Foglio 213  
567 - 230 “Genova”.
- 568 Cecchi, G., Marescotti, P., Piazza, S.D., Zotti, M., 2017a. Native fungi as metal remediators: Silver myco-  
569 accumulation from metal contaminated waste-rock dumps (Libiola Mine, Italy). J. Environ. Sci.  
570 Health Part B 52, 191–195. <https://doi.org/10.1080/03601234.2017.1261549>
- 571 Cecchi, G., Roccotiello, E., Piazza, S.D., Riggi, A., Mariotti, M.G., Zotti, M., 2017b. Assessment of Ni  
572 accumulation capability by fungi for a possible approach to remove metals from soils and waters. J.  
573 Environ. Sci. Health Part B 52, 166–170. <https://doi.org/10.1080/03601234.2017.1261539>
- 574 Chaney, R.L., Angle, J.S., Broadhurst, C.L., Peters, C.A., Tappero, R.V., Sparks, D.L., 2007. Improved  
575 Understanding of Hyperaccumulation Yields Commercial Phytoextraction and Phytomining  
576 Technologies. J. Environ. Qual. 36, 1429–1443. <https://doi.org/10.2134/jeq2006.0514>
- 577 Charlot, G., 1964. Colorimetric Determination of Elements, Principles and Methods: By G. Charlot. Elsevier.
- 578 Comerford, N.B., 2005. Soil Factors Affecting Nutrient Bioavailability, in: Nutrient Acquisition by Plants,  
579 Ecological Studies. Springer, Berlin, Heidelberg, pp. 1–14. [https://doi.org/10.1007/3-540-27675-0\\_1](https://doi.org/10.1007/3-540-27675-0_1)
- 580 de Souza, R., Ambrosini, A., Passaglia, L.M.P., 2015. Plant growth-promoting bacteria as inoculants in  
581 agricultural soils. Genet. Mol. Biol. 38, 401–419. <https://doi.org/10.1590/S1415-475738420150053>

- 582 Di Piazza, S., Cecchi, G., Cardinale, A.M., Carbone, C., Mariotti, M.G., Giovine, M., Zotti, M., 2017.  
583 *Penicillium expansum* Link strain for a biometallurgical method to recover REEs from WEEE.  
584 Waste Manag., Special Thematic Issue: Urban Mining and Circular Economy 60, 596–600.  
585 <https://doi.org/10.1016/j.wasman.2016.07.029>
- 586 Dimkpa, C. o., Merten, D., Svatoš, A., Büchel, G., Kothe, E., 2009. Siderophores mediate reduced and  
587 increased uptake of cadmium by *Streptomyces tendae* F4 and sunflower (*Helianthus annuus*),  
588 respectively. J. Appl. Microbiol. 107, 1687–1696. <https://doi.org/10.1111/j.1365-2672.2009.04355.x>
- 589 Dimkpa, C., Svatoš, A., Merten, D., Büchel, G., Kothe, E., 2008. Hydroxamate siderophores produced by  
590 *Streptomyces acidiscabies* E13 bind nickel and promote growth in cowpea (*Vigna unguiculata* L.)  
591 under nickel stress. Can. J. Microbiol. 54, 163–172. <https://doi.org/10.1139/W07-130>
- 592 Domsch, K.H., Gams, W., Anderson, T.H., 2007. Compendium of soil fungi, 2nd taxonomically revised  
593 edition by W. Gams IHW Eching.
- 594 Durand, A., Piutti, S., Rue, M., Morel, J.L., Echevarria, G., Benizri, E., 2016. Improving nickel  
595 phytoextraction by co-cropping hyperaccumulator plants inoculated by plant growth promoting  
596 rhizobacteria. Plant Soil 399, 179–192. <https://doi.org/10.1007/s11104-015-2691-2>
- 597 Dworkin, M., Foster, J.W., 1958. Experiments With Some Microorganisms Which Utilize Ethane And  
598 Hydrogen. J. Bacteriol. 75, 592–603.
- 599 Ent, A. van der, Baker, A.J.M., Reeves, R.D., Pollard, A.J., Schat, H., 2013. Hyperaccumulators of metal and  
600 metalloids trace elements: Facts and fiction. Plant Soil 362, 319–334. [https://doi.org/10.1007/s11104-](https://doi.org/10.1007/s11104-012-1287-3)  
601 [012-1287-3](https://doi.org/10.1007/s11104-012-1287-3)
- 602 Fitz, W.J., Wenzel, W.W., 2002. Arsenic transformations in the soil–rhizosphere–plant system: fundamentals  
603 and potential application to phytoremediation. J. Biotechnol., Highlights from ECB10 - Novel  
604 Bioactive Substances and Bioremediation Technologies 99, 259–278. [https://doi.org/10.1016/S0168-](https://doi.org/10.1016/S0168-1656(02)00218-3)  
605 [1656\(02\)00218-3](https://doi.org/10.1016/S0168-1656(02)00218-3)
- 606 Freitas, H., Prasad, M.N.V., Pratas, J., 2004. Analysis of serpentinophytes from north–east of Portugal for  
607 trace metal accumulation—relevance to the management of mine environment. Chemosphere 54,  
608 1625–1642. <https://doi.org/10.1016/j.chemosphere.2003.09.045>

- 609 Gadd, G.M., 2007. Geomycology: biogeochemical transformations of rocks, minerals, metals and  
610 radionuclides by fungi, bioweathering and bioremediation. *Mycol. Res.* 111, 3–49.  
611 <https://doi.org/10.1016/j.mycres.2006.12.001>
- 612 Gadd, G.M., 1992. Metals and microorganisms: a problem of definition. *FEMS Microbiol. Lett.* 100, 197–  
613 203.
- 614 Gadd, G.M., Fomina, M., 2011. Uranium and Fungi. *Geomicrobiol. J.* 28, 471–482.  
615 <https://doi.org/10.1080/01490451.2010.508019>
- 616 Gardes, M., Bruns, T.D., 1993. ITS primers with enhanced specificity for basidiomycetes-application to the  
617 identification of mycorrhizae and rusts. *Mol. Ecol.* 2, 113–118.
- 618 Ghaderian, S.M., Mohtadi, A., Rahiminejad, R., Reeves, R.D., Baker, A.J.M., 2007. Hyperaccumulation of  
619 nickel by two *Alyssum* species from the serpentine soils of Iran. *Plant Soil* 293, 91–97.  
620 <https://doi.org/10.1007/s11104-007-9221-9>
- 621 Ghasemi, R., Chavoshi, Z.Z., Boyd, R.S., Rajakaruna, N., 2015. Calcium: magnesium ratio affects  
622 environmental stress sensitivity in the serpentine-endemic *Alyssum inflatum* (Brassicaceae). *Aust. J.*  
623 *Bot.* 63, 39–46. <https://doi.org/10.1071/BT14235>
- 624 Ghasemi, R., Ghaderian, S.M., 2009. Responses of two populations of an Iranian nickel-hyperaccumulating  
625 serpentine plant, *Alyssum inflatum* Nyar., to substrate Ca/Mg quotient and nickel. *Environ. Exp. Bot.*  
626 67, 260–268.
- 627 Giammarino, S., Orezzi, S., Rosti, D., Fannucci, F., Morelli, D., De Stefanis, A., 2010. Note Illustrative della  
628 Carta Geologica d'Italia alla scala 1: 50.000-Foglio 258–271 Sanremo.
- 629 Glass, N.L., Donaldson, G.C., 1995. Development of primer sets designed for use with the PCR to amplify  
630 conserved genes from filamentous ascomycetes. *Appl. Environ. Microbiol.* 61, 1323–1330.
- 631 Glick, B.R., 2010. Using soil bacteria to facilitate phytoremediation. *Biotechnol. Adv.* 28, 367–374.  
632 <https://doi.org/10.1016/j.biotechadv.2010.02.001>
- 633 Glick, B.R., Stearns, J.C., 2011. Making Phytoremediation Work Better: Maximizing a Plant's Growth  
634 Potential in the Midst of Adversity. *Int. J. Phytoremediation* 13, 4–16.  
635 <https://doi.org/10.1080/15226514.2011.568533>

- 636 Goswami, D., Thakker, J.N., Dhandhukia, P.C., 2015. Simultaneous detection and quantification of indole-3-  
637 acetic acid (IAA) and indole-3-butyric acid (IBA) produced by rhizobacteria from l-tryptophan (Trp)  
638 using HPTLC. *J. Microbiol. Methods* 110, 7–14. <https://doi.org/10.1016/j.mimet.2015.01.001>
- 639 Gray, E.J., Smith, D.L., 2005. Intracellular and extracellular PGPR: commonalities and distinctions in the  
640 plant–bacterium signalling processes. *Soil Biol. Biochem.* 37, 395–412.  
641 <https://doi.org/10.1016/j.soilbio.2004.08.030>
- 642 Greco, G., Capello, M., Cecchi, G., Cutroneo, L., Di, P., Zotti, M., 2017. Another possible risk for the  
643 Mediterranean Sea? *Aspergillus sydowii* discovered in the Port of Genoa (Ligurian Sea, Italy). *Mar.*  
644 *Pollut. Bull.* 122, 470–474. <https://doi.org/10.1016/j.marpolbul.2017.06.058>
- 645 Hanirah, R., Piakong, M.T., Syaifi, L., 2015. Isolation, characterization and screening of rhizospheric  
646 bacteria of *Pittosferum resiniferum* Hemsl., in: *IOP Conference Series: Materials Science and*  
647 *Engineering*. IOP Publishing, p. 012036.
- 648 Husna, Tuheteru, F.D., Arif, A., 2017. Arbuscular Mycorrhizal Fungi and Plant Growth on Serpentine Soils,  
649 in: *Arbuscular Mycorrhizas and Stress Tolerance of Plants*. Springer, Singapore, pp. 293–303.  
650 [https://doi.org/10.1007/978-981-10-4115-0\\_12](https://doi.org/10.1007/978-981-10-4115-0_12)
- 651 Idris, R., Trifonova, R., Puschenreiter, M., Wenzel, W.W., Sessitsch, A., 2004. Bacterial Communities  
652 Associated with Flowering Plants of the Ni Hyperaccumulator *Thlaspi goesingense*. *Appl. Environ.*  
653 *Microbiol.* 70, 2667–2677. <https://doi.org/10.1128/AEM.70.5.2667-2677.2004>
- 654 Iram, S., Parveen, K., Usman, J., Nasir, K., Akhtar, N., Arouj, S., Ahmad, I., 2012. Heavy metal tolerance of  
655 filamentous fungal strains isolated from soil irrigated with industrial wastewater. *Biologija* 58.  
656 <https://doi.org/10.6001/biologija.v58i3.2527>
- 657 Jansson, J.K., Hofmockel, K.S., 2018. The soil microbiome—from metagenomics to metaphenomics. *Curr.*  
658 *Opin. Microbiol., Environmental Microbiology The New Microscopy* 43, 162–168.  
659 <https://doi.org/10.1016/j.mib.2018.01.013>
- 660 Jing, Y., He, Z., Yang, X., 2007. Role of soil rhizobacteria in phytoremediation of heavy metal contaminated  
661 soils. *J. Zhejiang Univ. Sci. B* 8, 192–207. <https://doi.org/10.1631/jzus.2007.B0192>



- 662 Kelly, J.J., Tate, R.L., 1998. Effects of Heavy Metal Contamination and Remediation on Soil Microbial  
663 Communities in the Vicinity of a Zinc Smelter. *J. Environ. Qual.* 27, 609–617.  
664 <https://doi.org/10.2134/jeq1998.00472425002700030019x>
- 665 Khan, M.U., Sessitsch, A., Harris, M., Fatima, K., Imran, A., Arslan, M., Shabir, G., Khan, Q.M., Afzal, M.,  
666 2015. Cr-resistant rhizo- and endophytic bacteria associated with *Prosopis juliflora* and their  
667 potential as phytoremediation enhancing agents in metal-degraded soils. *Front. Plant Sci.* 5.  
668 <https://doi.org/10.3389/fpls.2014.00755>
- 669 Kidd, P.S., Álvarez-López, V., Becerra-Castro, C., Cabello-Conejo, M., Prieto-Fernández, Á., 2017. Chapter  
670 Three - Potential Role of Plant-Associated Bacteria in Plant Metal Uptake and Implications in  
671 Phytotechnologies, in: Cuypers, A., Vangronsveld, J. (Eds.), *Advances in Botanical Research*,  
672 Phytoremediation. Academic Press, pp. 87–126. <https://doi.org/10.1016/bs.abr.2016.12.004>
- 673 Kidd, P.S., Bani, A., Benizri, E., Gonnelli, C., Hazotte, C., Kissler, J., Konstantinou, M., Kuppens, T.,  
674 Kyrkas, D., Laubie, B., 2018. Developing sustainable agromining systems in agricultural ultramafic  
675 soils for nickel recovery. *Front. Environ. Sci.* 6.
- 676 Kjøller, A., Struwe, S., 1982. Microfungi in ecosystems: fungal occurrence and activity in litter and soil.  
677 *Oikos* 391–422.
- 678 Kubatova, A., Prasil, K., Vanova, M., 2002. Diversity of soil microscopic fungi on abandoned industrial  
679 deposits. *Cryptogam. Mycol.* 23, 205–219.
- 680 Lebeau, T., Braud, A., Jézéquel, K., 2008. Performance of bioaugmentation-assisted phytoextraction applied  
681 to metal contaminated soils: A review. *Environ. Pollut.* 153, 497–522.  
682 <https://doi.org/10.1016/j.envpol.2007.09.015>
- 683 Li, K., Ramakrishna, W., 2011. Effect of multiple metal resistant bacteria from contaminated lake sediments  
684 on metal accumulation and plant growth. *J. Hazard. Mater.* 189, 531–539.  
685 <https://doi.org/10.1016/j.jhazmat.2011.02.075>
- 686 Liao, M., Xie, X.M., 2007. Effect of heavy metals on substrate utilization pattern, biomass, and activity of  
687 microbial communities in a reclaimed mining wasteland of red soil area. *Ecotoxicol. Environ. Saf.*  
688 66, 217–223. <https://doi.org/10.1016/j.ecoenv.2005.12.013>



- 689 Lima de Silva, A.A., de Carvalho, M.A.R., de Souza, S.A.L., Dias, P.M.T., da Silva Filho, R.G., de  
690 Meirelles Saramago, C.S., de Melo Bento, C.A., Hofer, E., 2012. Heavy metal tolerance (Cr, Ag  
691 AND Hg) in bacteria isolated from sewage. *Braz. J. Microbiol.* 43, 1620–1631.  
692 <https://doi.org/10.1590/S1517-838220120004000047>
- 693 Lopes, L.D., Pereira e Silva, M. de C., Andreote, F.D., 2016. Bacterial Abilities and Adaptation Toward the  
694 Rhizosphere Colonization. *Front. Microbiol.* 7. <https://doi.org/10.3389/fmicb.2016.01341>
- 695 Luo, S., Chen, L., Chen, J., Xiao, X., Xu, T., Wan, Y., Rao, C., Liu, C., Liu, Y., Lai, C., Zeng, G., 2011.  
696 Analysis and characterization of cultivable heavy metal-resistant bacterial endophytes isolated from  
697 Cd-hyperaccumulator *Solanum nigrum* L. and their potential use for phytoremediation.  
698 *Chemosphere* 85, 1130–1138. <https://doi.org/10.1016/j.chemosphere.2011.07.053>
- 699 Ma, Y., Prasad, M.N.V., Rajkumar, M., Freitas, H., 2011. Plant growth promoting rhizobacteria and  
700 endophytes accelerate phytoremediation of metalliferous soils. *Biotechnol. Adv.* 29, 248–258.  
701 <https://doi.org/10.1016/j.biotechadv.2010.12.001>
- 702 Ma, Y., Rajkumar, M., Freitas, H., 2009a. Isolation and characterization of Ni mobilizing PGPB from  
703 serpentine soils and their potential in promoting plant growth and Ni accumulation by *Brassica* spp.  
704 *Chemosphere* 75, 719–725. <https://doi.org/10.1016/j.chemosphere.2009.01.056>
- 705 Ma, Y., Rajkumar, M., Freitas, H., 2009b. Improvement of plant growth and nickel uptake by nickel  
706 resistant-plant-growth promoting bacteria. *J. Hazard. Mater.* 166, 1154–1161.  
707 <https://doi.org/10.1016/j.jhazmat.2008.12.018>
- 708 Ma, Ying, Rajkumar, M., Luo, Y., Freitas, H., 2011. Inoculation of endophytic bacteria on host and non-host  
709 plants—Effects on plant growth and Ni uptake. *J. Hazard. Mater.* 195, 230–237.  
710 <https://doi.org/10.1016/j.jhazmat.2011.08.034>
- 711 Marsili, S., Roccotiello, E., Carbone, C., Marescotti, P., Cornara, L., Mariotti, M.G., 2009. Plant  
712 Colonization on a Contaminated Serpentine Site. *Northeast. Nat.* 16, 297–308.  
713 <https://doi.org/10.1656/045.016.0522>
- 714 Massaccesi, G., Romero, M.C., Cazau, M.C., Bucsinszky, A.M., 2002. Cadmium removal capacities of  
715 filamentous soil fungi isolated from industrially polluted sediments, in La Plata (Argentina). *World*  
716 *J. Microbiol. Biotechnol.* 18, 817–820. <https://doi.org/10.1023/A:1021282718440>

- 717 Mengoni, A., Barzanti, R., Gonnelli, C., Gabbrielli, R., Bazzicalupo, M., 2001. Characterization of nickel-  
718 resistant bacteria isolated from serpentine soil. *Environ. Microbiol.* 3, 691–698.  
719 <https://doi.org/10.1046/j.1462-2920.2001.00243.x>
- 720 Morgan, J. a. W., Bending, G.D., White, P.J., 2005. Biological costs and benefits to plant–microbe  
721 interactions in the rhizosphere. *J. Exp. Bot.* 56, 1729–1739. <https://doi.org/10.1093/jxb/eri205>
- 722 Nkrumah, P.N., Echevarria, G., Erskine, P.D., van der Ent, A., 2018. Nickel hyperaccumulation in  
723 *Antidesma montis-silam*: from herbarium discovery to collection in the native habitat. *Ecol. Res.* 1–  
724 11.
- 725 Pal, A., Ghosh, S., Paul, A.K., 2006. Biosorption of cobalt by fungi from serpentine soil of Andaman.  
726 *Bioresour. Technol.* 97, 1253–1258. <https://doi.org/10.1016/j.biortech.2005.01.043>
- 727 Pal, A., Wauters, G., Paul, A.K., 2007. Nickel tolerance and accumulation by bacteria from rhizosphere of  
728 nickel hyperaccumulators in serpentine soil ecosystem of Andaman, India. *Plant Soil* 293, 37–48.  
729 <https://doi.org/10.1007/s11104-007-9195-7>
- 730 Penrose, D.M., Glick, B.R., n.d. Methods for isolating and characterizing ACC deaminase-containing plant  
731 growth-promoting rhizobacteria. *Physiol. Plant.* 118, 10–15. <https://doi.org/10.1034/j.1399-3054.2003.00086.x>
- 732
- 733 Pham, V.H.T., Kim, J., 2012. Cultivation of unculturable soil bacteria. *Trends Biotechnol.* 30, 475–484.  
734 <https://doi.org/10.1016/j.tibtech.2012.05.007>
- 735 Pollard, A.J., Reeves, R.D., Baker, A.J.M., 2014. Facultative hyperaccumulation of heavy metals and  
736 metalloids. *Plant Sci.* 217–218, 8–17. <https://doi.org/10.1016/j.plantsci.2013.11.011>
- 737 Rajkumar, M., Freitas, H., 2008a. Effects of inoculation of plant-growth promoting bacteria on Ni uptake by  
738 Indian mustard. *Bioresour. Technol.* 99, 3491–3498. <https://doi.org/10.1016/j.biortech.2007.07.046>
- 739 Rajkumar, M., Freitas, H., 2008b. Influence of metal resistant-plant growth-promoting bacteria on the  
740 growth of *Ricinus communis* in soil contaminated with heavy metals. *Chemosphere* 71, 834–842.  
741 <https://doi.org/10.1016/j.chemosphere.2007.11.038>
- 742 Rajkumar, M., Ma, Y., Freitas, H., 2013. Improvement of Ni phytostabilization by inoculation of Ni resistant  
743 *Bacillus megaterium* SR28C. *J. Environ. Manage.* 128, 973–980.  
744 <https://doi.org/10.1016/j.jenvman.2013.07.001>

- 745 Rascio, N., Navari-Izzo, F., 2011. Heavy metal hyperaccumulating plants: How and why do they do it? And  
746 what makes them so interesting? *Plant Sci.* 180, 169–181.  
747 <https://doi.org/10.1016/j.plantsci.2010.08.016>
- 748 Reeves, R.D., 2006. Hyperaccumulation of trace elements by plants, in: Morel, J.-L., Echevarria, G.,  
749 Goncharova, N. (Eds.), *Phytoremediation of Metal-Contaminated Soils*, NATO Science Series.  
750 Springer Netherlands, pp. 25–52.
- 751 Reeves, R.D., 1992. The hyperaccumulation of nickel by serpentine plants. *Veg. Ultramafic Serpentine Soils*  
752 253–277.
- 753 Reeves, R.D., Baker, A.J., 2000. Metal accumulating plants. *Phytoremediation Toxic Met. Using Plants*  
754 *Clean Environ. Raskin Ensley BD Eds John Wiley Sons Inc N. Y.*
- 755 Roccotiello, E., Marescotti, P., Di Piazza, S., Cecchi, G., Mariotti, M.G., Zotti, M., 2015. Biodiversity in  
756 Metal-Contaminated Sites—Problem and Perspective—A Case Study. *significance* 16, 26–28.
- 757 Roccotiello, E., Serrano, H.C., Mariotti, M.G., Branquinho, C., 2016. The impact of Ni on the physiology of  
758 a Mediterranean Ni-hyperaccumulating plant. *Environ. Sci. Pollut. Res.* 23, 12414–12422.  
759 <https://doi.org/10.1007/s11356-016-6461-3>
- 760 Roccotiello, Enrica, Serrano, H.C., Mariotti, M.G., Branquinho, C., 2015. Nickel phytoremediation potential  
761 of the Mediterranean *Alyssoides utriculata* (L.) Medik. *Chemosphere* 119, 1372–1378.  
762 <https://doi.org/10.1016/j.chemosphere.2014.02.031>
- 763 Roccotiello, E., Zotti, M., Mesiti, S., Marescotti, P., Carbone, C., Cornara, L., Mariotti, M.G., 2010.  
764 Biodiversity in metal-polluted soils. *Fresenius Environ. Bull.* 19, 2420–2425.
- 765 Rovira, A.D., 1965. Interactions Between Plant Roots and Soil Microorganisms. *Annu. Rev. Microbiol.* 19,  
766 241–266. <https://doi.org/10.1146/annurev.mi.19.100165.001325>
- 767 Rue, M., Vallance, J., Echevarria, G., Rey, P., Benizri, E., 2015. Phytoextraction of nickel and rhizosphere  
768 microbial communities under mono- or multispecies hyperaccumulator plant cover in a serpentine  
769 soil. *Aust. J. Bot.* 63, 92–102. <https://doi.org/10.1071/BT14249>
- 770 Schlegel, H.G., Cosson, J.-P., Baker, A.J.M., 1991. Nickel-hyperaccumulating Plants Provide a Niche for  
771 Nickel-resistant Bacteria. *Bot. Acta* 104, 18–25. <https://doi.org/10.1111/j.1438-8677.1991.tb00189.x>

- 772 Schwieger, F., Tebbe, C.C., 1998. A New Approach To Utilize PCR–Single-Strand-Conformation  
773 Polymorphism for 16S rRNA Gene-Based Microbial Community Analysis. *Appl. Environ.*  
774 *Microbiol.* 64, 4870–4876.
- 775 Schwyn, B., Neilands, J.B., 1987. Universal chemical assay for the detection and determination of  
776 siderophores. *Anal. Biochem.* 160, 47–56. [https://doi.org/10.1016/0003-2697\(87\)90612-9](https://doi.org/10.1016/0003-2697(87)90612-9)
- 777 Segura, A., Rodríguez-Conde, S., Ramos, C., Ramos, J.L., 2009. Bacterial responses and interactions with  
778 plants during rhizoremediation. *Microb. Biotechnol.* 2, 452–464. <https://doi.org/10.1111/j.1751-7915.2009.00113.x>
- 780 Sessitsch, A., Kuffner, M., Kidd, P., Vangronsveld, J., Wenzel, W.W., Fallmann, K., Puschenreiter, M.,  
781 2013. The role of plant-associated bacteria in the mobilization and phytoextraction of trace elements  
782 in contaminated soils. *Soil Biol. Biochem.* 60, 182–194.  
783 <https://doi.org/10.1016/j.soilbio.2013.01.012>
- 784 Shagol, C.C., Krishnamoorthy, R., Kim, K., Sundaram, S., Sa, T., 2014. Arsenic-tolerant plant-growth-  
785 promoting bacteria isolated from arsenic-polluted soils in South Korea. *Environ. Sci. Pollut. Res.* 21,  
786 9356–9365. <https://doi.org/10.1007/s11356-014-2852-5>
- 787 Sheng, X., He, L., Wang, Q., Ye, H., Jiang, C., 2008. Effects of inoculation of biosurfactant-producing  
788 *Bacillus* sp. J119 on plant growth and cadmium uptake in a cadmium-amended soil. *J. Hazard.*  
789 *Mater.* 155, 17–22. <https://doi.org/10.1016/j.jhazmat.2007.10.107>
- 790 Smalla, K., Wieland, G., Buchner, A., Zock, A., Parzy, J., Kaiser, S., Roskot, N., Heuer, H., Berg, G., 2001.  
791 Bulk and Rhizosphere Soil Bacterial Communities Studied by Denaturing Gradient Gel  
792 Electrophoresis: Plant-Dependent Enrichment and Seasonal Shifts Revealed. *Appl. Environ.*  
793 *Microbiol.* 67, 4742–4751. <https://doi.org/10.1128/AEM.67.10.4742-4751.2001>
- 794 Söderström, B.E., 1975. Vertical distribution of microfungi in a spruce forest soil in the south of Sweden.  
795 *Trans. Br. Mycol. Soc.* 65, 419–425.
- 796 Thijs, S., Langill, T., Vangronsveld, J., 2017. Chapter Two - The Bacterial and Fungal Microbiota of  
797 Hyperaccumulator Plants: Small Organisms, Large Influence, in: Cuypers, A., Vangronsveld, J.  
798 (Eds.), *Advances in Botanical Research, Phytoremediation*. Academic Press, pp. 43–86.  
799 <https://doi.org/10.1016/bs.abr.2016.12.003>

- 800 Tripathi, M., Munot, H.P., Shouche, Y., Meyer, J.M., Goel, R., 2005. Isolation and Functional  
801 Characterization of Siderophore-Producing Lead- and Cadmium-Resistant *Pseudomonas putida*  
802 KNP9. *Curr. Microbiol.* 50, 233–237. <https://doi.org/10.1007/s00284-004-4459-4>
- 803 Turgay, O.C., Görmez, A., Bilen, S., 2012. Isolation and characterization of metal resistant-tolerant  
804 rhizosphere bacteria from the serpentine soils in Turkey. *Environ. Monit. Assess.* 184, 515–526.  
805 <https://doi.org/10.1007/s10661-011-1984-z>
- 806 Urban, A., Puschenreiter, M., Strauss, J., Gorfer, M., 2008. Diversity and structure of ectomycorrhizal and  
807 co-associated fungal communities in a serpentine soil. *Mycorrhiza* 18, 339–354.  
808 <https://doi.org/10.1007/s00572-008-0189-y>
- 809 Vieira, F.C.S., Nahas, E., 2005. Comparison of microbial numbers in soils by using various culture media  
810 and temperatures. *Microbiol. Res.* 160, 197–202. <https://doi.org/10.1016/j.micres.2005.01.004>
- 811 Visioli, G., D'Egidio, S., Sanangelantoni, A.M., 2015. The bacterial rhizobiome of hyperaccumulators:  
812 future perspectives based on omics analysis and advanced microscopy. *Front. Plant Sci.* 5.  
813 <https://doi.org/10.3389/fpls.2014.00752>
- 814 Visioli, G., D'Egidio, S., Vamerali, T., Mattarozzi, M., Sanangelantoni, A.M., 2014. Culturable endophytic  
815 bacteria enhance Ni translocation in the hyperaccumulator *Noccaea caerulea*. *Chemosphere* 117,  
816 538–544. <https://doi.org/10.1016/j.chemosphere.2014.09.014>
- 817 Wenzel, W.W., Lombi, E., Adriano, D.C., 2004. Root and Rhizosphere Processes in Metal  
818 Hyperaccumulation and Phytoremediation Technology, in: *Heavy Metal Stress in Plants*. Springer,  
819 Berlin, Heidelberg, pp. 313–344. [https://doi.org/10.1007/978-3-662-07743-6\\_13](https://doi.org/10.1007/978-3-662-07743-6_13)
- 820 White, T.J., Bruns, T., Lee, S., Taylor, J.L., 1990. Amplification and direct sequencing of fungal ribosomal  
821 RNA genes for phylogenetics. *PCR Protoc. Guide Methods Appl.* 18, 315–322.
- 822 Xue, P.-P., Carrillo, Y., Pino, V., Minasy, B., McBratney, A.B., 2018. Soil Properties Drive Microbial  
823 Community Structure in a Large Scale Transect in South Eastern Australia. *Sci. Rep.* 8, 11725.  
824 <https://doi.org/10.1038/s41598-018-30005-8>
- 825 Zaidi, S., Usmani, S., Singh, B.R., Musarrat, J., 2006. Significance of *Bacillus subtilis* strain SJ-101 as a  
826 bioinoculant for concurrent plant growth promotion and nickel accumulation in *Brassica juncea*.  
827 *Chemosphere* 64, 991–997. <https://doi.org/10.1016/j.chemosphere.2005.12.057>

- 828 Zhang, Y., Lu, L., Chang, X., Jiang, F., Gao, X.-D., Peng, F., 2018. Small-scale soil microbial community  
829 heterogeneity linked to landforms on King George Island, maritime Antarctica. bioRxiv 310490.  
830 <https://doi.org/10.1101/310490>
- 831 Zhu, L.-J., Guan, D.-X., Luo, J., Rathinasabapathi, B., Ma, L.Q., 2014. Characterization of arsenic-resistant  
832 endophytic bacteria from hyperaccumulators *Pteris vittata* and *Pteris multifida*. Chemosphere 113,  
833 9–16. <https://doi.org/10.1016/j.chemosphere.2014.03.081>
- 834 Zhuang, X., Chen, J., Shim, H., Bai, Z., 2007. New advances in plant growth-promoting rhizobacteria for  
835 bioremediation. Environ. Int. 33, 406–413. <https://doi.org/10.1016/j.envint.2006.12.005>
- 836 Zotti, M., Di Piazza, S., Roccotiello, E., Lucchetti, G., Mariotti, M.G., Marescotti, P., 2014. Microfungi in  
837 highly copper-contaminated soils from an abandoned Fe–Cu sulphide mine: Growth responses,  
838 tolerance and bioaccumulation. Chemosphere 117, 471–476.  
839 <https://doi.org/10.1016/j.chemosphere.2014.08.057>
- 840

- Only 10% of hyperaccumulators have their rhizosphere examined
- *A. utriculata* is a facultative Ni-hyperaccumulator that thrives in serpentine soils
- Rhizobiota of *A. utriculata* seems to be limited by serpentine soil conditions
- *Pseudomonas*, *Streptomyces*, *P. ochrochloron* and *T. harzianum* group hypertolerate Ni
- Rhizobiota with PGP traits and high Ni tolerance can improve plant phytoextraction