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Corals in high diversity reefs resist human impact

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ABSTRACT

Coral reefs are amongst the most diverse ecosystems in the biosphere. However, they also represent one of the most threatened marine systems. Apart from global change, especially fishing and tourism affect coral reefs either with mechanical damage or with increase of pollution and sedimentation. Recently, the increase of disturbances has induced extensive changes in community structure and composition of coral reefs. Well-balanced and rich communities can better resist disturbances and show a more rapid recovery, compared to less biodiverse systems. This study assesses the status of coral reefs subjected to several anthropogenic pressures, using a modified version of Coral Condition Index (CCI) that takes into account all *Acropora* and *Pocillopora* growth forms and considers a further category of coral damage: the presence of disease. The investigations were carried out at Bangka Island (North Sulawesi, Indonesia), where some of the most flourishing reefs in the country are present. The CCI takes into account the extent of different damages on coral colonies, particularly of the genera *Acropora* and *Pocillopora*, being among the most widespread bioconstructors of local coral reefs and very sensitive to anthropogenic disturbances. The aim of the present work is to test whether the CCI is a reliable index in different coral reefs, and to evaluate if highly biodiverse reefs show a better resistance to several human stressors. Data showed high values of CCI (0.9 on average) at all the investigated sites and at the two depths (3 and 9 m) for each site, with the most abundant category represented by "healthy coral colonies". These data indicate a reasonably good health status of the reef in the study area (CCI > 0.8). The presence of different types of human pressure in the study area was evaluated through the use of metric proxies. Results do not seem to show any significant influence of such human activities on reef coral status, as shown by the low values of correlation between CCI values and the distances of the study sites from the three main sources of stress (Villages, Resorts and Other). Moreover, the present data seem to confirm that highly biodiverse and well-structured assemblages can resist disturbances more efficiently and that human pressure in the study area is sustainable. Compared to fishing activities, the impact of Scuba diving on coral reef is lower, resulting more sustainable and ecologically non-destructive. CCI summarizes many kinds of information and can be applicable in various areas with different pressures. It is a useful tool that might help to assist and guide management decisions towards alternative development models.

Introduction Coral reefs represent one of the most threatened marine ecosystems, due to either natural or anthropogenic disturbances, on both global and local scale (Lasagna et al., 2014). Global warming is resulting in widespread bleaching and mass mortality events (Corresponding author. E-mail address: federica.ferrigno@uniparthenope.it (F. Ferrigno). (Baker et al., 2008; Morri et al., 2015); similarly, ocean acidification is hampering coral growth and survival (Kleypas and Yates, 2009). On local scale, especially fishing and tourism affect coral reefs either with mechanical damage or with pollution and sedimentation (Bryant et al., 1998). The recent increase in scale and frequency of disturbances has resulted in extensive changes in community structure and composition of coral reefs (Done, 1992; Knowlton, 2001; Montefalcone et al., 2011), which dramatically decrease their recovery potential (Dudgeon et al., 2010). Some studies have also suggested that anthropogenic impacts can prevent

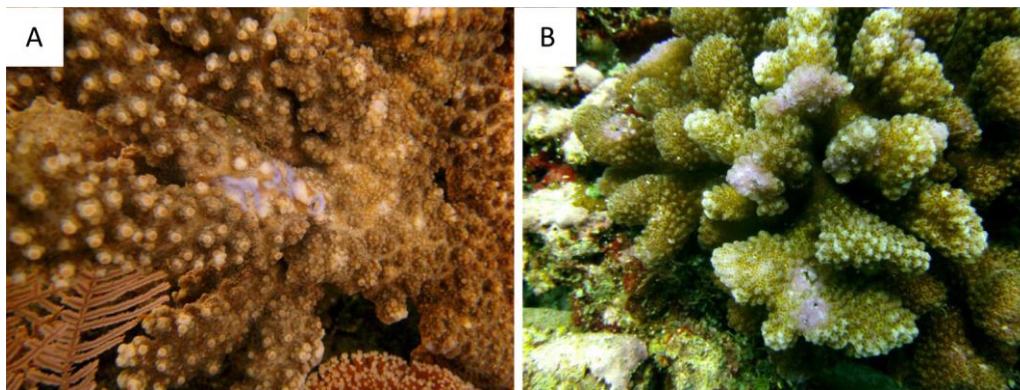


Fig. 2. Examples of diseased Acropora (A) and Pocillopora (B) colonies (DIC)

Recovery of coral assemblages from natural disasters (Hughes and Connell, 1999), and it is usually difficult to discern the different stressors increasing coral reef damage, since they act synergically (Nyström et al., 2000). Reducing local stressors could mitigate the impacts of global stressors, such as climate change. On the other hand, it has been said that continued degradation caused by local stressors induces coral communities to become dominated by tolerant species, making them more resilient to climate disturbance (Côté and Darling, 2010). Nonetheless, there is growing evidence that, following a range of disturbances, protected or less degraded reefs return more quickly to their original state than unprotected or more degraded reefs (Mumby and Harborne, 2010). Well-balanced communities can not only sustain their own resilience but also contribute to the resilience of other “downstream” communities (West and Salm, 2003; McClanahan et al., 2012). Coral reefs seem to shift their general structure in relation to physical disturbances leading to a loss of three-dimensional structural complexity, negatively affecting ecosystem functioning (Fava et al., 2009). For example, it was observed that loss of fast-growing branching corals of the families Acroporidae and Pocilloporidae induces a decrease of resilience of coral reef; particularly, highly diverse assemblages of Acropora exhibit rapid recovery after disturbance and prevent the shift of the ecosystem into a more vulnerable situation (Roff and Mumby, 2012). Other studies demonstrated that while extreme waves are needed to inflict damage on robust coral communities, much lesser forces can decimate fragile coral communities (Madin and Connolly, 2006). The threshold in stress intensity capable of causing severe loss in coral cover on undisturbed sites is approximately half that reported to cause coral loss in disturbed sites (Fabricius et al., 2008); particularly, in sites affected by a multitude of disturbances, resulting in low coral cover and simple framework structures, vulnerability of reefs increases (Gardner et al., 2003). The present paper evaluates the health status of corals thriving in reefs subjected to several human activities, using a modified version of the Coral Condition Index (CCI), originally developed by Lasagna et al. (2014) for Maldivian reefs. The study was carried out in Indonesia, a country with a recent and substantial growth and consequent urban development, but with also a high level of marine biodiversity, being part of the so-called Coral Triangle (Tomascik et al., 1997). The CCI, an extension of the Coral Damage Index (CDI) of Jameson et al. (1999), takes into account the extent of different damages on coral colonies, particularly of the genera Acropora and Pocillopora, because they are among the main and widespread bioconstructors of Indo-Pacific coral reefs and are sensitive to anthropogenic disturbances (Penin et al., 2007; Lasagna et al., 2010; Bigot and Amir, 2012). Indonesia is known to host 350 scleractinian coral species (Best et al., 1989), that is nearly the double of the 180 scleractinian coral species inventoried for the Maldives (Pichon and Benzoni, 2007). The present work aims at using the CCI in order to explore the

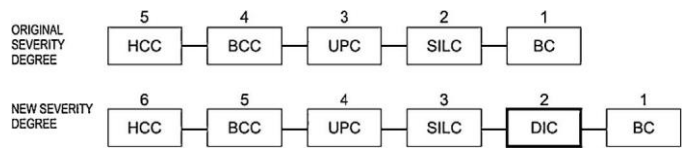


Fig. 3. Original (above) and new damage severity degree (below) for each CCI category. For explanation of abbreviations see text.

idea that corals in high diversity reefs resist several human stressors better than in comparatively low diversity reefs.

2. Materials and methods

2.1. Study area

Indonesia is part of the Coral Triangle, which covers 5.7 million square kilometers of ocean waters, has some of the world's largest coral reefs and is recognized as the global center of marine biodiversity (Tomascik et al., 1997). Unfortunately, Indonesia also hosts a very high number of species threatened with extinction (Ross and Wall, 1999). The big island of Sulawesi gathers a mixture of Australian and Asian species with also many endemic species. Particularly, North Sulawesi holds the greatest level of endemism, due to its long geological history of mountain and water barriers, and island bridges from north east and west (Whitten et al., 2002). The climate is greatly influenced by the system of monsoon winds that strongly influence the movement of water and productivity of coastal and marine systems. The general patterns of circulation of the seas are locally led by strong tidal regimes. Moreover, high seismic and volcanic activity plays a fundamental role in modeling land morphology (Tomascik et al., 1997). North Sulawesi is one of the six Sulawesi's provinces, with about 2,400,000 inhabitants (Badan Pusat Statistik, <http://www.bps.go.id/>). Its economy is mainly based on small-scale agriculture, particularly clove to produce cigarettes; other important and growing activities are coastal fishing farms, copper and gold mining, and tourism (Indonesia, <http://indahnesia.com/indonesia/SULECO/economy.php>). Bangka Island owns rich and diverse reefs, which are not too popular yet to Scuba diving. Thus, it may be an ideal site to test the validity of CCI. However, in the last few years, this island has been undergoing a continuous increase of anthropic development and its monitoring could help to understand the effects of human activities on the environment and, possibly, to protect it.

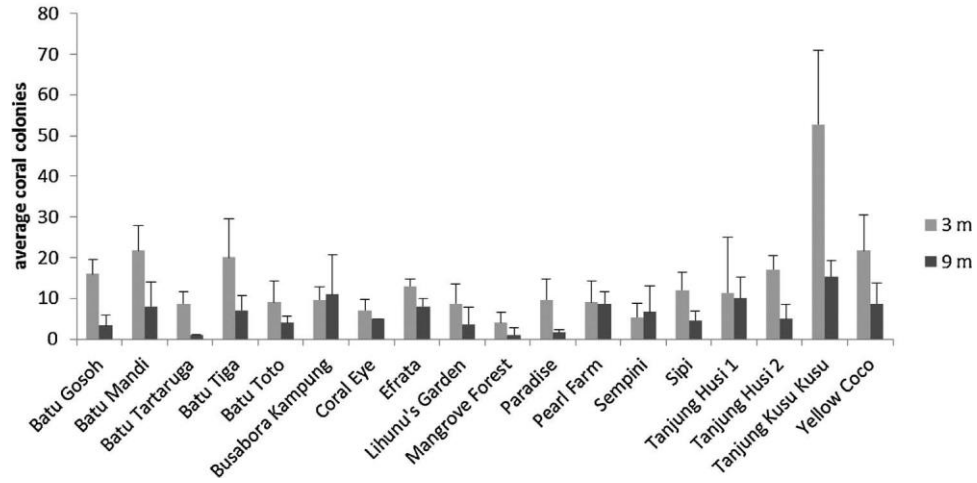
2.2. Field activities

Surveys were carried out in October and November 2014, at Bangka Island (North Sulawesi, Indonesia) (Fig. 1), by Scuba diving. Data were collected according to the methods for hard bottom described by Bianchi et al. (2004), at two different depths, corresponding to reef flat (3 m) and reef slope (9 m), along three replicates 20 × 2 m belt transects, parallel to the coastline. The investigation included 18 georeferenced sites with different types of potential human pressure (Table 1), randomly selected around the island, for a total of 108 transects. Since collecting quantitative data is often difficult, time-consuming and expensive, the use of proxies must be considered, whenever applicable. In this case, distance in kilometers from the major stress sources was used as a proxy for human pressure within the study sites (Hawkins et al., 1999; Parravicini et al., 2012). The three major stress distance proxies chosen were: Villages (fishing), Resorts (Scuba diving) and Other (see below). Villages (fishing): fishing villages can cause breaking and upturning of coral colonies with fishing gears and anchors, or smothering, diseases and bleaching due to release of pollutants and sediments in the water. Resorts (Scuba diving): the resorts, with their diving centers, could stress the reef through tourism activities of snorkeling and diving that can mainly cause mechanical damage as breaking and upturning of coral colonies. Other: activities related to metal mining cause smothering, diseases and bleaching due to release of pollutants and sediments in the water, whereas fixed fishing structures and a pearl farm may cause breaking and upturning of coral colonies. Only Acropora and Pocillopora colonies with diameters higher than 15 cm were counted in each transect, because they can be easily identified and are potentially sexually mature (Babcock et al., 2003; Lasagna et al., 2010). In the modified version of CCI, suggested here, all the morphologies of Acropora were considered, while the original CCI accounted only for tabular Acropora colonies.

2.3. Data management

According to Lasagna et al. (2014), Acropora and Pocillopora colonies were classified into the following categories: healthy coral colonies (HCC), broken coral colonies (BCC), upturned coral colonies (UPC), smothered coral colonies (SILC), bleached coral colonies (BC) and recently dead corals (RDC). An important change of this new version of the index is the addition of another category, for a more complete overview of all possible damages of corals: diseased

Healthy coral colonies (HCC)



Damaged coral colonies (DCC)

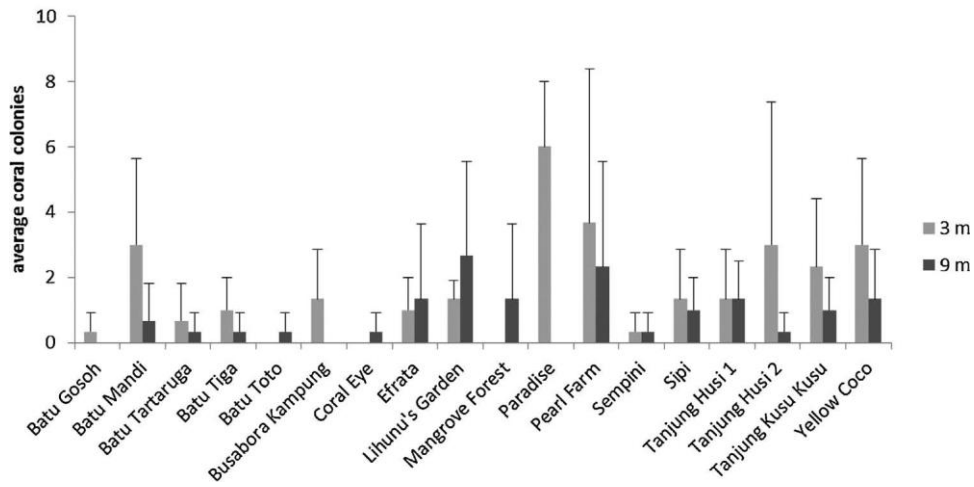


Fig. 4. Average number (\pm SD) of healthy coral colonies (HCC), at each site.

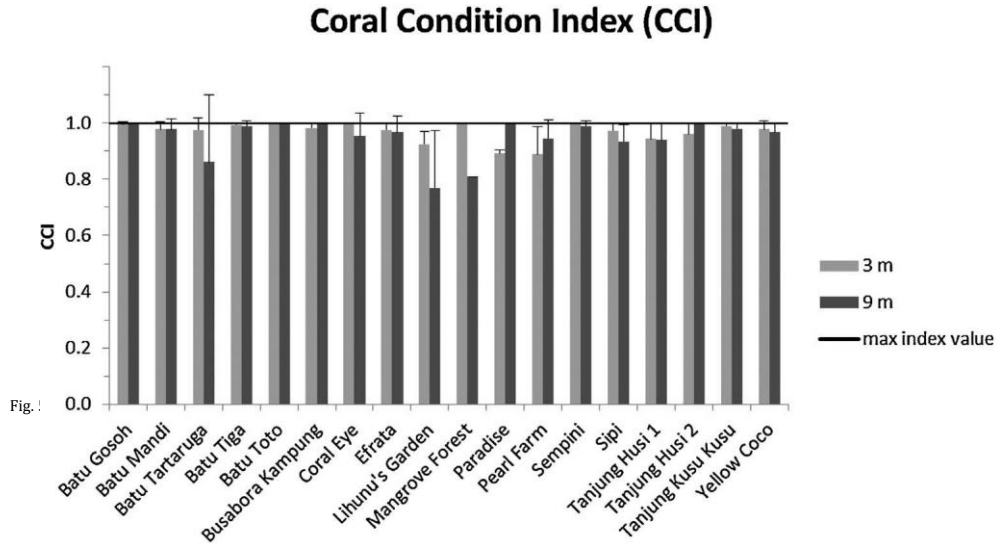


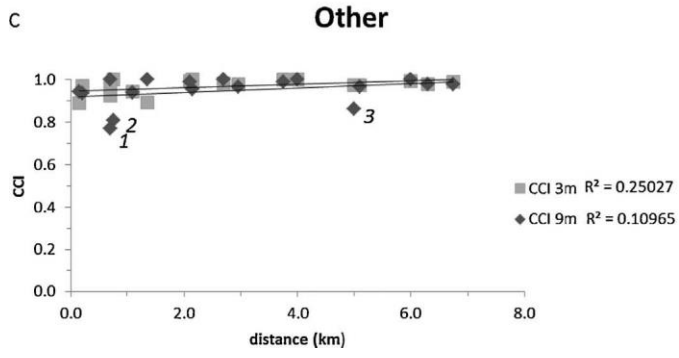
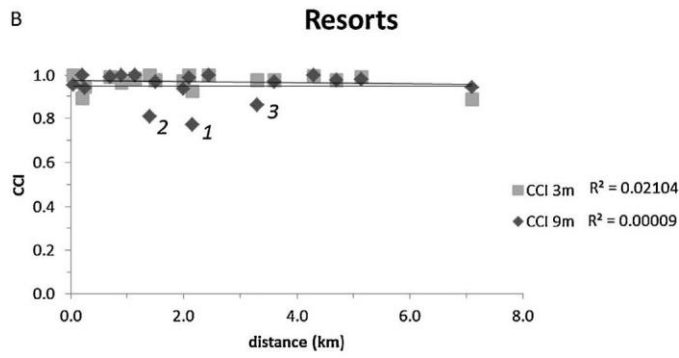
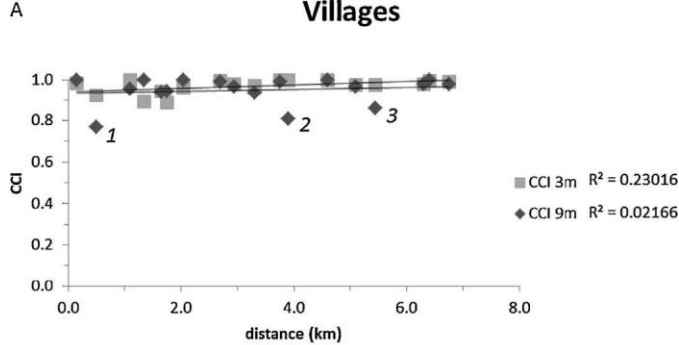
Fig. 6

Fig. 6. Average values (\pm SD) of Coral Condition Index (CCI), at each site. coral colonies (DIC), including those colonies that exhibited evident abnormal pigmentation (Fig. 2).

Coral diseases could induce tissue degradation and skeleton exposure, potentially resulting in colony death (Richardson, 1998; Weil, 2004; Kumaraguru et al., 2005). The modified CCI was computed according to the new severity degree scale suggested (Fig. 3): where TCC is the total number of coral colonies. The sum of all damaged (BCC, UPC, SILC, DIC and BC) and recently dead coral colonies (RDC) is the category damaged coral colonies (DCC). Categories were assigned a different degree of severity (1–6), based on literature information about the putative recovery capacity following damage (Wittenberg and Hunte, 1992; Anthony et al., 2009; Maina et al., 2011; Anthony et al., 2009; Maina et al., 2011), where lower degree corresponds to higher damage (Fig. 3). CCI results can be represented by values ranging from 0 to 1, where lower values suggest the impact of large scale disturbances, and comparatively higher values suggest the impact of local scale disturbances; a value of 1 indicates a state of optimal health of reef. Finally, linear regressions between the distance from the sources of stress and CCI values were calculated.

3. Results

On a total of 1241 colonies of *Acropora* and *Pocillopora*, the majority (69%) was found at 3 m depth. At both depths, most coral colonies were in healthy condition (90% at 3 m and 88% at 9 m depth) and the main damage was UPC (5% at 3 m and 7% at 9 m), followed by BCC (2% at 3 m and SILC (2%) at 9 m (Table 2). Only 0.2% of coral colonies resulted diseased (DIC) at 3 m depth, while no diseased colony was found at 9 m depth. At 3 m depth, coral colonies in healthy conditions (HCC) were on average 14.24 ± 4.26 , with the lowest value in Mangrove Forest (4.00 ± 2.65) and the highest in Tanjung Kusu Kusu (52.67 ± 18.18); at 9 m depth, coral colonies were on average 6.26 ± 2.46 , with the lowest value in Batu Tartaruga (1.00 ± 0.00) and Mangrove Forest (1.00 ± 1.73), and the highest in Tanjung Kusu Kusu (15.33 ± 4.04) (Fig. 4). At 3 m depth, damaged coral colonies (DCC) were on average 1.65 ± 1.36 , with the lowest value in Batu Toto, where no damaged coral colony was found, and the highest value in Paradise (6.00 ± 2.00); at 9 m depth, damaged coral colonies were on average 0.83 ± 0.97 , with the lowest value in Batu Gosoh, Busabora Kampung and Paradise, where no damaged coral colony was found, and the highest value in Lihunu's Garden (2.67 ± 2.89) (Fig. 5). The calculation of CCI was then performed on the 18 sites at each depth (3 and 9 m) (Fig. 6). The values ranged from a minimum of 0.77 to a maximum of 1.00, which corresponds to a good health status of the coral reef. Particularly, at 3 m depth, the average value was 0.97 ± 0.03 , with the lowest value in Paradise (0.89 ± 0.01) and Pearl Farm (0.89 ± 0.10), and the highest value (1.00 ± 0.00) in Batu Toto, Coral Eye, Mangrove Forest and Sempini; at 9 m depth, the average value was 0.95 ± 0.07 , with the lowest value in Lihunu's Garden (0.77 ± 0.20), Mangrove Forest (0.81 ± 0.01) and Batu Tartaruga (0.86 ± 0.24), and the highest value (1.00 ± 0.01) in Batu



Compared to fishing activities, Scubadiving has a lower impact on coral reefs and is less harmful to coral health, being therefore more sustainable (Walters and Samways, 2001). Nevertheless, unsustainable rates of attendance and inappropriate behavior of diving tourists may cause changes in the marine environment, mainly through mechanical damage (Medio et al., 1997; Barker and Roberts, 2004). This study showed that this activity is underexploited in the study area and can be considered sustainable. Indeed, the number of dives per site per year, according to the data provided by the local Scuba diving centers, does not exceed the value of 1000. This figure is lower than the 6000 dives per site and per year, estimated by Hawkins and Roberts (1997) as impacting on coral reefs. The threshold of 6000 dives per site and per year has been established for the Caribbean Sea and the Red Sea (Tratalos and Austin, 2001; Zakai and Chadwick-Furman, 2002), two coral reef areas characterized by lower biodiversity than the Coral Triangle (Tomascik et al., 1997). The number of Scuba dives in the area might therefore be sensibly increased, provided effective monitoring programs are carried out. Monitoring the effects of Scuba diving on the reef is necessary to protect the ecosystem and also to direct the entire socio-economic sector towards sustainable management approaches. Additionally, small-scale management strategy aimed at the protection of territory and maintenance of ecosystems, could result also in a larger scale protection from global disturbances. The measure beyond which community begins to change as a response to stress, is hard to assess but it is necessary to know for the conservation of high biodiversity areas. Estimate the tolerance threshold of ecosystem to each disturbance is mandatory. For example, a way to estimate the sustainable level of diving tourism attendance by a marine ecosystem is to calculate its diving "carrying capacity" (Davis and Tisdell, 1995). However, in order to fully understand the ecosystem functioning of a reef, it is necessary to consider the influence of multiple stressors acting together, because this might increase coral reef damage (Nyström et al., 2000). Studying single events in isolation can be misleading, therefore a longer term approach is necessary to understand the responses of reef corals to multiple stressors (Hughes and Connell, 1999). Co-occurrence of different damage types could produce an increase of instability and fragility of the entire reef framework (Bellwood et al., 2004).

5. Conclusions A modified version of Coral Condition Index was applied to assess the ecological status of coral colonies in a high diversity coral reef, taking into account also the effects of different stressors. The modified index gives additional information on reef coral health than the original version, by adding an additional coral damage (disease), and also with respect to other measures based only on live coral cover. Diseases are in fact known to be an early signal of coral stress (Harvell et al., 1999). In this study several possible damages on the corals were considered, obtaining a more complete overview of the effects of human pressure on rich and sensitive communities. CCI is a very fast and easy method; it supplies valuable information on various types of damage at different severity degrees, and, finally, does not require highly specialized operators. The index, previously tested in some reefs of Maldives (Lasagna et al., 2014) and in Indonesia (present study), implies the use of two of the main and widespread genera of hard coral of the Indo-Pacific and sensitive to anthropogenic disturbances (Penin et al., 2007; Lasagna et al., 2010; Bigot and Amir, 2012). The positive outcome shows that it might work also in other reefs where the two genera are amongst the dominant. Different kinds of human pressures were considered in the study. Particularly, distances of stress sources from study sites were used as proxies and correlated to CCI values. The results

Fig. 7. Linear regression between the Coral Condition Index (CCI) and the distance of sites from the source of anthropogenic stress Villages (A), Resorts (B) and Others (C), at both depth of 3 m and 9 m. 1: Lihunu's Garden; 2: Mangrove Forest; 3: BatuTartaruga. Distances of the study sites from the three main sources of anthropogenic stress ranged from a minimum of 50 m to a maximum of 7 km. No correlation between CCI and distance from stress sources, at either 3 m or 9 m depth, was detected at almost all sites (Fig. 7). However, Lihunu's Garden, Mangrove Forest and Batu Tar-taruga, at 9 m depth, may be considered outliers showing lower CCI values. Coral damage at the first site consists in smothered coral colonies (SILC), upturned coral colonies (UPC) and bleached coral colonies (BC); at the second site the damage consists in upturned coral colonies (UPC), and at the third site it consists in bleached coral colonies (BC).

4. Discussion Compared to the original version (Lasagna et al., 2014), the modified CCI suggested here gives information on the status of individual coral colonies, and, according to the use of ecological indicators (Dale and Beyeler, 2001; Jameson et al., 2001), provides a tool for an even earlier detection of change. Unlike the most commonly used metrics of reef health, such as percent-age live coral cover (Clarke et al., 1993; McManus et al., 1997; Sweatman et al., 2011), CCI gives more detailed information, taking into account the effects of various stress and at different severity degree to understand better the real possible damage and how to act when a reef is endangered. Moreover, it is a very fast and easy method to be applied based only on field activities to collect data (directly by Scuba diving), and does not require extensive taxonomic knowledge, since the two coral genera used as indicators (Acropora and Pocillopora), are very common and easily recognizable. Furthermore, the two genera have been chosen because they are representative of Indo-Pacific hard coral, being among the main bioconstructors of coral reefs and being sensitive to human pressure (Penin et al., 2007; Lasagna et al., 2010; Bigot and Amir, 2012). In this study, a modified version of the index was applied to assess the ecological status of coral colonies in a high diversity coral reef and to test whether they resist several human stressors better than in less diverse reefs. In this new version of the CCI, as mentioned above, all Acropora and Pocillopora forms were considered, and the presence of disease was evaluated. CCI was modified according to the types of damage and re-assigning a suitable severity degree. Considering as many as possible coral damages might allow to obtain a more complete overview of human pressure causes and effects. So, the integration of the category "diseased coral colonies" (DIC) in the index calculation, provides important information and might be useful for a better understanding of the health status of reefs. Diseases are an early signal of coral stress; indeed, it was observed that both climate and human activities may induce physiological stress, compromising host resistance and increasing frequency of opportunistic diseases (Harvell et al., 1999). Anthropogenic pollutants, habitat degradation and overfishing have led to a recent increase in disease outbreaks in many reef ecosystems and organisms (Harvell et al., 2007). The several stressors affecting coral reefs, particularly along heavily urbanized coastlines, as well as introductions of new species to distant reefs by global transport, are contributing to concerns about extinction risks and loss of biodiversity (Peters, 2015). More anthropic areas may be threatened by acute stresses, including destructive fishing practices, as well as anchor damage and ship groundings, and chronic stresses, including sewage pollution, increased sedimentation, nearshore eutrophication, and industrial pollution (Edinger et al., 1998). In the present study, data showed high values of CCI (0.9 on average) in all the investigated sites and at the two depths for each site, with the most abundant category represented by healthy coral colonies (HCC). These data translate into a good health status of the reef corals in the study area (CCI > 0.8). The lowest CCI value was detected at Lihunu's Garden, a site stressed by the nearby village with its run-off that causes sedimentation and bleaching of coral colonies, and with anchoring and transition of boats that may induce upturning of coral colonies. The other lower CCI values were detected at 9 m depth at Mangrove Forest with upturned coral colonies probably due to fishing activities, and at Batu Tar-taruga with bleached coral colonies, even if few hard coral colonies were present while soft corals were dominant. Quantitative and accurate data for evaluation of human pressure are often difficult to obtain, due to the interaction of multiple factors; in this study, the distances of stress sources from study sites were used as proxies of human pressure. Proxy metrics are used to reduce time and money required for data collection (Richards, 2013) and are representative of trends in biodiversity and significant challenge for ecology and conservation issues (Baillie et al., 2008). There are also new methodologies, used for evaluation of the environmental impact, such as the Multiple Criteria Data Envelopment Analysis (MCDEA), a model to rank and select the best alternative considering both qualitative and quantitative criteria (Zhao et al., 2006). Other data-driven models, including Machine Learning (ML) techniques, have been employed to identify the ecologically significant variables, allowing to predict some ecological damages (Muttil and Chau, 2007), even though they might appear more time-consuming and less direct approaches in this context. Although different kinds of human pressures are present in the study area, they do not seem to exert a marked influence on reef coral health, as indicated by the low values of correlation between CCI and the distances of sites from the three main sources of stress (Villages, Resorts and Other). This could mean that time and/or intensity of disturbances are not enough to affect coral health (Done, 1992; Knowlton, 2001; Montefalcone et al., 2011); on the other hand, these results suggest corals in high diversity reefs are less prone to damage, as Indonesia lies at the center of the world's tropical biodiversity and has extremely rich and diverse coral reefs (Veron, 1993). Thus, the present work may provide an indication that corals in highly biodiverse and well-structured assemblages can resist disturbances more efficiently and have a more rapid recovery, compared to less biodiverse environments (Tomascik et al., 1997; McClanahan et al., 2012). CCI can be applied to coral reefs of different areas, to assess the possible impact that human activities may produce according to the socio-economic context. Comparing the present data from Indonesia with those from the Maldives by Lasagna et al. (2014), Indonesian corals result in a healthier status, which is consistent with a general lower population density of Indonesia, about 135 inhabitants per square kilometer, compared with the Maldives, about 1213 inhabitants per square kilometer (United Nations Population Division, <http://www.un.org/en/development/desa/population/theme/trends/>). Overall, the human pressure in the study area seems to be sustainable, even if the growing population is placing greater pressure on the services from the environment, e.g. fisheries, and contribute to increased pollution, damage to habitats and illegal practices (DeVantier et al., 2004). Besides fishing activities, tourism is becoming of increasing interest for local economy, particularly Scuba diving, one of the main touristic activities in tropical seas and it is a significant source of income for the most flourishing coral reef sites (Brown et al., 2001). Well-planned tourism provides economic and political incentives for management and conservation, and may bring additional benefits to local communities and regional economies (Agardy, 1993).