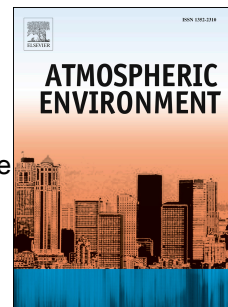


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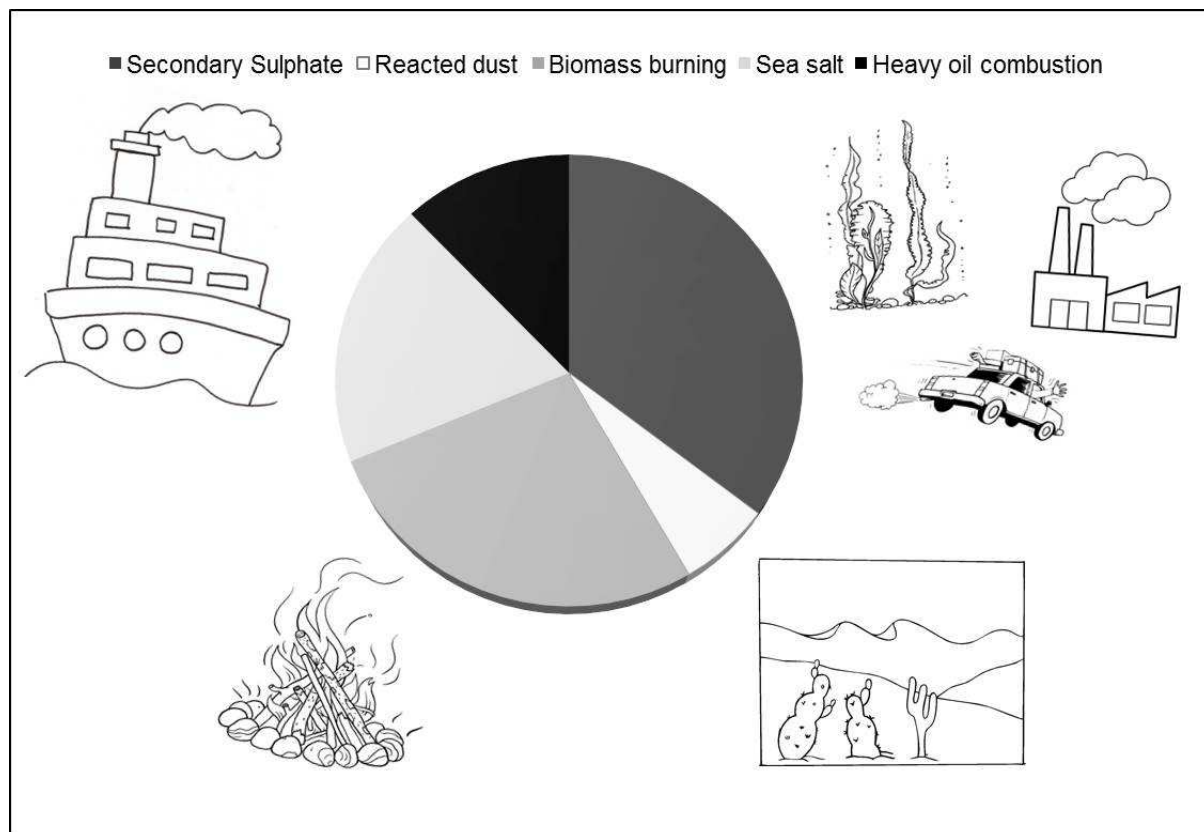
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1 **PM10 source apportionment applying PMF and chemical tracer analysis to ship-borne**
2 **measurements in the Western Mediterranean**

3
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15
16 **Abstract**

17
18 A PM10 sampling campaign was carried out on board the cruise ship Costa Concordia during
19 three weeks in summer 2011. The ship route was Civitavecchia-Savona-Barcelona-Palma de
20 Mallorca-Malta (Valletta)-Palermo-Civitavecchia. The PM10 composition was measured and
21 utilized to identify and characterize the main PM10 sources along the ship route through
22 receptor modelling, making use of the Positive Matrix Factorization (PMF) algorithm. A
23 particular attention was given to the emissions related to heavy fuel oil combustion by ships,
24 which is known to be also an important source of secondary sulphate aerosol. Five aerosol
25 sources were resolved by the PMF analysis. The primary contribution of ship emissions to
26 PM10 turned out to be $(12 \pm 4)\%$, while secondary ammonium sulphate contributed by $(35 \pm$
27 $5)\%$. Approximately, 60% of the total sulphate was identified as secondary aerosol while
28 about 20% was attributed to heavy oil combustion in ship engines. The measured
29 concentrations of methanesulphonic acid (MSA) indicated a relevant contribution to the
30 observed sulphate loading by biogenic sulphate, formed by the atmospheric oxidation of
31 dimethyl sulphide (DMS) emitted by marine phytoplankton.

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36
37 *Keywords:* PM10, ship emissions, Mediterranean Basin, source apportionment

39 1. Introduction

40

41 The contribution of diverse anthropogenic and natural emissions sources, such as highly
42 populated and industrial coastal areas, intense ship traffic, forest fire emissions and Sahara
43 dust outbreaks, together with meteorological and geographical peculiarities, make the
44 Mediterranean Basin one of the most polluted regions on Earth in terms of ozone
45 concentrations and aerosol loading (Lelieveld et al, 2002, Velchev et al, 2011). This is caused
46 by local emissions as well as transport of air pollution from outside the Mediterranean region.
47 Ship emissions are an important source of pollution in this region and represent significant
48 and growing contributors to air quality degradation in coastal areas (Van Aardenne et al.,
49 2013). Emissions of exhaust gases and particles from the oceangoing ships affect the chemical
50 composition of the atmosphere, climate and regional air quality (Eyring et al., 2005). In recent
51 years, particle emissions from ships and harbour activities became a concern for air quality
52 and object of several scientific investigations (Moreno et al., 2010, Becagli et al., 2012, Cesari
53 et al., 2014, Bove et al., 2014). A number of studies have shown that ship exhaust particles
54 contain V and Ni and these elements have been used as markers to investigate primary ship
55 emissions using receptor models (Mazzei et al., 2008, Viana et al., 2009; Cuccia et al., 2010,
56 Pandolfi et al., 2011, Salameh et al., 2015). The Joint Research Centre of the European
57 Commission (JRC, EC) has carried out an air quality monitoring program from 2006 to 2014,
58 based on observations from a cruise ship following a regular route in the Western
59 Mediterranean. In the framework of a collaboration agreement between the JRC and Costa
60 Crociere, continuous measurements of atmospheric pollutants were carried out on cruise ships
61 from spring to autumn. During two campaigns in particular, in 2009 and 2010, a two-stage
62 streaker sampler (Formenti et al., 1996) was installed on the ship. The elemental composition
63 of the fine and coarse fraction of PM₁₀, separately collected by the streaker on an hourly
64 basis, was measured by PIXE analysis (Schembari et al., 2014). These datasets were used for
65 an investigation of the influence of ship emissions on the composition of aerosols over the sea
66 through a source apportionment analysis by PMF as well as by chemical marker compounds.
67 The ship emissions were found to be an important source of aerosols in the Western
68 Mediterranean, however a quantification of their impacts by PMF was not obtained. That
69 experiment did not disentangle primary and secondary sources of sulphate and did not resolve
70 the contribution of primary aerosol from ships, presumably because of the insufficient
71 chemical speciation of PM₁₀. A mixed combustion source, which showed evidence of a direct

72 connection with ship emissions was found to contribute by 55%, 63% and 80% to PM10,
73 Black Carbon and sulphate, respectively (Schembari et al., 2014). In summary, the results of
74 the previous campaigns indicated a significant impact of ship emissions to PM levels in the
75 explored area but were not conclusive. In this context, a new PM10 sampling campaign was
76 organized in the summer of 2011, to complete the information of the previous studies and to
77 get a better description of PM sources. An extensive characterisation of PM10 samples,
78 collected using a sequential filter sampler, was addressed; the obtained data were analysed by
79 PMF and used to identify and characterize the main PM10 sources met along the route.

80

81 **2. Material and methods**

82

83 *2.1 Monitoring campaign*

84 The monitoring station was placed in a cabin at the front of the top deck of the ship
85 “Costa Concordia”. It permitted to perform continuous measurements of NO_x, SO₂, O₃ and
86 Black Carbon (BC), the last one by means of an Aethalometer (AE 21, 2 wavelengths, Magee
87 Scientific, USA) (Schembari et al., 2014). The aerosol sampling campaigns were carried out
88 during three weeks of summer 2011: July 18-25, August 15-22 and September 12-19. PM10
89 samples were collected on Quartz filters (47mm diameter, flow rate 2.3 m³ h⁻¹) using a Sven
90 Leckel Ingenieurburo sequential sampler, placed on the top of the cabin where the monitoring
91 and meteorological station were also located. The sampling was carried out on a variable time
92 basis: the sampler was started 1 h after the departure from each harbour and stopped 1 h
93 before the arrival in the next harbour. Each leg was then divided in periods of about 4-5 h with
94 one filter sampled per each period. This resulted in a variable number of filters per open-sea
95 leg and in a total number of about 20 filters per week.

96

97 *2.2 Analytical methods*

98 All filters were pre-conditioned for two days in a controlled room (temperature:
99 20±1°C, relative humidity: 50±5%) before and after the sampling and then weighed using an
100 analytical balance (sensitivity: 1 µg). Field blank filters were used to monitor possible
101 artefacts. The compositional analyses were conducted using different methods. The elemental
102 composition of filters sampled in August and September weeks, were measured by ED-XRF
103 (Energy Dispersive - X Ray Fluorescence) using an ED-2000 spectrometer from Oxford
104 Instruments (Ariola et al., 2006) for S, Cl, K, Ca, Ti, V, Cr, Mn, Fe, Ni, Cu, Zn, Br, Ba, Pb.
105 For technical reasons, the concentrations of the same elements in the samples collected during

106 the July cruise were indeed determined by PIXE analysis at the HVEE 3 MV Tandatron
107 accelerator, installed at the LABEC (LABoratorio BENi Culturali) laboratory of INFN in
108 Florence (Calzolari et al., 2006; Lucarelli et al., 2013). The concentration values of S and K
109 determined using ED-XRF were corrected for an average attenuation factor (Bove et al.,
110 2014) to determine their mean values, whereas S, Cl, K resulted to be always below their
111 Minimum Detection Limit when measured by PIXE. The Minimum Detection Limits obtained
112 for both the techniques are shown in Table E1 in the electronic supplementary material.
113 Finally, the analytical uncertainties are the sum of the systematic term on the calibration
114 standards (5%) and of the statistical fluctuation on peak areas.

115 The water-soluble inorganic components of the PM₁₀ were determined by Ion
116 Chromatography (IC) using an ICS-1000 Ion Chromatography System (Dionex) at the
117 University of Milan. In particular, for the extraction of the PM, a quarter of each filter was
118 wetted previously and then three times with MilliQ water in an ultrasonic bath for 20 min
119 (complete recovery, $98\% \pm 3\%$), renewing the water at each step (Piazzalunga et al., 2013).
120 The extracts were analysed using IC to identify the major ionic species (i.e., Na^+ , NH_4^+ , K^+ ,
121 Mg^{2+} , Ca^{2+} , Cl^- , SO_4^{2-} , NO_3^-) with an overall 10% uncertainty for the ionic concentrations. The
122 MSA (methanesulfonic acid) concentration values were also measured by IC with the same
123 uncertainty. The lack of quantification of low-Z elements (due to the X-ray self-absorption
124 and the high Si concentration in the quartz filters) was partially recovered by Ionic
125 Chromatography analysis which was finally considered more accurate for such elements.

126 Information on meteorological parameters (wind speed and direction, temperature,
127 humidity from the meteorological station of the ship) and on the ships position, speed and
128 sailing direction, were also available (in 10 min intervals) and used to identify situations
129 where the PM sampling might be influenced by the emissions of Costa Concordia itself.
130 When the inlets of the measurement station were downwind the ship stack within an angle of
131 $\pm 40^\circ$, the data were discarded to avoid any risk of contamination.

132 Air mass back-trajectories were calculated using the US NOAA HYSPLIT model
133 (<http://ready.arl.noaa.gov/HYSPLIT.php>) with GDAS meteorological data. For each filter,
134 five-day back trajectories arriving at 50 m and 500 m above sea level were calculated for the
135 positions where the filter sampling ended, to evaluate the different air masses arriving over the
136 sea in the three cruise weeks. During summer 2011, the route of the ship was Civitavecchia-
137 Savona-Barcelona-Palma de Mallorca-Malta (Valletta)-Palermo-Civitavecchia (see Figure E1
138 in the electronic supplementary material).

139

140 2.3 Aerosol composition: mass closure

141 Details on the method to obtain the aerosol composition is described in Schembari et
142 al. (2014). Briefly, concentration values of SO_4^{2-} , NH_4^+ and NO_3^- were directly retrieved from
143 the IC analysis, while sea salt and dust were obtained from raw data and conversion factors:
144 sea salt was calculated from Na^+ and Cl^- concentration values, taking into account the
145 seawater composition (Seinfeld and Pandis, 1998); dust was obtained by multiplication of
146 non-sea-salt calcium nssCa^{2+} (calculated by subtracting from total measured Ca^{2+} , the fraction
147 in sea salt given by multiplication of Na^+ by $\text{Ca}^{2+}:\text{Na}^+$ ratio in seawater composition, Seinfeld
148 and Pandis, 1998) by 5.6 (the value retrieved by Putaud et al. (2004) for a background site,
149 would vary for any other kind of station).

150

151 2.4 Receptor model-PMF

152 Positive Matrix Factorization (PMF) was used to identify and characterize the major
153 PM10 sources along the ship route. PMF has been described in detail by its developers
154 (Paatero and Tapper, 1994), it has been adopted in several studies for PM receptor modelling
155 and has rapidly become a reference tool in this research field (e.g., Qin et al., 2006; Escrig et
156 al., 2009; Contini et al., 2012; Cuccia et al., 2013). In this work, the PMF2 program (Paatero,
157 2010) and the methodology described in Bove et al. (2014) was used. The PMF analyses were
158 carried out using the data collected from the three weeks of the summer 2011. The variables
159 were selected according to the signal-to-noise criterion (Paatero and Hopke, 2003) and 14
160 series of concentration values were finally retained for the PMF study: Ti, V, Fe, Ni, MSA,
161 Cl^- , NO_3^- , SO_4^{2-} , Na^+ , NH_4^+ , K^+ , Mg^{2+} , Ca^{2+} , BC. The Polissar et al. (1998) procedure was
162 used to assign concentration data and their associated uncertainties; the Cl^- , NO_3^- , Na^+ , Mg^{2+}
163 uncertainties only were increased of 20% in the PMF runs for down weighting these elements
164 which resulted ubiquitous among the factors. The number of samples considered in the PMF
165 run (55) satisfies the criteria set in Thurston and Spengler, (1985). PMF results are affected by
166 the rotational ambiguity (Paatero et al., 2002) and rotations are directly implemented in the
167 minimisation algorithm using the FPEAK parameter (Paatero, 1997). In the analysis, the
168 parameters obtained from the scaled residual matrix, IM (the maximum individual column
169 mean), and IS (the maximum individual column standard deviation), together with Q-values
170 (goodness of fit parameter) were examined to find the most reasonable solution. The best
171 rotation for each factor was chosen in the FPEAK range from -2 to +2 by discarding the
172 solutions corresponding to profiles without physical meaning (i.e., the sum of elemental
173 concentrations exceeded 100%) and selecting those generating concentration ratios between

174 the tracer elements of the natural sources (e.g., sea salt, crustal matter) comparable to
175 literature values (Bove et al., 2014).

176

177 **3. Results and discussion**

178

179 *3.1 Meteorological conditions*

180 The sea level pressure composite mean and anomalies over the Mediterranean basin
181 during the three campaigns according to the NCEP/NCAR Reanalysis (Kalnay et al. 1996),
182 are shown in Figure E2 in the electronic supplementary material. While in August and in
183 September the synoptic conditions were characterized by the expansion towards the
184 Mediterranean of the Azores Anticyclone, in line with seasonal climatology (especially in
185 August, whereas a slightly negative anomaly is found in September), in July the situation was
186 very peculiar. In this case, the anticyclonic system is confined over the Atlantic, favouring the
187 development of low-pressure systems across Central Europe and the Mediterranean Basin,
188 where a strong negative pressure anomaly can be seen.

189 The meteorological parameters recorded during the three cruises by the on-board
190 instrumentation are reported in Figure E3 in the electronic supplementary material and
191 confirm what is suggested by the synoptic analysis. In particular, pressure exhibited lower
192 average values and larger variability in July, associated to episodes of strong wind and, as a
193 consequence, rough sea along the route. On the contrary, during the two campaigns in August
194 and September, more stable conditions were encountered, with higher pressure values and
195 generally lighter winds, apart from the last leg of the September cruise, when the passage of
196 an Atlantic frontal system determined a sudden pressure drop and wind speed increase.

197 The meteorological conditions along the ship route during the most relevant strong
198 wind episodes were also assessed using a 32-year hindcast, recently realized at the University
199 of Genoa by means of simulations with the Weather Research and Forecasting (WRF,
200 Skamarock et al. 2008) model on a domain covering the entire Mediterranean with a
201 horizontal grid spacing of 10 km. Details about the modelling system are given in Mentaschi
202 et al. (2015).

203

204 *3.2 PM10 composition*

205 The average PM10 concentration and its composition are reported in Table 1 whereas
206 in Figure 1 the chemical composition as described in 2.3 Section for the three 2011 cruises, is
207 shown. The nssSO_4^{2-} , NO_3^- , sea salt seem to be quite different between the July campaign and

208 the other two cruise weeks (Figure 1). Such a discrepancy is attributable to the peculiar
209 meteorological conditions occurred in July, as discussed in the previous section. The balance
210 between nitrate and ammonium sulphate also shows two different well defined situations.
211 During July and for some samples collected in September, ammonium and nitrate ions exactly
212 balance, this highlighting the lack of ammonium sulphate. On the contrary, in August the sum
213 of nitrate and sulphate ions are completely balanced by ammonium (Figure 2a). Furthermore,
214 the ratio $(\text{SO}_4^{2-} + \text{NO}_3^-) : \text{NH}_4^+$ shows increases correlated with Cl^- concentration values
215 (Figure 2b). Figure 3 shows the anti-correlation between chlorine and sulphate concentration
216 values: all these pieces of information point at the presence of air masses of two different
217 origin which affect our samples as also indicated by back trajectory analysis. The air masses
218 reaching the ship route in July, had been mainly over the sea for at least the previous 24 h;
219 during the August and September cruises, the impacting air masses passed mostly over the
220 continental areas, suggesting a larger contribution from the transport of terrestrial pollutants to
221 the open sea. An example is shown in Figure E4 of the electronic supplementary material. In
222 conclusion, when the air masses reached the ship coming from the continent, sulphate
223 concentrations increased and the ratio $(\text{SO}_4^{2-} + \text{NO}_3^-) : \text{NH}_4^+$ approached 1. On the contrary,
224 sulphate concentration values remained low without any sizeable presence of ammonium
225 sulphate when the aerosol impacting the ship was mainly of marine origin.

226 The primary contribution of ship emissions to PM10 can be calculated on the basis of
227 previous research works (Agrawal et al., 2009; Zhao et al., 2013) and using the equation:

$$228 \quad 229 \quad PM_a = R \frac{V_a}{F_{V,HFO}} \quad (1)$$

230
231 $R = 8205.8$ is the average ratio of PM2.5 to normalized V emitted (ppm) suggested in
232 Agrawal et al. (2009), which could be universally applied to other locations with HFO
233 burning ship emissions; V_a is the ambient concentration of V (ng m^{-3}), whilst $F_{V,HFO}$ is a term
234 indicating the typical V content (in ppm) in HFOs used by vessels. We used the same average
235 value of $F_{V,HFO} = (65 \pm 25)$ ppm, in agreement with Cesari et al., (2014). According to eq.
236 (1), the primary PM10 from ship traffic ranged from 0.7 to 3.4 $\mu\text{g m}^{-3}$; similar values had been
237 previously obtained in some port sites (Viana et al., 2009).

238

239 *3.3 Sulphate apportionment*

240 The contributions of different sources to the sulphate concentration was evaluated on
241 the basis of specific markers as described in Table E2 in the electronic supplementary
242 material. The main components of the sulphate are the sea salt sulphate ($ssSO_4^{2-}$), that is the
243 amount of sulphate present in sea salt particles, and non-sea-salt sulphate. The non-sea-salt
244 sulphate ($nssSO_4^{2-}$) is defined as the amount of the sulphate present in particles in excess of
245 what expected from sea salt particles, and has three contributions: anthropogenic, biogenic
246 and crustal $nssSO_4^{2-}$. According to some literature works, methanesulfonic acid can be used as
247 a marker for quantifying the biogenic non-sea salt sulphate ($nssSO_4^{2-}_{bio}$). The ratio between
248 MSA and $nssSO_4^{2-}_{bio}$ depends on the season (Kouvarakis et al., 2002), latitude (Chen et al.,
249 2012) and temperature (Bates et al. 1992). In a previous work (Schembari et al., 2014), the
250 $nssSO_4^{2-}_{bio}$ was estimated starting from the measurement of MSA concentration in the samples
251 through the relation by Bates et al. (1992) even if that equation was obtained during of field
252 campaign in the eastern Pacific Ocean.

253 In our data set we identified a sample marked in Figure 3, that seems to be affected by
254 a strong presence of fresh marine aerosols: the ratio $Cl^-:Na^+ = 1.094$ and $Mg^+:Na^+ = 0.24$ are
255 very similar to those reported in literature for the fresh marine aerosol, respectively 1.17, and
256 0.25 (Keene et al., 1986). Furthermore, the ionic balance is fully respected: 162 neq m^{-3} of
257 anions vs. 160 neq m^{-3} of cations and the sulphate is not fully balanced from ammonium, this
258 highlighting the presence of a sulphates source different from the anthropogenic ones. With
259 this sample, the direct calculation (details are shown in Table E2b in the electronic
260 supplementary material) of the $nssSO_4^{2-}_{bio}$ concentration using the diagnostic ratios reported
261 in literature work (Keene et al., 1986), is possible and the $MSA:nssSO_4^{2-}_{bio}$ ratio was found to
262 be 0.08 against the value of 0.03 which would result from the Bates's formula. Finally, we
263 adopted $MSA:nssSO_4^{2-}_{bio} = 0.08$ for the whole campaign and to derive the sulphate
264 apportionment.

265 The results of such calculations for the three cruises are reported in Table 2 and Figure
266 8. Large concentration values of $nssSO_4^{2-}$ were obtained for all the three weeks, while highest
267 values of $ssSO_4^{2-}$ and lowest values of $nssSO_4^{2-}$ were observed in July (Figure 1). The latter
268 were in coincidence with a quite high wind speed, in particular during the Savona-Barcelona
269 and Palermo-Civitavecchia legs. The analysis of wind speed and direction, both measured on
270 board and obtained by hindcast simulations with the WRF-ARW model (see also Figure E5
271 and E6 in the electronic supplementary material), highlighted that it blew from the sea and its
272 velocity increased rapidly during the last part of the routes, close to the Barcelona coast and to
273 Civitavecchia, respectively. This observation confirms the sea salt dependence on the local

274 wind speed in the Mediterranean Basin (Bergametti et al. (1989a) Chabas and Lefèvre (2000);
275 Contini et al., 2010). The main contribution to nssSO_4^{2-} was of anthropic origin in August and
276 September, whereas in July $\text{nssSO}_4^{2-}_{\text{bio}}$ was prevailing, probably due to the particular
277 meteorological conditions that determined high sea salt concentrations. The $\text{nssSO}_4^{2-}_{\text{anthr}}$
278 contributed to PM10 by 5%, 25% and 18%, in July, August and September, respectively,
279 while the $\text{nssSO}_4^{2-}_{\text{bio}}$ was on average 9%, 5% and 3% of PM10 in the same periods. The
280 $\text{nssSO}_4^{2-}_{\text{crust}}$, as estimated by this approach, remained always around 1% of PM10.

281 The above-discussed results can be compared with those collected in the similar cruise
282 in June 2010 (Schembari et al., 2014), also reported in Table 2. The previous sulphate
283 apportionment showed similar contributions to 2011 data for $\text{nssSO}_4^{2-}_{\text{crust}}$ and ssSO_4^{2-} . Indeed,
284 the meteorological conditions in June 2010 were quite resembling those found in July 2011,
285 with cool temperatures and intense winds (see Schembari et al. 2014). The $\text{nssSO}_4^{2-}_{\text{bio}}$ showed
286 higher average contributions if compared to values obtained in the three 2011 cruises. Higher
287 values of $\text{nssSO}_4^{2-}_{\text{anth}}$ were also obtained in 2010 with respect to the 2011 campaigns, with the
288 only exception of the August cruise, when higher levels of $\text{nssSO}_4^{2-}_{\text{anth}}$ were found.

289

290 3.4 PMF results

291 The database used as input to PMF included data obtained by the analysis of filters
292 sampled along open-sea legs while samples collected when the ship was manoeuvring or
293 hotelling in the harbours and when the sampling station was downwind the ship stack, were
294 excluded. The database was completed with the time series of hourly BC concentration values
295 and PM10 mass concentration.

296 Five factors were resolved and identified by PMF for PM10 obtaining the best solution
297 with $\text{FPEAK} = 0$: *Secondary Sulphate*, *Reacted dust*, *Biomass burning*, *Sea salt* and *Heavy oil*
298 *combustion*. Source profiles and explained variations (EV) parameters are shown in Figure 4,
299 while the average PM10 apportionment is given in Figure 5.

300 PMF-Factor 1 was identified as the contribution due to *Secondary Sulphate* looking at
301 the high EVs for SO_4^{2-} and NH_4^+ and the relevance of these compounds in the chemical profile
302 (Figure 4). The average concentration ratio for $\text{SO}_4^{2-}:\text{NH}_4^+$ in the factor is 2.1 ± 0.1 , which is
303 slightly lower than the stoichiometric figure for ammonium sulphate (i.e. $\text{SO}_4^{2-}:\text{NH}_4^+ = 2.7$).
304 The average relative contribution of this factor to the PM10 mass is $(35 \pm 5)\%$, with highest
305 concentrations observed during August and lowest in July as reported in Table 3. The PMF
306 result is comparable, within its uncertainty, with the direct calculation of the average
307 abundance of ammonium non-sea-salt sulphate in PM10 of $(39 \pm 4)\%$, discussed in Section

308 3.2. The quite low concentration value of July reported in Figure 6 confirms the observations
309 obtained by the measurements.

310 PMF-Factor 2 was characterised by high EV values for Ti and Fe, this suggesting a
311 contribution by mineral dust, and by a relevant fraction of SO_4^{2-} , NO_3^- , NH_4^+ and BC in the
312 source profile (Figure 4). The mineral particles aged in the atmosphere and then changed their
313 original composition, getting mixed/coated with organic and inorganic ions (sulphate and
314 nitrate) and BC (Fairlie et al., 2010). For this reason, this factor was labelled as *Reacted dust*,
315 also in agreement with other source profiles obtained by PMF in Mediterranean sites (Perrone
316 et al., 2013, Cesari et al., 2014). The temporal pattern of this factor showed highest
317 concentrations along the Barcelona-Palma legs (see also Figure 6), in particular near the
318 Palma coast. Moreover, this source profile is quite similar to the mineral dust profile obtained
319 by PMF analysis of the data sampled in a site located at Palma de Mallorca (Pey et al., 2013),
320 which includes anthropogenic dust emissions from the harbour too. For this reason, the
321 fraction of PM10 attributed by PMF to *Reacted dust*, even if it appears consistent with the
322 “chemical” apportionment described in Section 2.3, is not comparable to the pure dust
323 composition.

324 PMF-Factor 3 was assigned to *Biomass burning* because it was characterized by high
325 contributions of BC, SO_4^{2-} , NH_4^+ and K^+ in the source profile (Figure 4) and by high EV
326 values for BC and K^+ in agreement with other works which adopted K^+ as tracer of biomass
327 burning (Belis et al., 2011). High concentration values were detected along the Malta-Palermo
328 leg, both in August and September (see also Figure 6). Maximum values were observed with
329 high wind speed and prevailing direction from the Sicilian coast and from the city of Palermo.
330 The contribution of the source, on average (27 ± 5)% of PM10, seems to be excessive
331 considering the summer period in which the measurements were performed. Actually, this is
332 not a pure profile because includes the mixing with other sources like re-suspended dust
333 coming from the continents nearby. This aspect is confirmed also by the presence of Ca^{2+}
334 element, which instead is absent in the “Reacted dust” profile.

335 PMF-Factor 4 was identified as *Sea salt* since it was characterized by high EV values
336 for NO_3^- , Cl^- , Na^+ , Mg^{2+} and MSA (Figure 4). The $\text{Cl}^-:\text{Na}^+$ ratio in the profile is equal to 0.2,
337 which is much smaller than both the 0.9 mean ratio obtained in the 2009 and 2010 cruises
338 (Schembari et al., 2014) and the 1.17 ratio of fresh sea salt particles (Keene et al., 1986). This
339 can be due to evaporation of HCl to the atmosphere which occurs in marine air samples
340 (Perrone et al., 2013, Cuccia et al., 2013). The PMF algorithm could not distinguish fresh and
341 aged sea salt: in the *Sea salt* source profile (Figure 4), the presence of the secondary nitrates

342 and MSA^- due to the oxidation of dimethyl sulphide emitted from the sea suggested the
343 mixing with secondary components of PM_{10} . The average fraction of PM_{10} attributed to this
344 factor was $(19 \pm 4)\%$, in agreement with the $(27 \pm 5)\%$ value obtained as the sum of Sea salt
345 and Nitrates components obtained evaluated by chemical analysis (Section 2.3). The sea salt
346 concentration was higher in July than in August and September as highlighted in Table 3: this
347 confirms the occurrence of sea salt events during the Savona-Barcelona and Palermo-
348 Civitavecchia legs as described in Section 3.2.

349 PMF-Factor 5 was finally identified as *Heavy oil combustion* because it was
350 characterized by high EV values for V and Ni, typical tracers of heavy oil combustion
351 (Mazzei et al., 2008, Viana et al., 2009). The V:Ni concentration ratio in the source profile is
352 2.6 ± 0.1 , in agreement with the 2.9 ± 0.4 value obtained by PMF during the previous
353 campaigns (Schembari et al., 2014) and with the conclusions of several other literature works
354 which recognized such value as typical of ship emissions (Agrawal et al., 2008, Mazzei et al.,
355 2008, Cuccia et al., 2010, Pandolfi et al., 2011, Bove et al., 2014). The source profile was
356 enriched in sulphate with $\text{SO}_4^{2-}:\text{V} = 67 \pm 4$. The initial $\text{SO}_4^{2-}:\text{V}$ ratio in the particulate exhaust
357 ($\text{PM}_{2.5}$) of the main engine of different oceangoing container vessels is reported to be in the
358 range 11–27 (Agrawal et al., 2008). However, the amount of SO_4^{2-} in the air mass is expected
359 to grow fast due to SO_2 conversion into sulphate; this conversion is faster in high UV
360 radiation and high humidity conditions (Restad et al., 1998, Becagli et al., 2012). Actually, the
361 measured $\text{SO}_4^{2-}:\text{V}$ ratio (similar to the $\text{SO}_4^{2-}:\text{V}$ ratio in the profile) is lower in July than the
362 other two cruise weeks, confirming the higher influence of marine air masses as observed in
363 Section 3.2 and, therefore, of the ship emissions along the route. Ship emissions contributed
364 on average to $(12 \pm 4)\%$ of PM_{10} . This figure is in agreement with the $(16 \pm 11)\%$ percentage
365 evaluated considering the measured V as a marker for the combustion in ship engines (3.2
366 Section).

367 The apportionment of single PM_{10} species is given in Figure 7. Notably, NO_3^- was
368 mainly associated with *Sea salt* (on average 95%) supporting the nature of aged marine source
369 (Cuccia et al., 2013), whereas NH_4^+ was primarily associated with one of the secondary
370 components of PM_{10} , i.e. *Secondary Sulphate* (on average: 80%). On average, $(23 \pm 9)\%$ of
371 the SO_4^{2-} was attributed to *Heavy oil combustion*. The Sulphate apportionment resolved by
372 PMF appears to be different in the three cruises (see also Figure 8). The apportionment seems
373 to be quite similar in August and September while an increase of the total SO_4^{2-} attributed to
374 *Heavy oil combustion* in association with the Sea salt events (3.3 Section) was observed in
375 July. The latter can be explained by the possible contamination in the *Heavy oil combustion*

376 profile of the biogenic fraction of the sulphates (the measured biogenic sulphate was much
377 larger than the anthropogenic one in July); this is probably due to the influence of
378 meteorological conditions and air masses which remained over the sea for several hours,
379 producing the association of both sources. Moreover, the average measured MSA: nssSO_4^{2-}
380 ratio for the three cruise weeks is the same value found in the *Heavy oil combustion* factor
381 obtained by PMF analysis to support the biogenic contamination of the sulphate in the profile.

382

383 3.3.1 Sources comparison

384 The new study provided more complete and clear information than the analysis
385 performed in the past years (Schembari et al., 2014). Due to the lack of a complete chemical
386 speciation, only four sources were resolved in 2010 and in particular the PMF did not resolve
387 secondary and primary sources of sulphate. A *Combustion* source only, which showed
388 evidence of a contribution by ship emissions, was found to contribute by $(55 \pm 4)\%$ to PM10.
389 The main scope of the 2011 experiment was to separately quantify the contribution of ship
390 emissions and of secondary sulphate to PM10. This objective was achieved: in 2011 the
391 *Secondary Sulphate* and *Heavy oil combustion* were found to account for $(35 \pm 5)\%$ and $(12 \pm$
392 $4)\%$ of PM10, respectively. The *Combustion* factor identified in the previous campaigns is
393 comparable with the sum of *Secondary Sulphate* and *Heavy oil combustion* sources in 2011.
394 Moreover, the source *Not identified* by PMF in 2009-2010 was recognized as *Biomass*
395 *burning* with the 2011 dataset since it is characterized and traced by high contributions of BC
396 and K^+ . For *Sea salt* and *Reacted dust* a similar mean contribution to PM10 was obtained in
397 2010 and 2011.

398 The names given to the sources of the five PMF factors obviously represent a
399 simplification; it is clear that there must be several additional minor sources that have
400 contributed to the observed aerosol composition; in particular, land-based traffic and
401 industrial sources. Thus, the five source profiles are not representing 'pure' sources and the
402 names given to them will only reflect what is believed to be the principal source contributing
403 to this profile.

404

405 4. Conclusions

406

407 PM10 aerosol samples collected during three campaigns on board a cruise ship from
408 July to September 2011 were analysed to determine their chemical composition and to
409 improve the source apportionment obtained during previous studies performed on board

410 cruise ships in the Western Mediterranean. The biogenic fraction of the sulphate was
411 prevailing during the July campaign, together with a higher contribution of the ship emissions,
412 probably due to the influence of predominantly marine air masses along the ship route. Five
413 sources were resolved and identified by PMF analysis with the new data sets: *Secondary*
414 *Sulphate*, *Reacted dust*, *Biomass burning*, *Sea salt* and *Heavy oil combustion*. Heavy oil
415 combustion by ship engines was identified using V and Ni as tracers. Secondary ammonium
416 sulphate was found to be an important source of aerosol in Western Mediterranean. The
417 experiment allowed the identification of a contribution of primary ship emissions to PM10.
418 This contribution turned out to be $(12 \pm 4)\%$, while secondary ammonium sulphate
419 contributed by $(35 \pm 5)\%$. Approximately 60% of the total sulphate was attributed to
420 secondary sources and around 20% was attributed to *Heavy oil combustion* considering the
421 measuring campaigns not influenced by strong sea salt events.

422

423

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427

428

References

429

430 Agrawal, H., Welch, W.A., Miller, J.W., Cocker, D.R., 2008. Emission measurements from a
431 crude oil tanker at sea. *Environmental Science Technology* 42, 7098-7103.

432

433 Agrawal, H., Eden, R., Zhang, X., Fine, P.M., Katzenstein, A., Miller, J.W., Ospital J.,
434 Teffera S., Cocker D.R., 2009. Primary particulate matter from ocean-going engine in the
435 southern California air basin. *Environmental Science Technology* 43, 398-402.

436

437 Ariola, V., D'Alessandro, A., Lucarelli, F., Marcazzan, G., Mazzei, F., Nava, S., Garcia
438 Orellana, I., Prati, P., Valli, G., Vecchi, R., Zucchiatti, A. 2006. Elemental characterization of
439 PM10, PM2.5 and PM1 in the town of Genoa (Italy). *Chemosphere* 62, 226-232.

440

441 Bates, T., Calhoun, J., Wang, Y., Quinn, P., 1992. Variations in the methanesulphonate to
442 sulphate molar ratio in submicrometer marine aerosol particles over the south pacific ocean.
443 *Journal of Geophysical research* 97, 9859-9865.

- 444 Becagli, S, Sferlazzo, D.M., Pace, G., Di Sarra, A., Bommarito, C., Calzolari, G., Ghedini,
445 C., Lucarelli, F., Meloni, D., Monteleone, F., Severi, M., Traversi, R. and Udisti R., 2012.
446 Evidence for heavy fuel oil combustion aerosols from chemical analyses at the island of
447 Lampedusa: a possible large role of ships emissions in the Mediterranean. *Atmospheric*
448 *Chemistry and Physics* 12, 3479–92.
- 449
- 450 Belis, C.A., Cancelinha, J., Duane, M., Forcina V, Pedroni V, Passarella R, Tanet, G.,
451 Douglas, K., Piazzalunga, A., Bolzacchini, E., Sangiorgi, G.M.L., Perrone, M.G., Ferrero, L.,
452 Fermo, P., Larsen, B.R., 2011. Sources for PM air pollution in the Po Plain, Italy: I. Critical
453 comparison of methods for estimating biomass burning contributions to benzo(a)pyrene.
454 *Atmospheric Environment* 45, 7266–75.
- 455
- 456 Belis C.A., Karagulian, F., Larsen, B.R., Hopke, P.K, 2013. Critical review and meta-analysis
457 of ambient particulate matter source apportionment using receptor models in Europe.
458 *Atmospheric Environment* 69, 94-108.
- 459
- 460 Bergametti, G., Dutot, A.-L., Buat-Menard, P., Losno, R., and Remoudaki, E., 1989a.
461 Seasonal variability of the elemental composition of atmospheric aerosol particles over the
462 north western Mediterranean, *Tellus* 41B, 353–361.
- 463
- 464 Birch, M. E., Cary, R. A., 1996. Elemental carbon-based method for monitoring occupational
465 exposures to particulate diesel exhaust. *Aerosol Science and Technology* 25, 221-241.
- 466
- 467 Bove, M.C., Brotto, P., Cassola, F., Cuccia, E., Massabò, D., Mazzino, A., Piazzalunga, A.,
468 Prati, P., 2014. An integrated PM_{2.5} source apportionment study: Positive Matrix
469 Factorisation vs. the Chemical Transport Model CAMx. *Atmospheric Environment* 94, 274-
470 286.
- 471
- 472 Calzolari, G., Chiari, M., García Orellana, I., Lucarelli, F., Migliori, A., Nava, S., Taccetti, F.,
473 2006. The new external beam facility for environmental studies at the Tandatron accelerator
474 of LABEC. *Nuclear Instrument Methods* B249 (1-2), 928-931.
- 475

- 476 Cavalli, F., Putaud, J.P., Viana, M., Yttri, K.E., Gernberg, J., 2010. Toward a standardised
477 thermal-optical protocol for measuring atmospheric organic and elemental carbon: the
478 EUSAAR protocol. *Atmospheric Measurement Techniques* 3, 79-89.
- 479
- 480 Cesari, D., Genga, A., Ielpo, P., Siciliano, M., Mascolo, G., Grasso, F.M., Contini, D., 2014.
481 Source apportionment of PM_{2.5} in the harbour–industrial area of Brindisi (Italy):
482 Identification and estimation of the contribution of in-port ship emissions. *Science of the*
483 *Total Environment* 497–498, 392–400.
- 484
- 485 Chabas, A. and Lefèvre, R. A., 2000. Chemistry and microscopy of atmospheric particulates
486 at Delos (Cyclades-Greece). *Atmospheric Environment* 34, 225–238.
- 487
- 488 Chen, L. Wang, J., Gao, Y., Xu, G., Yang, X., Lin, Q., Zhang, Y., 2012. Latitudinal
489 distributions of atmospheric MSA and MSA/nss-SO₄²⁻ ratios in summer over the high latitude
490 regions of the Southern and Northern Hemispheres. *Journal of geophysical research* 117,
491 D10306, doi:10.1029/2011JD016559.
- 492
- 493 Chiari, M., Lucarelli, F., Mazzei, F., Nava, S., Paperetti, L., Prati, P., Valli, G., Vecchi, R.
494 2005. Characterization of airborne particulate matter in an industrial district near Florence by
495 PIXE and PESA. *X-Ray Spectrometry* 34, 323-329.
- 496
- 497 Chow, J.C., Watson, J.G., 1999. Ion chromatography in elemental analysis of airborne
498 particles. In: Landsberger, S., Creatchman, M. (Eds.), *Elemental Analysis of Airborne*
499 *Particles*, vol. 1. Gordon and Breach Science, Amsterdam, 97-137.
- 500
- 501 Contini, D., Genga, A., Cesari, D., Siciliano, M., Donateo, A., Bove, M.C., Guascito, M.R.,
502 2010. Characterisation and source apportionment of PM₁₀ in an urban background site in
503 Lecce. *Atmospheric Research* 95 (1), 40-54.
- 504
- 505 Contini, D., Belosi, F., Gambaro, A., Cesari, D., Stortini, A., Bove, M.C., 2012. Comparison
506 of PM₁₀ concentrations and metal content in three different sites of the Venice Lagoon: an
507 analysis of possible aerosol sources. *Journal of Environmental Science* 24 (11), 1954-1965.
- 508

- 509 Cuccia, E., Bernardoni, V., Massabò, D., Prati, P., Valli, G., Vecchi, R., 2010. An alternative
510 way to determine the size distribution of airborne particulate matter. *Atmospheric*
511 *Environment* 44, 3304-3313.
- 512
- 513 Cuccia, E., Massabò, D., Ariola, Bove, M.C., Fermo, P., Piazzalunga, A., Prati, P., 2013. Size
514 resolved comprehensive characterization of airborne particulate matter. *Atmospheric*
515 *Environment* 67, 14-26.
- 516
- 517 Eyring, V., Köhler, H.W., Van Aardenne, J., Lauer, A., 2005. Emissions from international
518 shipping: 1. The last 50 years. *Journal of Geophysical Research Atmosphere* 110, D17305.
- 519
- 520 Escrig, A., Monfort, E., Celades, I., Querol, X., Amato, F., Mingiullon, M.C., Hopke, P.K.,
521 2009. Application of optimally scaled target factor analysis for assessing source contribution
522 of ambient PM10. *Journal of Air Waste Management Association* 59, 1296-1307.
- 523
- 524 Fairlie, T.D., Jacob, D.J., Dibb, J.E., Alexander, B., Avery, M.A., Van Donkelaar, A., Zhang,
525 L., 2010. Impact of mineral dust on nitrate, sulphate, and ozone in transpacific Asian pollution
526 plumes. *Atmospheric Chemistry and Physics* 10, 3999-4012.
- 527
- 528 Formenti, P., Prati, P., Zucchiatti, A., Lucarelli, F., Mandò, P.A., 1996. Aerosol study in the
529 Genova area via a two stage streaker and PIXE analysis. *Nuclear Instrument Methods B113*,
530 359-362.
- 531
- 532 Kalnay, E., Kanamitsu, M., Kistler, R., Collins, W., Deaven, D., Gandin, L., Iredell, M., Saha,
533 S., White, G., Woollen, J., Zhu, Y., Leetmaa, A., Reynolds, R., Chelliah, M., Ebisuzaki, W.,
534 Higgins, W., Janowiak, J., Mo, K. C., Ropelewski, C., Wang, J., Jenne, R., Joseph, D., 1996.
535 The NCEP/NCAR 40-Year Reanalysis Project, *Bulletin of the American Meteorological*
536 *Society* 77, 437-471.
- 537
- 538 Keene, W. C., Pszenny, A. A. P., Galloway, J. N. Hawley M. E., 1986. Sea-Salt Corrections
539 and Interpretation of Constituent Ratios in Marine Precipitation. *Journal of Geophysical*
540 *research* 91, 6647-6658.
- 541

- 542 Kouvarakis, G. and Mihalopoulos, N., 2002. Seasonal variation of dimethylsulfide in the gas
543 phase and of methanesulfonate and non-sea-salt sulphate in the aerosols phase in the Eastern
544 Mediterranean atmosphere. *Atmospheric Environment* 36, 929–938.
- 545
- 546 Lelieveld, J., Berresheim, H., Borrmann, S., Crutzen, P.J., Dentener, F., 2002. Global air
547 pollution crossroads over the Mediterranean. *Science* 298, 794-799.
- 548
- 549 Lucarelli, F., Calzolari, G., Chiari, M., Giannoni, M., Mochi, D., Nava, S., Carraresi, L., 2013.
550 The upgraded external-beam PIXE/PIGE set-up at LABEC for very fast measurements on
551 aerosol samples. *Nuclear Instruments and Methods in Physics Research*. [http://
552 dx.doi.org/10.1016/j.nimb.2013.05.099](http://dx.doi.org/10.1016/j.nimb.2013.05.099).
- 553
- 554 Mazzei, F., D'Alessandro, A., Lucarelli, F., Nava, S., Prati, P., Valli, G., Vecchi, R., 2008.
555 Characterization of particulate matter sources in an urban environment. *Science of Total
556 Environment* 401, 81-89.
- 557
- 558 Mentaschi, L., Besio, G., Cassola, F., Mazzino, A., 2015. Performance evaluation of
559 WavewatchIII in the Mediterranean Sea. *Ocean Modelling* 90, 82-94.
- 560
- 561 Moreno, T., Perez, N., Querol, X., Amato, F., Alastuey, A., Bhatia, R., Spiro, B., Hanvey, M.,
562 Gibbons, W., 2010. Physicochemical variations in atmospheric aerosols recorded at sea on
563 board the Atlantic-Mediterranean 2008 Scholar Ship cruise (Part II): natural versus
564 anthropogenic influences revealed by PM10 trace element geochemistry. *Atmospheric
565 Environment* 44, 2563-2576.
- 566
- 567 Paatero, P., Tapper, U., 1994. Positive matrix factorization: a non-negative factor model with
568 optimal utilization of error estimates of data values. *Environmetrics* 5, 111-126.
- 569
- 570 Paatero, P., 1997. Least squares formulation of robust, non-negative factor analysis.
571 *Chemometrics Intelligent Laboratory Systems* 37, 23-35.
- 572
- 573 Paatero, P., Hopke, P.K., Song, X.H., Ramadan, Z., 2002. Understanding and controlling
574 rotations in factor analytic models. *Chemometrics and Intelligent Laboratory Systems* 60,
575 253-264.

- 576 Paatero, P., Hopke, P.K., 2003. Discarding or down weighting high-noise variables in factor
577 analytic models. *Analytica Chimica Acta* 490, 277-289.
- 578
- 579 Paatero, P., 2010. User's Guide for Positive Matrix Factorization Programs PMF2 and PMF3,
580 Part 1: Tutorial. University of Helsinki, Finland.
- 581
- 582 Perrone, M.R., Becagli, S., Garcia Orza, J.A., Vecchi, R., Dinoi A, Udisti, R., Cabello, M.,
583 2013. The impact of long-range-transport on PM1 and PM2.5 at a Central Mediterranean site.
584 *Atmospheric Environment* 71, 176-186.
- 585
- 586 Pey, J, Alastuey, A. Querol, X., 2013. PM10 and PM2.5 sources at an insular location in the
587 western Mediterranean by using source apportionment techniques. *Science of the total*
588 *Environment* 456-457, 267-277.
- 589
- 590 Piazzalunga, A., Bernardoni, V., Fermo, P., Vecchi, R., 2013. Optimisation of analytical
591 procedures for the quantification of ionic and carbonaceous fractions in the atmospheric
592 aerosol and applications to ambient samples. *Analytical and Bioanalytical Chemistry* 405,
593 1123-1132.
- 594
- 595 Polissar, A.V., Hopke, P.K., Paatero, P., Malm, W.C., Sisler, J.F., 1998. Atmospheric aerosol
596 nucleation and primary emission rates. *Atmospheric Chemistry and Physics*, 1339-1356.
- 597
- 598 Putaud, J.P., Van Dingenen, R., Dell'Acqua, A., Raes, F., Matta, E., Decesari, S., Facchini,
599 M.C., Fuzzi, S., 2004. Size-segregated aerosol mass closure and chemical composition in
600 Monte Cimone (I) during MINATROC. *Atmospheric Chemistry and Physics* 4, 889-902.
- 601
- 602 Qin, Y., Kim, E., Hopke, P.K., 2006. The concentration and sources of PM2.5 in metropolitan
603 New York City. *Atmospheric Environment* 40, 312-332.
- 604
- 605 Restad, K., Isaksen, I. S. A., and Berntsen, T. K., 1998. Global distribution of sulphate in the
606 troposphere: A three-dimensional model study. *Atmospheric Environment* 32, 3593-3609.
- 607
- 608 Salameh, D., Detournay A., Pey, J., Pérez, N., Liguori, F., Saraga D., Bove, M.C., Brotto P,
609 Cassola F., Massabò, D., Latella A., Pillon, S., Formenton, G., Patti, S., Armengaud, A., Piga,

- 610 D. Jaffrezou, J.L. Bartzisf, J., Tolis, E., Prati, P. Querol, X., Worthama, H., Marchand, N.,
611 2015. PM_{2.5} chemical composition in five European Mediterranean cities: A 1-year study.
612 Atmospheric research 155, 102-117.
613
- 614 Schembari, C., Bove, M.C., Cuccia, E., Cavalli, F., Hjorth, J., Massabò, D., Nava, S., Udisti,
615 R., Prati, P. 2014. Source apportionment of PM₁₀ in the Western Mediterranean based on
616 observations from a cruise ship. Atmospheric Environment 98, 510-518.
617
- 618 Seinfeld, J.H., Pandis, S.N., 1998. Atmospheric Chemistry and Physics. John Wiley and Sons.
619
- 620 Skamarock, W.C., Klemp, J.B., Dudhia, J., Gill, D.O., Barker, D.M., Huang, X.Z., Wang, W.,
621 Powers, J.G., 2008. A Description of the Advanced Research WRF Version 3. Technical
622 report. Mesoscale and Microscale Meteorology Division, NCAR, Boulder, Colorado.
623
- 624 Thurston, G.D., Spengler, J.D., 1985. A quantitative assessment of source contributions to
625 inhalable particulate matter pollution in metropolitan Boston. Atmospheric Environment 19,
626 9-25.
627
- 628 Van Aardenne, J., A. Colette, B. Degraeuwe, P. Hammingh, M. Viana, I. de Vlieger, 2013.
629 The impact of international shipping on European air quality and climate forcing. EEA
630 Technical Report no. 4.
631
- 632 Vecchi, R., Valli, G., Fermo, P., D'alessandro, A., Piazzalunga, A., Bernardoni, V., 2009.
633 Organic and inorganic sampling artefacts assessment. Atmospheric Environment 43, 1713-
634 1720.
635
- 636 Velchev, K., Cavalli, F., Hjorth, J., Vignati, E., Dentener, F., Raes, F., 2011. Ozone over
637 Western Mediterranean Sea-results from two years of shipborne measurements. Atmospheric
638 Chemistry and Physics 11, 675-688.
639
- 640 Viana, M., Amato, F., Alastuey, A., Querol, X., 2009. Chemical Tracers of particulate
641 emissions from commercial shipping. Environmental Science Technology 43, 7472-7477.

- 642 Zhao, M., Zhang, Y., Maa, W., Fu, Q., Yang, X., Li, C., Zhou, B., Yu, Q., Chen., L., 2013.
643 Characteristics and ship traffic source identification of air pollutants in China's largest port.
644 Atmospheric Environment 64, 277–86.

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645 **FIGURE AND TABLE CAPTIONS**

646 **Figure 1.** PM10 chemical composition obtained from raw data and conversion factors during
647 July (top), August (centre) and September (bottom) campaigns. PM10 gravimetric values,
648 which were affected by large uncertainties, are also shown in each panel as 1-sigma band
649 delimited by the two dashed lines.

650 **Figure 2. a)** Ionic balance for the three cruise campaigns in neq m^{-3} **b)** the ratio between the
651 sum of NO_3^- and SO_4^{2-} with NH_4^+ is compared with the Cl^- concentration.

652 **Figure 3.** Scatter plot between SO_4^{2-} and Cl^- concentrations. When the Cl^- concentrations are
653 high (greater than 10 neq m^{-3}) the SO_4^{2-} concentrations are low. The marked data is the sample
654 used to calculate the $\text{MSA/nssSO}_4^{2-}\text{-bio}$ ratio.

655 **Figure 4.** PMF profiles (left axis, coloured bars) and explained variation factors, EV (right
656 axis, white circles) of the PM10 sources resolved in all the three cruise weeks in summer
657 2011.

658 **Figure 5.** Average source apportionment obtained by the PMF analysis of the PM10 data sets
659 collected during the summer 2011.

660 **Figure 6.** Average apportionment of elements/compounds concentration obtained by PMF
661 analysis calculated with the PM10 data sets of the whole field campaign.

662 **Figure 7.** Time trends of the five pollutant sources (factors) obtained by PMF analysis during
663 the three cruise weeks in summer 2011.

664 **Figure 8.** Average apportionment of the total sulphate (relative values) obtained by PMF
665 (right histograms) and by the chemical approach described in Section 3.3 (left histograms) for
666 the three cruises of the 2011 campaign.

667 **Table 1.** Average PM10 composition and BC obtained by Aethalometer for the three
668 campaigns in summer 2011: average (A) and standard deviation (St. Dev) of concentration
669 values were calculated with the samples (reported as percentage frequency, F) with
670 concentration values above their Minimum Detection Limit (MDL). For Cl, K and Ca both the
671 total concentration by ED-XRF and the soluble fraction by IC are reported.

672 **Table 2.** Contributions to the total SO_4^{2-} concentration in absolute values, for the three cruise
673 weeks in 2011 and for the previous campaign in June 2010 determined with the method
674 described in Section 3.3.

675 **Table 3.** Average source apportionment obtained by the PMF analysis of the PM10 data sets
676 collected during the summer 2011 separately for the three cruise campaigns. The average
677 source apportionment is reported in absolute and relative values.

678

679 **Electronic supplement material**

680 **Figure E1.** Route of Costa Concordia during the three campaigns in summer 2011.

681 **Figure E2.** Sea level pressure composite mean (left) and anomalies (right) with respect to the
682 1981-2010 climatology for the July (top panel), August (center) and September (bottom)
683 campaigns, obtained from the NCEP/NCAR Reanalysis (images provided by the
684 NOAA/ESRL Physical Sciences Division, Boulder Colorado, from their web site at
685 <http://www.esrl.noaa.gov/psd/>).

686 **Figure E3.** Times series of temperature, relative humidity and pressure (top), and wind speed
687 (bottom) recorded by the meteorological instrumentation on board the ship during the three
688 campaigns.

689 **Figure E4.** Air mass back trajectories calculated using the HYSPLIT model, related to 21
690 July (right) and 16 August 2011 (left).

691 **Figure E5.** Time trends of sea salt component of PM10 obtained as described in Section 2.3
692 and correlation between wind velocity (km/h) (top figure) and wind prevalent direction
693 (bottom) along the open sea tracks considered.

694 **Figure E6.** 10-m wind fields simulated by the non-hydrostatic mesoscale model WRF-ARW
695 relative to the sea salt events: Savona-Barcelona tracks of the July 18 (top) and Palermo-
696 Civitavecchia tracks of the July 24 (bottom).

697 **Table E1.** Minimum detection limits for each species in ED-XRF and PIXE analysis; PIXE
698 analysis have been used for the samples relative to July week only.

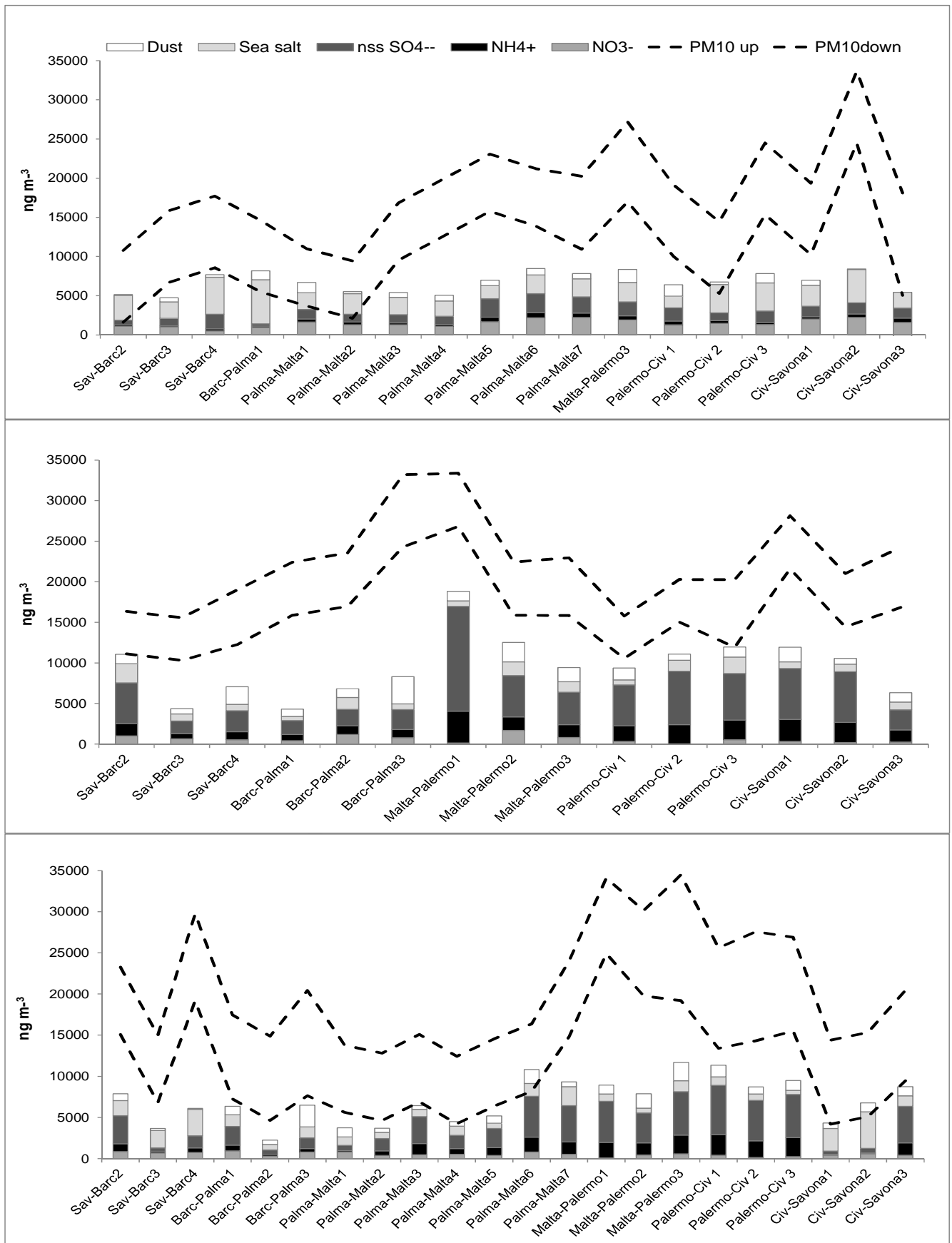
699 **Table E2. a)** Relations used to calculate the different contributions to the total SO_4 ; sea salt
700 sulphate and not sea salt sulphate, divided in: anthropogenic, biogenic and crustal. A value of

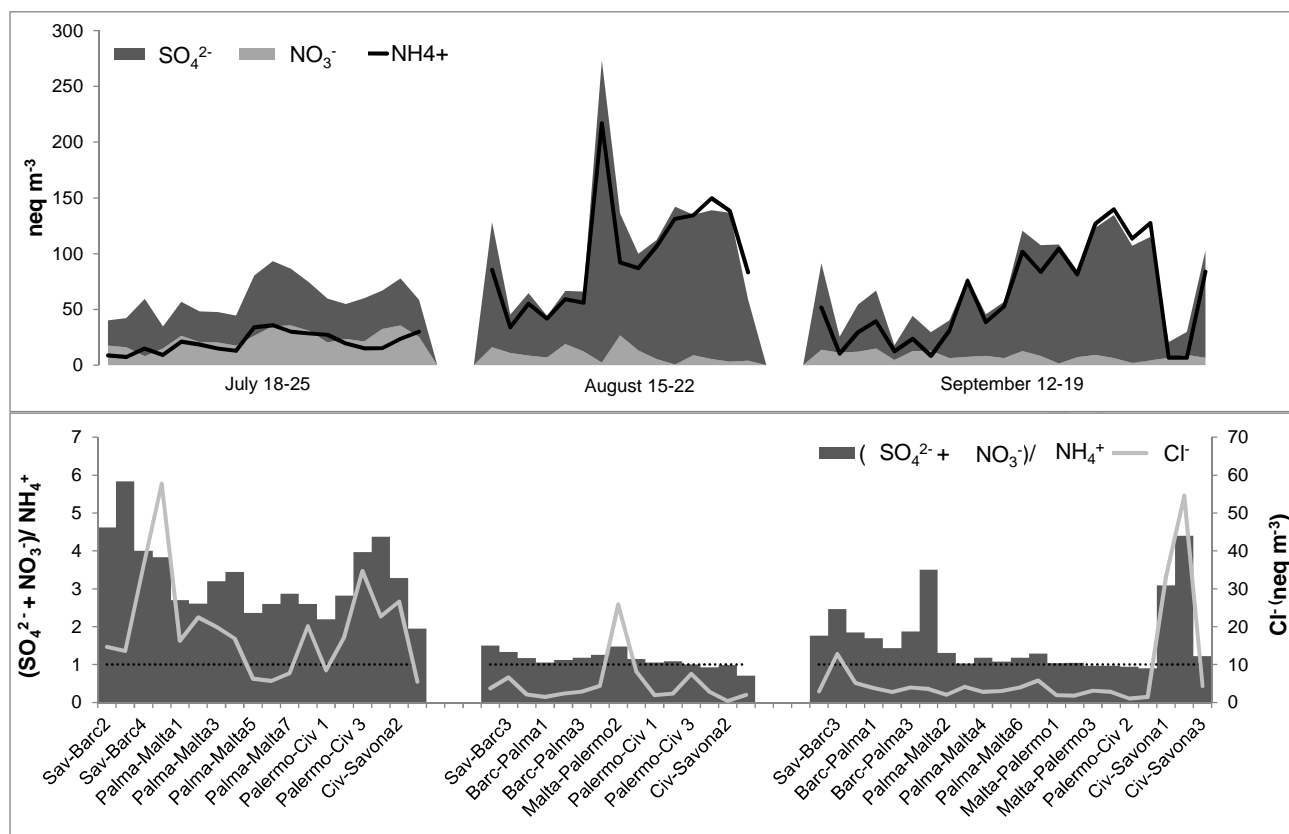
701 0.498 was used for the ratio between SO₄ and Mg in seawater (Keene et al., 1986). Based on
702 the sulphur content in crustal material (Seinfeld and Pandis, 1988), the ratio between SO₄ and
703 Ca in dust was estimated to be 0.097 g/g. Sea salt Ca was obtained as $ssCa = 0.0437 * Na$ (
704 Keene et al., 1986) **b)** details of the calculation to extract the $nssSO_{4bio}$ concentration in a
705 sample influenced by fresh marine air (see text, Section 3.3).

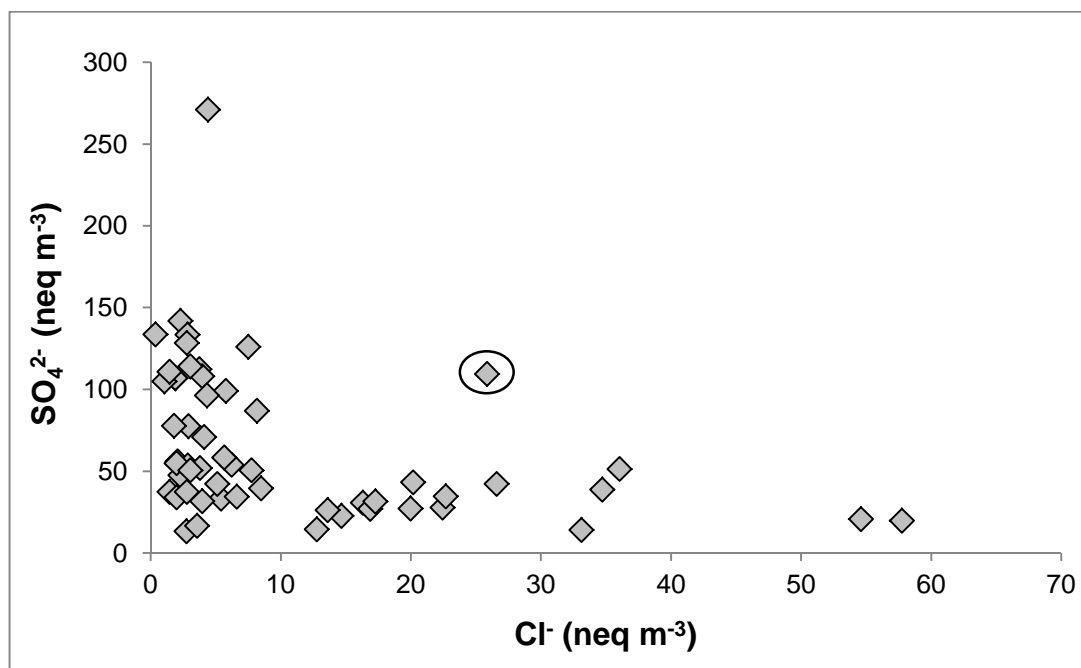
ng m ⁻³			
	A	St. Dev	F
PM10	13113	4778	100%
S	1684	933	67%
Cl	209	376	38%
K	340	291	65%
Ca	151	120	93%
Ti	31	19	98%
V	16	13	95%
Cr	10	5	58%
Mn	5	4	75%
Fe	164	101	98%
Ni	7	5	87%
Cu	5	3	64%
Zn	16	15	87%
Br	7	5	58%
Ba	15	7	27%
Pb	4	3	16%
MSA	54	28	93%
Cl ⁻	381	452	98%
NO ³⁻	882	584	98%
SO ₄ ²⁻	3216	2254	100%
Na ⁺	1003	566	100%
NH ₄ ⁺	1043	869	100%
K ⁺	151	150	27%
Mg ²⁺	139	79	100%
Ca ²⁺	222	114	100%
BC	570	501	100%

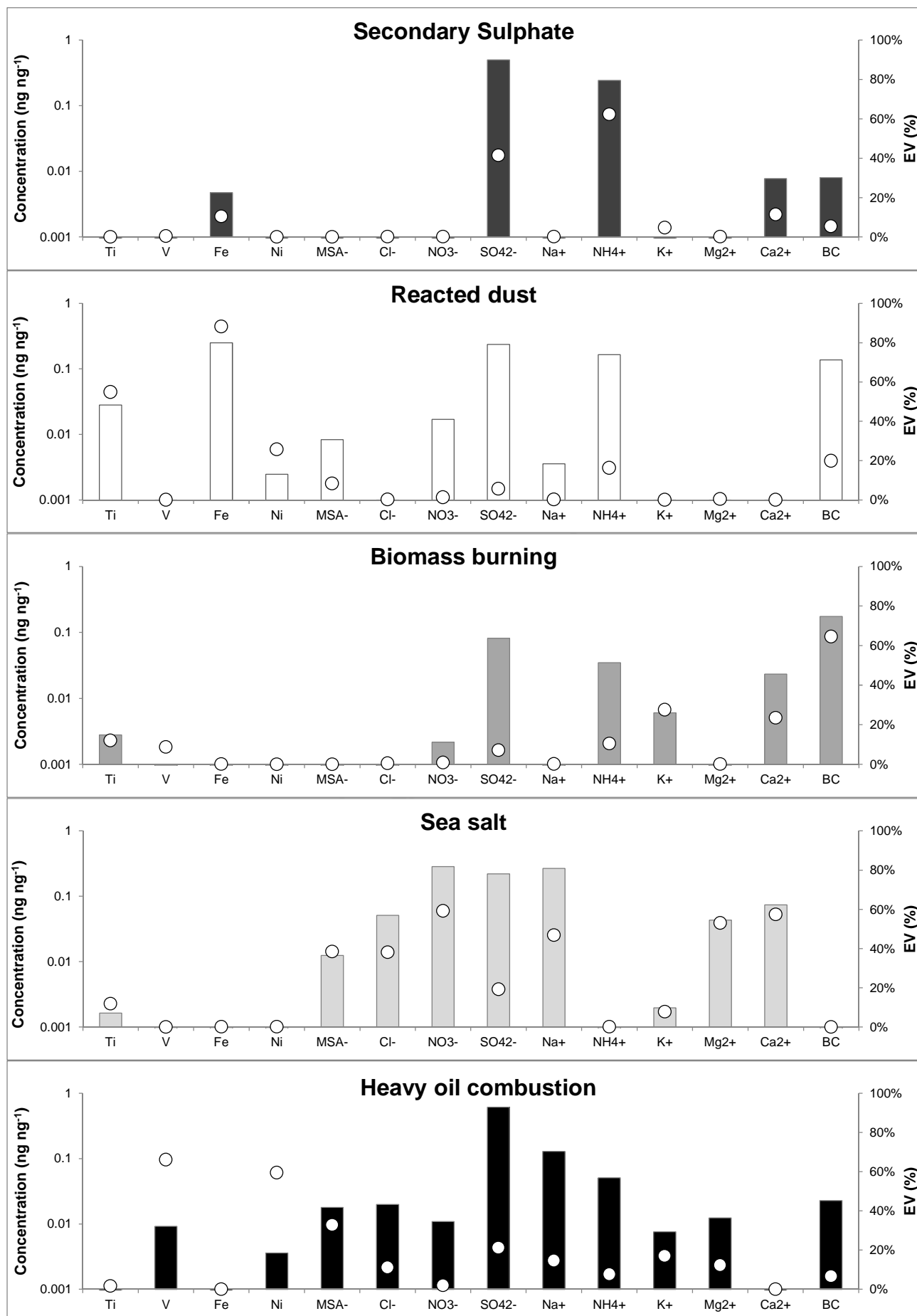
ng m ⁻³	2010 June 7-14	2011 July 18-25	2011 August 15-22	2011 September 12-19
totSO₄²⁻	4550	1760	4820	3100
ssSO₄²⁻	280	420	150	250
nssSO₄²⁻_{crustal}	30	30	60	40
nssSO₄²⁻_{biogenic}	1290	870	750	400
nssSO₄²⁻_{anthropogenic}	2950	440	3850	2410

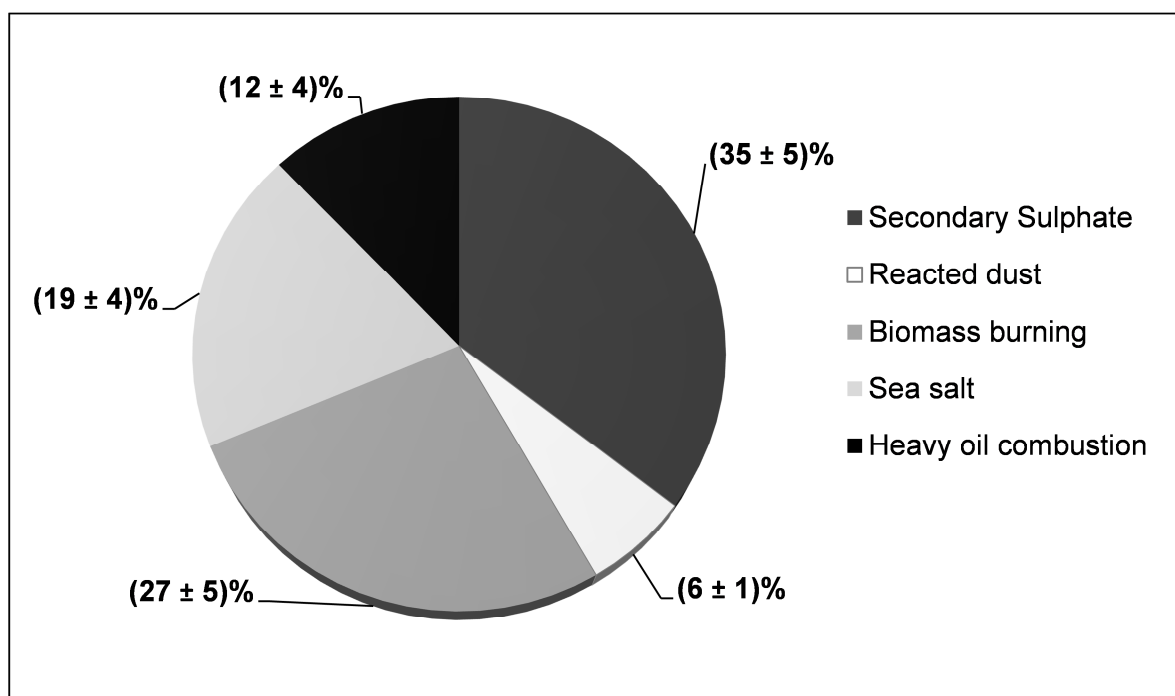
Source	July 18-25		August 15-22		September 12-19	
	(ng m ⁻³)	(%)	(ng m ⁻³)	(%)	(ng m ⁻³)	(%)
<i>Secondary Sulphate</i>	730 ± 170	14 ± 3	5730 ± 720	41 ± 5	3730 ± 500	39 ± 5
<i>Reacted dust</i>	650 ± 40	12 ± 1	980 ± 80	7 ± 1	250 ± 40	3 ± 1
<i>Biomass burning</i>	540 ± 200	10 ± 4	3890 ± 600	28 ± 4	3170 ± 490	33 ± 5
<i>Sea salt</i>	2260 ± 330	43 ± 6	1740 ± 440	13 ± 3	1420 ± 380	15 ± 4
<i>Heavy oil combustion</i>	1110 ± 150	21 ± 3	1470 ± 410	11 ± 4	910 ± 410	10 ± 4



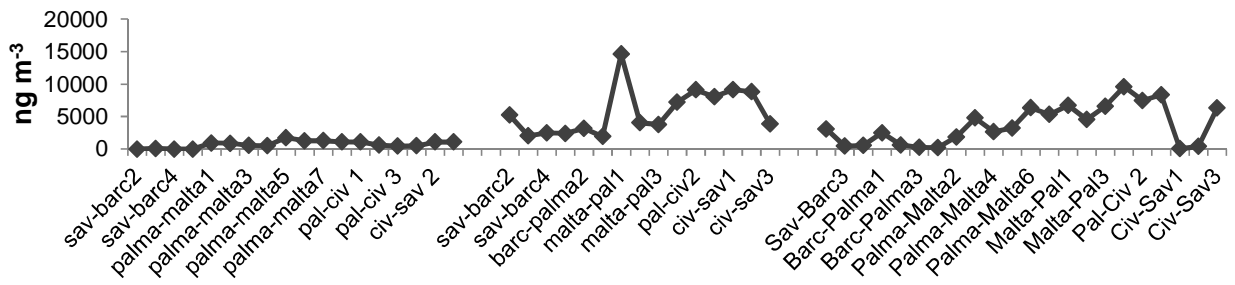




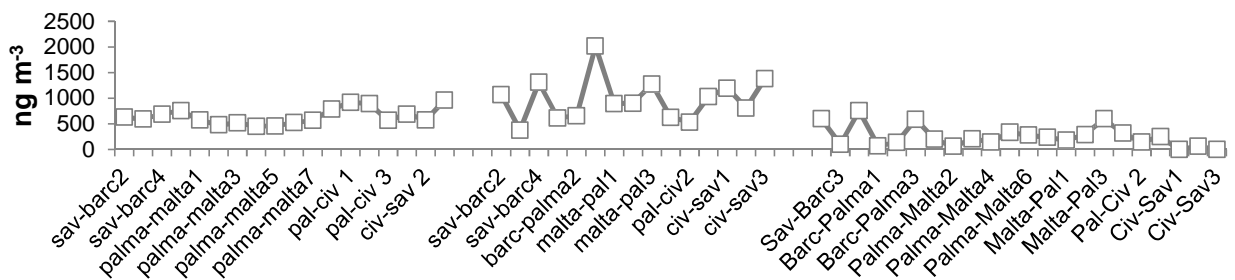




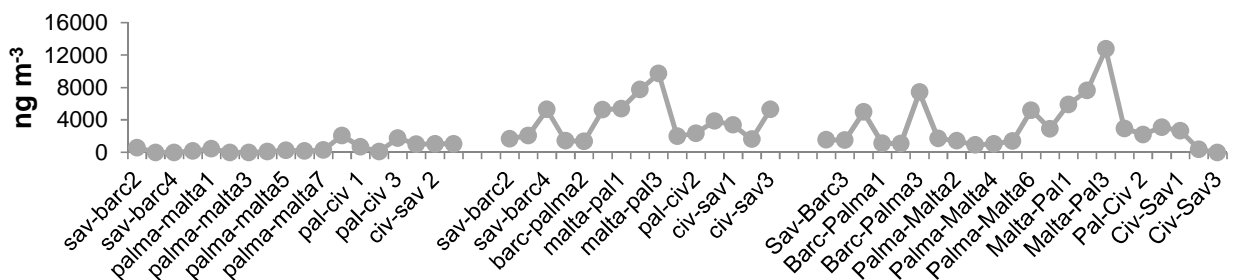
Secondary Sulphate



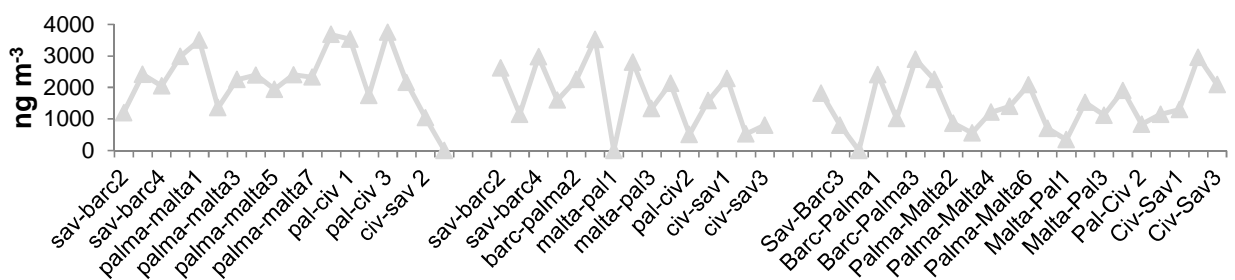
Reacted dust



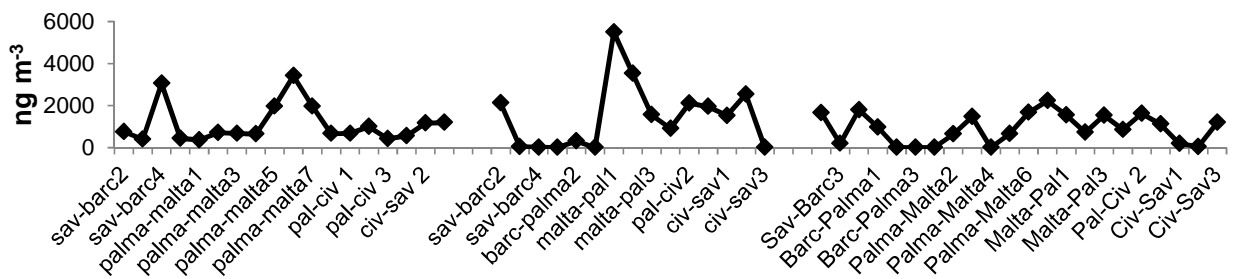
Biomass burning

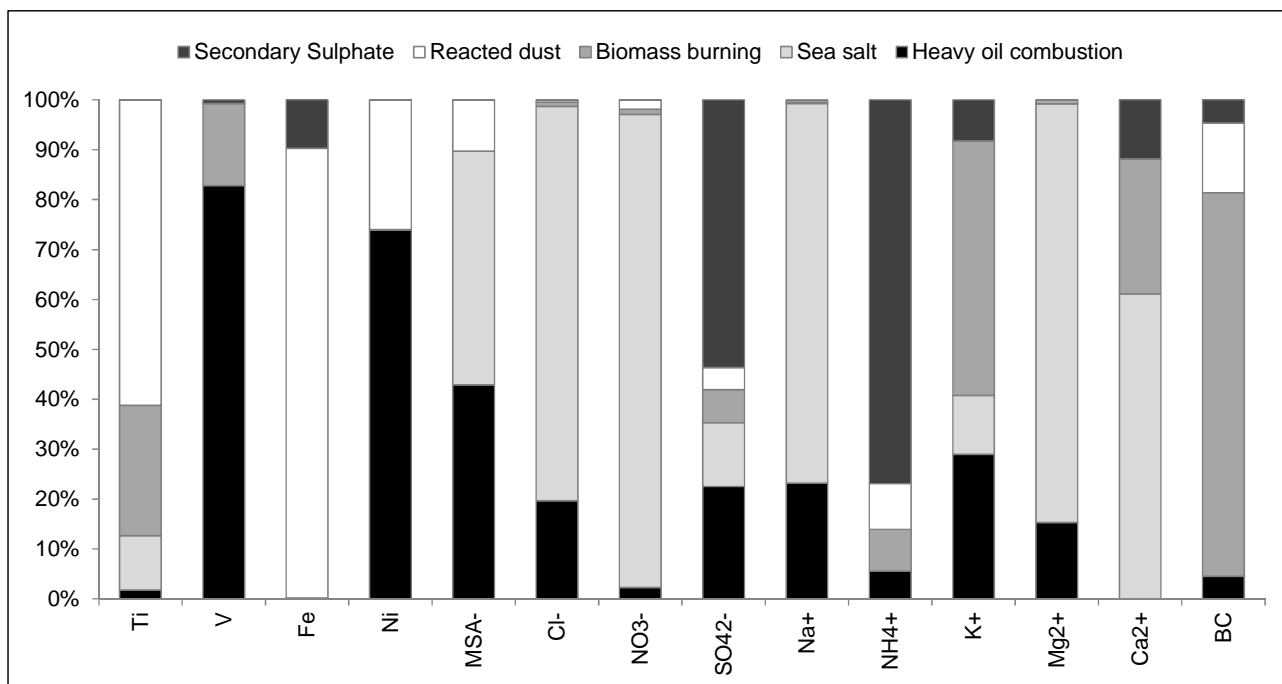


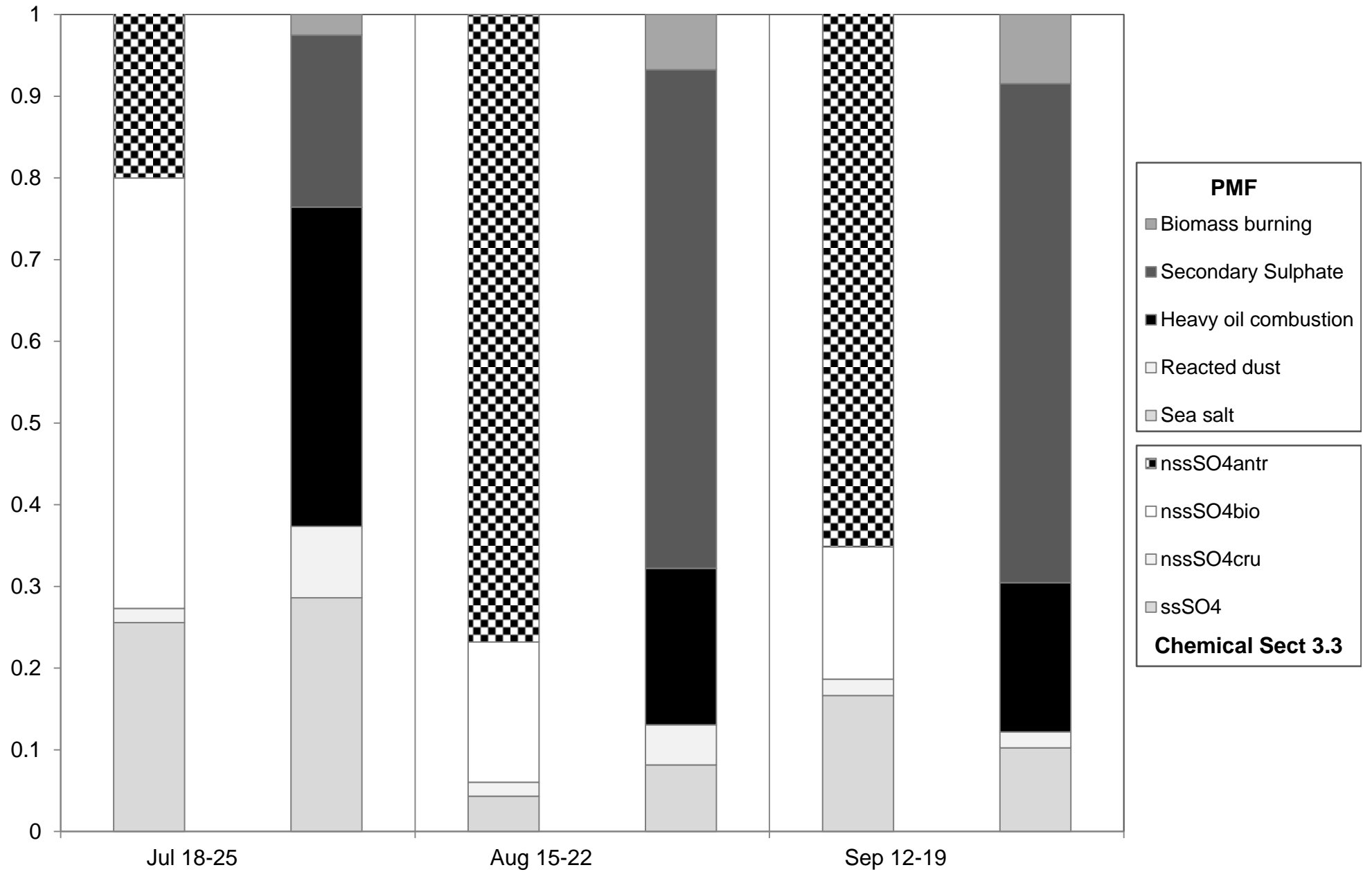
Sea salt



Heavy oil combustion







HIGHLIGHTS

- A conclusive PM₁₀ sampling campaign on a cruise ship was performed in summer 2011
- PMF analysis allowed evaluating the main PM₁₀ sources met along the ship route
- Large marine biogenic sulphur production was identified as function of strong winds
- The study disentangles primary ship emissions and secondary sulphates
- Primary ship emissions contributed on average to $(12 \pm 4)\%$ of PM₁₀